

61        **The green thorns of *Ulex europaeus* play both defensive and photosynthetic roles:**  
62                    **consequences for predictions of the Enemy Release Hypothesis**

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85

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106 **Abstract** The widespread invasive success of *Ulex europaeus*, a thorny shrub native to NW  
107 Europe, remains to be understood from a functional perspective. According to the Enemy  
108 Release Hypothesis (ERH), lower pressure by vertebrate herbivores in the invaded areas  
109 should lead to lower investment in (costly) physical defenses, allowing plants to invest more in  
110 growth and/or reproduction. While *U. europaeus* seedlings have spines, adult plants have  
111 thorns, which are the main photosynthetic tissue (leaves are reduced to small phyllodes).  
112 Therefore, reduced biomass investment in thorns could compromise photosynthesis and  
113 growth in the invaded range. We hypothesized that *U. europaeus* plants in invaded ranges  
114 should show a reduction in the defensive components of thorns (e.g., softer and less fibrous  
115 tissues), but not reduced biomass allocation. We compared *U. europaeus* plants from the  
116 invaded (Chile) and native (Spain) distribution ranges regarding: i) spinescence traits (thorn  
117 length, width, biomass, slenderness and bending strength) in adult plants, ii) thorn fiber content  
118 and digestibility (proxies for palatability) in adult plants, and iii) spine density in seedlings grown  
119 in a common garden. As expected, plants in the invaded range showed larger thorns, which  
120 contained less cellulose, were slenderer and easier to bend than those from plants in the native  
121 range. Likewise, seedlings from the invaded range showed lower spine density and more  
122 diameter growth, thus supporting the ERH. We found functional changes in spinescence traits  
123 between distribution ranges consistent with the dual role of thorns in *U. europaeus*, and these  
124 changes may partly explain its invasiveness.

125

126 **Keywords:** cellulose; gorse; physical defenses; plant-herbivory interactions; spines, thorns

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129 **Introduction**

130 A major goal of invasion ecology is to understand why some exotic species are  
131 successful invaders (Catford et al. 2009), and to this end, studies comparing native and invaded  
132 distribution ranges are especially relevant (Hierro et al. 2005; Cronin et al. 2015). The enemy  
133 release hypothesis (ERH) is one of the main hypotheses with a biogeographical approach to  
134 explain the success of exotic invasive plant species (invasive plants, hereafter) (Keane and  
135 Crawley 2002; Colautti et al. 2004; Jeschke et al. 2012). The ERH posits that the reduced  
136 abundance or absence of natural enemies in the colonized range may allow the invasive plants  
137 to invest more in growth and/or reproduction and less in defense (Blossey and Notzold 1995;  
138 Keane and Crawley 2002; Colautti et al. 2004). Several studies, particularly evaluating  
139 chemical defenses, have verified the predictions of the ERH (e.g., Joshi and Vrieling 2005; Liu  
140 and Stilling 2006; Hull-Sanders et al. 2007; Gruntman et al. 2017), but fewer studies have found  
141 shifts toward both lower defense and higher growth in introduced populations (Bossdorf et al.  
142 2005).

143 Plant physical and structural defenses (e.g., spinescence, pubescence, sclerophylly)  
144 respond to vertebrate herbivory pressure (Obeso 1997; Hanley et al. 2007; Gómez and Zamora  
145 2002; Musariri et al. 2018). Spinescence, a term including sharp-pointed modifications of leaf  
146 parts (spines), stems or branches (thorns) and epidermal outgrowths (prickles), is an effective  
147 physical defense against vertebrate herbivores (Cooper and Owen-Smith 1986; Cooper and  
148 Ginnett 1998; Milewski et al. 1991; Hanley et al. 2007; Wigley et al. 2019). Sclerophylly  
149 (hardened leaves) is a defensive trait effective against both vertebrate and invertebrate  
150 herbivores, decreasing the palatability and digestibility of plant tissues through a high content  
151 of digestibility-reducing compounds, such as lignin or cellulose fibers, and a low nutrient content

152 (Lundberg and Åström 1990; Turner 1994; Hanley et al. 2007). Some mechanical properties  
153 related to sclerophylly, e.g., flexural stiffness and fracture toughness, are predictors of  
154 resistance to herbivores (Choong et al. 1992; Lucas et al. 2000; Hanley et al. 2007; Guerra et  
155 al. 2010). Plant physical defenses may be affected by extrinsic and intrinsic factors, such as  
156 environmental conditions (González et al. 2008) and plant ontogeny (Hanley et al. 2007),  
157 respectively. For instance, due to limitations in resource acquisition, seedlings often show less  
158 investment in physical defenses than mature plants (Boege and Marquis 2005; Hanley et al.  
159 2007; Mediavilla et al. 2018; but see Clark and Burns 2015). Nonetheless, physical defenses  
160 in seedlings are effective against vertebrate and invertebrate herbivores, yet costly (Cooper  
161 and Ginnett 1998; Boege and Marquis 2005; Cash and Fulbright 2005; Barton and Hanley  
162 2013).

163 In line with the ERH, a meta-analysis reported that invasive plants in the invaded range  
164 generally show lower investment in digestibility-reducing compounds, such as tannins and  
165 lignin, compared to the native range (Zhang et al. 2018). Likewise, a perennial daisy showed  
166 decreased leaf toughness in the invaded range compared to the native range (Brandenburger  
167 et al. 2020). Yet, spinescence traits, such as thorns and spines, have not been compared  
168 between invaded and native ranges in any plant invader, to the best of our knowledge. Invasive  
169 thorny plants, originally well-armed against vertebrate herbivores, are particularly suitable  
170 model species to test predictions of the ERH.

171 *Ulex europaeus* L. (Fabaceae), a thorny shrub native to NW Europe, is considered  
172 among the world's worst invasive species (Lowe et al. 2000; Clements et al. 2001). Once  
173 established in the invaded ecosystems, through mechanisms not completely understood, *U.*  
174 *europaeus* dramatically increases in size and abundance, creating thick monocultures (Muñoz

175 2009). Some characteristics that may favor its invasiveness are high growth rate, strong ability  
176 to resprout, high seed output, persistent seed bank, and substantial genetic and phenotypic  
177 diversity (Reyes et al. 2009; Atlan et al. 2010; Hornoy et al. 2011, 2013; Atlan et al. 2015). In  
178 a common garden experiment, *U. europaeus* seedlings from the invaded range were taller than  
179 plants from the native range but had similar insect infestation rates and similar concentrations  
180 of defensive alkaloids (Hornoy et al. 2011, 2012). Because these studies did not consider  
181 physical defenses, it remains an open question whether changes in physical defenses could  
182 be associated with the increased growth of *U. europaeus* in the invaded range. In this study,  
183 we aimed to answer this question.

184 In the native range, *U. europaeus* is consumed by wild vertebrate herbivores, such as  
185 red deer (*Cervus elaphus*), roe deer (*Capreolus capreolus*) and rabbit (*Oryctolagus cuniculus*)  
186 (González-Hernandez and Silva-Pando 1996, 1999; Alves et al. 2006), and by domestic  
187 ungulates, such as goats, sheep, cows, and horses (Putman et al. 1987; Howe et al. 1988;  
188 Clements et al. 2001; Atlan et al. 2015). In contrast, in the invaded range, specifically in Chile,  
189 there is no overlap between the distribution of *U. europaeus*, always close to human-impacted  
190 ecosystems, and the distribution of native big vertebrate herbivores, such as guanaco (*Lama*  
191 *guanicoe*), vicuña (*Vicugna vicugna*) or huemul (*Hippocamelus bisulcus*), which are generally  
192 restricted to remote, relatively pristine areas (Jaksic 2001; GBIF.org 2021). As to exotic  
193 vertebrate herbivores in Chile, those herbivores possibly coexisting with *U. europaeus* in the  
194 native range were introduced at least 50 years after *U. europaeus* (Jaksic 1998), a time period  
195 long enough for several invasive species to evolve traits related to invasive potential (e.g.,  
196 growth rate) (Whitney and Gabler 2008; Buswell et al. 2011; Gianoli and Molina-Montenegro  
197 2021). Some of those introduced herbivores hardly subsist in the wild (e.g., European fallow

198 deer, *Dama dama*), while others (e.g., red deer, *Cervus elaphus*) are not present in most of the  
199 areas invaded by *U. europaeus* (Iriarte et al. 2005; GBIF.org 2021). In addition, introduced  
200 vertebrate herbivores, such as wild rabbits and domestic ungulates, often prefer native plant  
201 species over exotics, thus favoring invasion by exotic plants (Jaksic 1998; Holmgren et al.  
202 2000; Holmgren 2002; Jaksic et al. 2002; Vázquez 2002; Figueroa et al. 2004; Mitchell et al.  
203 2006; Parker et al. 2006; Vavra et al. 2007; Oduor et al. 2010). Moreover, in the invaded range,  
204 *U. europaeus* is not commonly used as fodder species for domestic ungulates, as is the case  
205 in the native range (Atlan et al. 2015; Dietl et al. 2009). For all the above reasons, *U. europaeus*  
206 likely faces lower pressure by vertebrate herbivores in the invaded range, compared to the  
207 native range. According to the ERH, the lower herbivory pressure in the invaded range should  
208 lead to lower investment in physical defenses and/or increased growth in invasive populations  
209 of *U. europaeus*, but this specific hypothesis has not been experimentally tested.

210         Concerning the evaluation of the ERH predictions, *U. europaeus* has a particularly  
211 interesting biological feature: since leaves in mature plants are reduced to spine-like phyllodes,  
212 the green thorns –together with the green branches– are at the same time physical defenses  
213 and the main photosynthetic tissue, in terms of mass and surface (Clements et al. 2001; Cubas  
214 2008). Therefore, the ERH prediction of a reduction in physical defenses in the invaded range  
215 should not result from reduced biomass allocation to thorns, which could hinder plant growth.  
216 Instead, structural changes associated with a reduction in the (costly) defensive components  
217 of thorns, together with increased growth, should be expected.

218         In the present study, we compared physical defenses between *U. europaeus*  
219 populations in the native (Spain) and invaded (Chile) distribution ranges. In adult plants  
220 collected from natural field populations, we measured spinescence traits, thorn allocation

221 parameters and thorn bending strength, as well as fiber content and digestibility of thorns (as  
222 proxies for palatability). In addition, we measured spine density and growth in seedlings from  
223 both origins (Spain and Chile) grown in a common garden. In the invaded range, we expected  
224 to find, compared to the native range: i) higher seedling growth, and ii) spinescence traits  
225 associated with a reduction in the defensive components of thorns (e.g., softer, less fibrous  
226 tissues with decreased content of digestibility-reducing compounds), which may reduce  
227 effectiveness against herbivores.

228

## 229 **Material and Methods**

230

### 231 *Study species*

232 *Ulex europaeus* (gorse) is a shrub that can be 2.5 m tall in the native range of  
233 distribution, but it can reach over 5 m in the invaded range (Clements et al. 2001; Cubas 2008,  
234 Barrera-Cataño et al. 2019). It has two types of branches, one of limited growth that lignifies  
235 fast and another of prolonged growth that develops when conditions are favorable and lignifies  
236 the next year. The latter type is softer and more flexible. Young plants develop alternate  
237 petiolate leaves with 1-3 leaflets, while leaves in the adult plants are modified and reduced to  
238 triangular rigid spiny phyllodes (5-12 mm). Therefore, green branches and stems are the main  
239 photosynthetic tissues in adult plants. Branches have a spiny end, and each branch develops  
240 sharp-pointed stems (primary thorns), which in turn develop secondary and tertiary thorns.  
241 Each thorn has a phyllode in its base. See Cubas (2008) for a detailed description and  
242 schematic pictures of the described plant parts and Figures 1 and 2 of the present manuscript.



243 Domestic animals (e.g., sheep and goat) can ingest different parts of *U. europaeus*,  
244 including green and lignified thorns, new vegetative buds, bark, and flowers (Atlan et al. 2015;  
245 Radcliffe 1986). However, though *U. europaeus* was introduced worldwide as forage,  
246 ornamental or hedge plant in the 1800s and 1900s, it is not used as fodder in the invaded  
247 range, which, together with the absence of management practices, may have favored its  
248 expansion (Atlan et al. 2015). In Chile, it was first reported in 1847, likely introduced from  
249 England (Hornoy et al. 2013), being now spread from 33°S to 41°S, reaching high densities  
250 and large sizes, and even invading areas of conservation value (Clements et al. 2001; Muñoz  
251 2009; Fuentes et al. 2014; Altamirano et al. 2016). The fact that this perennial species has  
252 long-lived seeds (more than 30 years) allows it to keep a significant genetic diversity (Atlan et  
253 al., 2010; Hornoy et al. 2013).

254

#### 255 *Plant material collection*

256 The collection of branches, thorns and seeds was done during spring 2018 in three *U.*  
257 *europaeus* populations each with similar climatic conditions within the native (Pontevedra,  
258 Spain) and invaded (Concepción, Chile) ranges (Online resources 1 and 2). Climatic data for  
259 the period 1982-2012 were obtained from <https://es.climate-data.org>. All the sampled  
260 populations were located in roadside areas, which reduces the likelihood of the presence of  
261 ungulate herbivores, thus resulting in the observed low browsing damages in all populations at  
262 both native and invaded ranges (S. Medina-Villar, personal observation). In Spain, *U.*  
263 *europaeus* populations located far from roads and urban areas can be heavily browsed by  
264 domestic and wild ungulates (S. Medina-Villar, personal observation). We chose *U. europaeus*  
265 populations with low herbivory because we did not aim to record the short-term responses to

266 continuous high herbivory (i.e., acclimatization). We rather sought to detect the consequences  
267 of long-term contrasting levels of herbivory in the native and invaded ranges, expecting  
268 functional changes in *U. europaeus* traits in response to decreased herbivory pressure in Chile.  
269 Thus, albeit lightly browsed, the sampled populations in Spain certainly had a longer  
270 evolutionary history facing herbivory (around 200 years more) than their Chilean counterparts.  
271 In addition, by sampling roadside populations in both ranges (Chile and Spain) we could keep  
272 the factor “land use” constant. To measure spinescence traits and fiber content in adult  
273 individuals, a total of 60 branches (1 branch x 10 individuals x 3 populations x 2 distribution  
274 ranges) were collected. Besides, a pool of seeds from 30 individuals (ca. 25 g of seeds: slightly  
275 over 5,000 seeds) were collected in each distribution range. The minimum distance between  
276 individuals in each of the six populations sampled was 1 m.

277

#### 278 *Spinescence traits*

279 Branches collected in the field were oven-dried at 60 °C for at least 48 h before we  
280 measured spinescence traits. Measurements were made in green thorns, avoiding senesced  
281 parts. Thorn length (L) and thorn-base width perpendicular to the branch (W) were measured  
282 in a total of 1200 primary thorns (20 thorns x 1 branch x 10 individuals x 3 populations x 2  
283 distribution ranges) using a digital caliper (Figs. 1A, B). Thorn length was measured from the  
284 base of the thorn to its tip (Figs. 1A, B). Thorn length has been shown to increase in response  
285 to herbivory in other *Fabaceae* species (Young et al. 2003; Wigley et al. 2019). Thorn width  
286 can be used to estimate defense effectiveness against herbivores (Pérez-Harguindeguy et al.  
287 2013). The L:W ratio was calculated, which is analogous to the slenderness ratio

288 (height:diameter; the inverse of sturdiness) commonly measured in plants (Haase 2008;  
289 Valladares et al. 2012).

290 In each branch, four branch segments were randomly subsampled and cut to calculate  
291 thorn density (number of thorns per cm of branch segment) and biomass allocation to thorns,  
292 i.e., the ratio between the thorn dry weight and the branch segment dry weight (480 branch  
293 segments in total = 4 branch segments x 2 branches x 10 individuals x 3 populations x 2  
294 distribution ranges). Previous studies in other shrub species have reported a positive  
295 correlation between spine density and herbivory pressure (Cavagnaro et al. 2017; Musariri et  
296 al. 2018). All the above-mentioned variables were averaged by individual (10 per population).  
297 To standardize measurements, all the spinescence traits were divided by branch length.

298

#### 299 *Bending strength of thorns*

300 In freshly collected primary thorns, flexural stiffness (elastic resistance to bending) was  
301 estimated by the following formula:  $(F/D) \times B^3$ , where  $F$  is the force,  $D$  is the displacement, and  
302  $B$  is the span length between supports in the bending test (Hanley et al. 2007). We kept  
303 parameters  $B$  and  $D$  constant and estimated  $F$  in a total of 144 thorns (6 thorns x 4 randomly  
304 selected individuals x 3 populations x 2 distribution ranges). Using a vise, thorns were fixed at  
305 3 cm from the thorn tip by pressing the thorn side (width) that is perpendicular to the branch.  
306 At 1 cm from the thorn tip, a plastic container was hanged –using a thread as handle– and an  
307 increasing weight was applied to the container (by adding rice) until the thorn bent at a 10°  
308 angle. Therefore,  $B$  was 2 cm,  $D$  was 10° angle, and  $F$  was the weight necessary to bend a  
309 thorn 10°, which we termed bending strength (Van Gelder et al. 2006; Shah et al. 2017). A  
310 scheme of the bending test performed is shown in Figure 1C. The total length of each thorn

311 submitted to the bending test was also recorded because thorn size may influence its bending  
312 strength (Van Gelder et al. 2006).

313

#### 314 *Fiber content of thorns*

315 To analyze fiber content, individual thorn samples from each population were pooled,  
316 dried, and then pulverized with a mixer mill (Retsch MM 301). A total of 24 samples (4 pooled  
317 samples x 3 populations x 2 distribution ranges) were analyzed for neutral detergent fiber  
318 (NDF), acid detergent fiber (ADF), lignin, cellulose, and dry matter digestibility (DMD) using the  
319 filter bag technique; this was done in an Ankom Automated Fiber Analyzer A2000 (Ankom,  
320 USA), based on the analytical method of Göering and Van Soest (1970).

321

#### 322 *Common garden experiment*

323 In June 2019, air-dried seeds of *U. europaeus* from Concepción (Chile) and Pontevedra (Spain)  
324 were sown in square pots (11 x 11 x 12 cm) containing a 90:10 (v/v) mixture of organic substrate  
325 (50:50 white peat moss and black peat; Projar Professional Seed Pro 5050;) and vermiculite.  
326 Due to seed coat hardness, seeds were pretreated at 80 °C for 30 min to ensure germination  
327 success. A total of 120 seedlings per distribution range grew for six months in the pots in  
328 common environmental conditions (minimum temperature 18 °C, maximum temperature 25 °C,  
329 watered when needed) in a greenhouse at the Alfonso XIII Botanical Garden (Universidad  
330 Complutense de Madrid, Spain). We considered the six-month plants as seedlings because  
331 most of them still had their cotyledons. In each seedling the following parameters were  
332 measured: the stem diameter below cotyledon insertion, plant height and spine density, which  
333 was visually estimated as the percentage of plant height that developed spines. Note that

334 seedlings do not develop thorns, but they show leaves and spines (i.e., sharp-pointed modified  
335 leaves). Figure 2 shows seedlings of *U. europaeus* from Spain and Chile.

336

### 337 *Statistical analyses*

338 Linear Mixed Models (LMM) were performed to assess differences in standardized  
339 spinescence variables (L, W, L:W, density and biomass), fiber contents and DMD between  
340 distribution ranges (invaded vs native range; fixed factor), using population as random factor  
341 to account for population variability. To evaluate the relationship between spinescence  
342 variables and branch length in *U. europaeus* from each range, LMM were performed using  
343 distribution range as fixed factor, branch length as covariate and population as random factor.

344 Principal Component Analyses were performed to visualize relationships among fiber-related  
345 variables (lignin, cellulose, ADF, NDF and DMD) and among spinescence variables in *U.*  
346 *europaeus* thorns from native and invaded ranges. Permutational multivariate analysis of  
347 variance (PERMANOVA) was used to assess differences between distribution ranges  
348 considering all fiber-related variables on the one hand, and all spinescence variables on the  
349 other hand. To evaluate the effect of distribution range (fixed factor) on the bending strength of  
350 thorns (dependent variable), a LMM was performed using L as covariate and population as  
351 random factor.

352 A one-way ANOVA was used to compare the diameter and height of seedlings between  
353 distribution ranges (fixed factor). To deal with excess of zeros in spine density of seedlings  
354 (dependent variable), we performed a Generalized Linear Model (GLM, family quasibinomial)  
355 with distribution range as fixed factor (Zuur et al. 2009). When needed, dependent variables  
356 were transformed with log, root-square or power functions to meet homoscedasticity and

357 normality in LMM and ANOVA. Statistical analyses were performed in R software 3.4.3 (R Core  
358 Team 2017) using “lmer” function (lme4 package) for LMM, “adonis2” function (vegan package)  
359 for PERMANOVA, the built-in R function “prcomp” for PCA, and “glm” function for GLM (Zeileis  
360 et al. 2008; Bates et al. 2015; Jari et al. 2020).

361

## 362 **Results**

363 Spinescence traits in *U. europaeus* plants from the invaded range differed from those of the  
364 native range (PERMANOVA, pseudo- $F_1 = 30.6$ ,  $P < 0.001$ ; PCA in Online Resource 3). The  
365 effect of distribution range was significant for all the branch-length standardized spinescence  
366 traits except for thorn density (LMM, Table 1). Specifically, thorn length (L), thorn width (W),  
367 L:W, and thorn biomass allocation were greater in the invaded range (Fig. 3). Thorn length and  
368 L:W were greater in the invaded range and increased with branch length in both the native and  
369 the invaded ranges (Table 2; Fig. 3). Thorn density did not differ between plants from Chile and  
370 Spain, decreased with increasing branch length in both distribution ranges, but the slope was  
371 greater in plants from the invaded range (Table 2; Fig. 3). Thorn biomass allocation was,  
372 overall, greater in the invaded range, decreased with increasing branch length in the invaded  
373 range and did not vary with branch length in the native range (Table 2; Fig. 3). Thorns from the  
374 native range showed a greater length-corrected bending strength than thorns from the invaded  
375 range, being this difference minimized in the longest thorns (Range,  $F_1 = 9.1$ ,  $P = 0.0035$ ; Thorn  
376 length,  $F_1 = 7.2$ ,  $P = 0.0084$ ; Range x Thorn length,  $F_1 = 6.0$ ,  $P = 0.0156$ , Fig. 4).

377 Overall, *U. europaeus* plants from the invaded range differed from those of the native  
378 range in fiber-related traits (PERMANOVA,  $F = 10.9$ ,  $P < 0.001$ ; PCA in Online Resource 4).  
379 Fiber contents of thorns tended to be greater in the native range, the differences being

380 significant for NDF and cellulose, marginally significant for ADF and non-significant for lignin  
381 (Table 3). Our estimate of digestibility (DMD) tended to be lower in thorns from the native range,  
382 but differences were not significant (Table 3).

383 Seedlings of *U. europaeus* from the native range showed higher spine density than  
384 seedlings from the invaded range ( $t = 5.5$ ,  $P < 0.001$ , GLM; Figures 2 and 5). Interestingly,  
385 whereas 6-month seedlings did not differ in height ( $F_1 = 0.11$ ,  $P = 0.7369$ , ANOVA; Fig. 5),  
386 those from the invaded range showed thicker stems ( $F_1 = 4.5$ ,  $P = 0.0357$ , ANOVA; Fig. 5),  
387 which indicates greater biomass production.

388

## 389 **Discussion**

390 Our results show that, in the invaded range, adult plants of *U. europaeus* shift towards  
391 i) decreased investment in the (costly) defensive components of thorns, i.e., less fibrous,  
392 slenderer and easier to bend thorns, which can result in reduced effectiveness against  
393 herbivores (Lambert et al. 1989; Lucas et al. 2000; Hanley et al. 2007), and ii) increased  
394 investment in photosynthetic tissues: increased thorn size and biomass. Plant stems may have  
395 both structural and functional roles, which may be in conflict with one another (Speck and  
396 Burgert 2011). In the case of adult plants of *U. europaeus*, stems are at the same time physical  
397 anti-herbivore defenses, thorns, and the main photosynthetic tissue, as leaves are reduced to  
398 small triangular squamous phyllodes (Cubas 2008). This particular feature of *U. europaeus*  
399 somewhat challenges the predictions of the Enemy Release Hypothesis (ERH) (Keane and  
400 Crawley 2002; Colautti et al. 2004). Thus, a lower herbivory pressure in the invaded range  
401 should not lead to a lower investment in thorn size or biomass, because it would reduce  
402 interception of photosynthetically active radiation and hence decrease plant growth.

403 Nonetheless, consistent with the fundamental tenets of ERH, the changes in structural design,  
404 fiber composition, biomass allocation and strength in *U. europaeus* thorns in the invaded range  
405 resulted in these physical defenses likely being less efficient against vertebrate herbivores  
406 (Lambert et al. 1989; Read and Stokes 2006; Xu et al. 2013).

407 All the morphological, compositional and mechanical changes observed in thorns in the  
408 invaded range may favor the invasive success of *U. europaeus*. For instance, increased size  
409 and biomass of the main photosynthetic tissues (green thorns) in the invaded range may favor  
410 plant growth because the photosynthetic area (usually leaf area) is the main driver of light  
411 interception and ultimately plant productivity (Gifford et al. 1984; Koester et al. 2014;  
412 Weraduwage et al. 2015). Moreover, a lower investment in costly fibers, such as cellulose  
413 (Poorter and De Jong 1999), in the invaded range leaves more resources available to invest in  
414 other plant functions that contribute to vital rates. In line with our results, an experimental study  
415 with an evergreen tree species and a meta-analysis showed that the content of digestibility-  
416 reducing compounds was lower in plants from the invaded range compared to those from the  
417 native range (Montti et al. 2016, Zhang et al. 2018).

418 The reduced bending strength of thorns in the invaded range, a key result concerning  
419 physical defense effectiveness, was likely associated with differences in the content of  
420 structural compounds (cellulose) and patterns of biomass allocation, resulting in less fibrous  
421 and slenderer thorns. We emphasize that, to our knowledge, the reduction in bending strength  
422 of a physical defense in the invaded range has not been documented for any invasive plant  
423 species. This is rather surprising, given the relevance of physical defenses, such as  
424 spinescence traits, for plant protection against vertebrate herbivores (Obeso 1997; Gómez and  
425 Zamora et al. 2002; Hanley et al. 2007). It is much more common to find results linked to leaf



426 toughness, which is generally lower in the invaded range (Agrawal et al. 2005; Liao et al. 2013;  
427 but see: Franks et al. 2008; Ridenour et al. 2008). We hypothesize that the reduction in bending  
428 strength in *U. europaeus* may be a consequence of the lower selective pressure exerted by  
429 herbivores in the invaded range (Jacksic 1998, 2001; Atlan et al. 2015). However, the specific  
430 mechanical properties of physical defenses that deter vertebrate herbivores are not completely  
431 understood (Read and Stokes 2006; Sanson 2006).

432         Unlike adult plants, seedlings of *U. europaeus* do not develop thorns, but they do show  
433 leaves and spines. Therefore, seedlings are amenable to “traditional” tests of the ERH. We  
434 found a negative relationship between physical defense (spine density) and growth (in  
435 diameter) in seedlings, with more investment in growth and less in defense in the invaded  
436 range, as predicted by the ERH. This shift in resource allocation between defense and growth  
437 was not found in earlier studies addressing the relationship between chemical defenses and  
438 growth in *U. europaeus* (Hornoy et al. 2011, 2012). In fact, few studies have reported both  
439 ERH-predictions, namely decreased defense and increased growth, in the same invasive  
440 species (Bossdorf et al. 2005). Seedlings are generally considered to rely more on chemical  
441 than on physical defenses against herbivores (Barton and Hanley 2013). However, our results  
442 suggest that physical defenses may be relevant during the early establishment of this invasive  
443 shrub. Seedlings are commonly highly susceptible to big vertebrate herbivores, such as  
444 ungulates, which can easily kill the whole seedling (Rafferty et al. 2005). Nevertheless, spiny  
445 seedlings from different species have been shown to effectively deter other herbivores, such  
446 as lagomorphs, rodents and invertebrates (Cooper and Ginnett 1998; Boege and Marquis  
447 2005; Cash and Fulbright 2005; Kariyat et al. 2017). Thus, the lower spine density of *U.*  
448 *europaeus* seedlings in the invaded range could result from the release of a variety of

449 herbivores, not only big vertebrate herbivores, which prefer to feed on adult than young plants  
450 (Boege and Marquis 2005).

451 From an applied perspective, the results of this study suggest that land managers can  
452 take advantage of the high palatability and (likely) reduced effectiveness of physical defenses  
453 in plants of *U. europaeus* in Chile to use domestic animals, especially goats, as a measure to  
454 control this invasive shrub. The intensive use of goats to control *U. europaeus* resprouts is an  
455 effective management practice widely used in New Zealand (Radcliffe 1985, 1986; Howe et al.  
456 1988; Popay and Field 1996); however, this control procedure requires clipping of adult plants  
457 and numerous goats to eat the resprouts. On the other hand, a high invasive potential could  
458 be expected for other *Fabaceae* species with photosynthetic thorns, such as *Ulex galli*, *U. minor*  
459 (Stokes et al. 2003) and *Prosopis kuntzei* (Böcher 1977), because they could also be partially  
460 released from the growth-defense tradeoff in the invaded ecosystems. Finally, we have  
461 captured only a subset of the geographic distribution of the study species in its native and  
462 invaded range, which calls for caution in extrapolating our results to other geographic areas.  
463 Some traits, such as the contents of lignin and cellulose in the plant cell wall, may respond to  
464 abiotic conditions, such as drought, high and low temperatures (Le Gall et al., 2015; Liu et al.  
465 2018). Sampling populations of *U. europaeus* from areas with different climatic conditions,  
466 within both native and invaded ranges, would broaden our understanding of the contribution of  
467 the ERH to explain its invasion patterns and the underlying plant traits.

468

## 469 **Conclusions**

470 This study evidences the role of spinescence traits in the invasiveness of an aggressive  
471 and successful thorny shrub, *U. europaeus*. In particular, it shows that in the invaded range

472 (Chile) adult plants of *U. europaeus* shift resource allocation, increasing the amount of  
473 photosynthetic tissue (thorn size and biomass) and decreasing the investment in the defensive  
474 components of thorns by producing less fibrous, slenderer and easier to bend thorns. The latter  
475 may translate into reduced effectiveness against herbivores. In the early stages of  
476 establishment (seedling stage), lower investment in spines and increased growth in the invaded  
477 range may partly explain the invasion success of *U. europaeus*, according to the ERH. As far  
478 as we know, this is the first report of such modification in thorns or spines of a woody species  
479 between invaded and native ranges. Whereas higher investment in growth in the invaded range  
480 (larger thorns) may provide competitive advantages over native plants, higher investment in  
481 defense in the native range (stronger, more fibrous, and robust thorns) may better protect *U.*  
482 *europaeus* against a higher herbivory pressure. The present study illustrates the need of  
483 refinement of ERH predictions to consider additional functions of plant antiherbivore defenses,  
484 such as photosynthesis in the case of *U. europaeus* thorns.

485

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Table 1. Linear Mixed Models assessing the effects of distribution range (native vs. invaded, fixed factor) on branch length-standardized thorn length (L, cm), thorn width (W, mm), L:W ratio, thorn density (number of thorns per cm of branch segment) and thorn biomass allocation (thorn:branch segment mass ratio) in adult plants of *Ulex europaeus* from the invaded and the native range. Population was entered as random factor.

	<i>F</i> ( <i>df</i> = 1)	P
<i>Thorn length (L)</i>	54.04	<b>&lt;0.0001</b>
<i>Thorn width (W)</i>	11.72	<b>0.0011</b>
<i>L:W</i>	35.40	<b>&lt;0.0001</b>
<i>Thorn density</i>	1.21	0.3334
<i>Thorn biomass allocation</i>	19.13	<b>&lt;0.0001</b>



Table 2. Linear Mixed Models assessing the effect of distribution range (invaded vs. native, fixed factor), branch length (covariate) and their interaction on thorn length (L, cm), thorn width (W, mm), L:W ratio, thorn density (number of thorns per cm of branch segment) and thorn biomass allocation (thorn:branch segment mass ratio) in *Ulex europaeus*. Population was entered as random factor.

	<i>Distribution Range (R)</i>		<i>Branch length (BL)</i>		<i>R x BL</i>	
	F ( <i>df</i> = 1)	<i>P</i>	F ( <i>df</i> = 1)	<i>P</i>	F ( <i>df</i> = 1)	<i>P</i>
<i>Thorn length (L)</i>	10.46	<b>0.0023</b>	22.63	<b>&lt;0.0001</b>	1.12	0.2946
<i>Thorn width (W)</i>	3.67	0.0618	3.02	0.0882	1.07	0.3062
<i>L:W</i>	5.63	<b>0.0213</b>	14.40	<b>0.0004</b>	0.19	0.6647
<i>Thorn density</i>	3.17	0.0818	24.43	<b>&lt;0.0001</b>	5.43	<b>0.0235</b>
<i>Thorn biomass allocation</i>	12.19	<b>0.0009</b>	4.13	<b>0.0467</b>	5.79	<b>0.0193</b>

Table 3. Content (%) of lignin, acid detergent fiber (ADF), neutral detergent fiber (NDF), cellulose and dry matter digestibility (DMD) in *Ulex europaeus* thorns from the invaded (Chile) and the native range (Spain). Results of the Linear Mixed Models (LMM) assessing the effect of distribution range (invaded vs native) as fixed factor are also shown. Population was entered as random factor.

	<i>Mean ± SE</i>		<i>LMM</i>	
	Invaded range	Native range	F	<i>P</i>
<i>Lignin</i>	16.2 ± 1.9	21.7 ± 2.1	2.10	0.2236
<i>ADF</i>	34.6 ± 3.4	45.1 ± 1.4	7.35	0.0543
<i>NDF</i>	45.5 ± 2.2	58.8 ± 1.9	10.70	<b>0.0291</b>
<i>Cellulose</i>	28.9 ± 2.6	46.7 ± 2.3	16.99	<b>0.0131</b>
<i>DMD</i>	47.0 ± 2.1	38.3 ± 2.0	3.34	0.1384

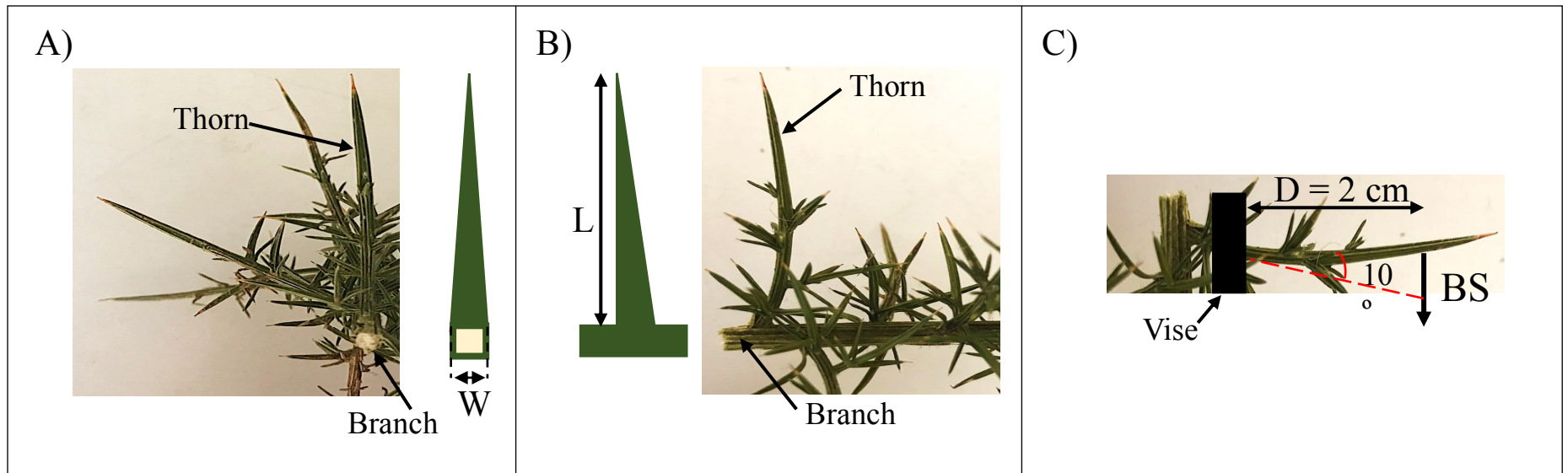


Fig. 1. Proximal (A) and lateral (B) view of a branch segment of *Ulex europaeus* (picture) complemented with a basic scheme of a single thorn showing some spinescence traits measured: thorn length (L) and, thorn base width perpendicular to the branch (W). Scheme showing the bending test applied to the thorns (C). D: the displacement from the vise to the point where the weight is applied in the bending test; BS: bending strength (flexural resistance, see text for additional details).



Fig. 2. Seedlings of *U. europaeus* from the invaded (Chile, left panel) and the native range (Spain, right). Note that the seedling from Spain developed more spines than that from Chile.

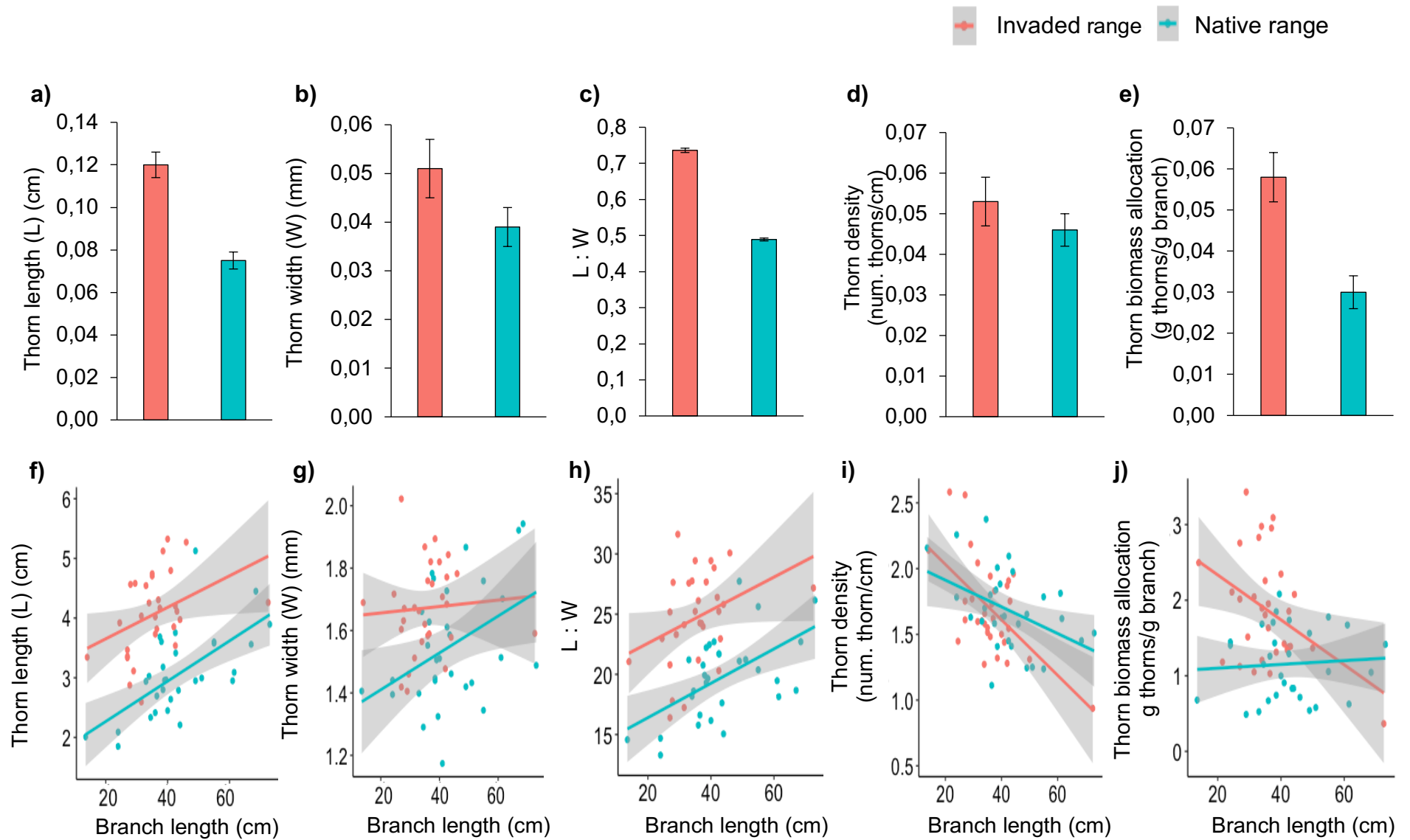


Fig. 3. Mean ( $\pm$  SE) branch length-standardized spinescence variables thorn length (L), thorn width (W), L:W (thorn slenderness), thorn density and thorn biomass allocation (**a**, **b**, **c**, **d** and **e**, respectively) and the linear regressions of the same spinescence variables (no

branch length-standardized) as a function of branch length (**e**, **f**, **g**, **h**, **i** and **j**, respectively) in *Ulex europaeus* individuals from the invaded and native ranges (Chile and Spain, respectively). Statistical results are shown in Table 1 and 2.

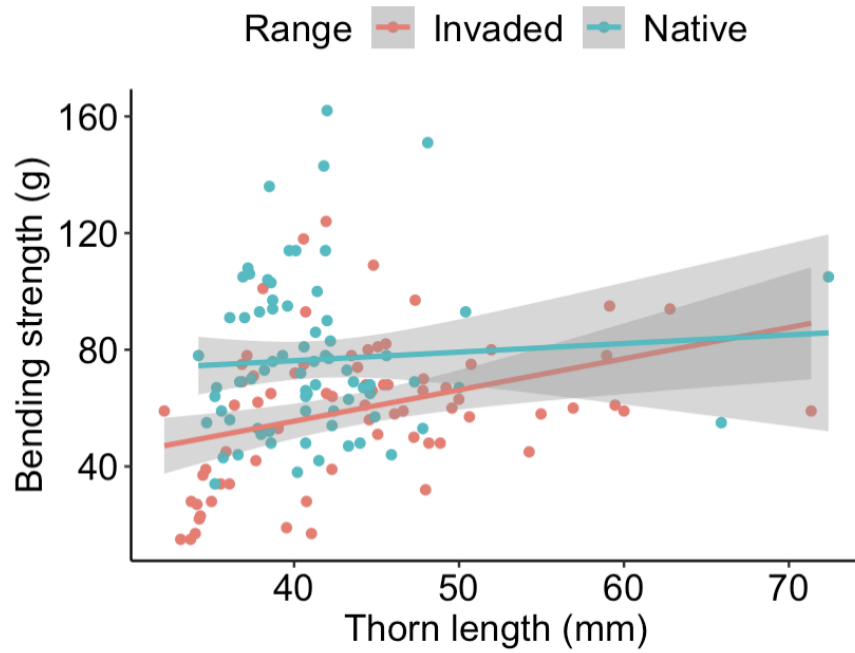


Fig. 4. Linear regression for bending strength as a function of thorn length in *Ulex europaeus* from invaded and native distribution ranges (Chile and Spain, respectively; Range). Bending strength was significantly influenced by Range, Thorn length and their interaction (statistical results in the text).

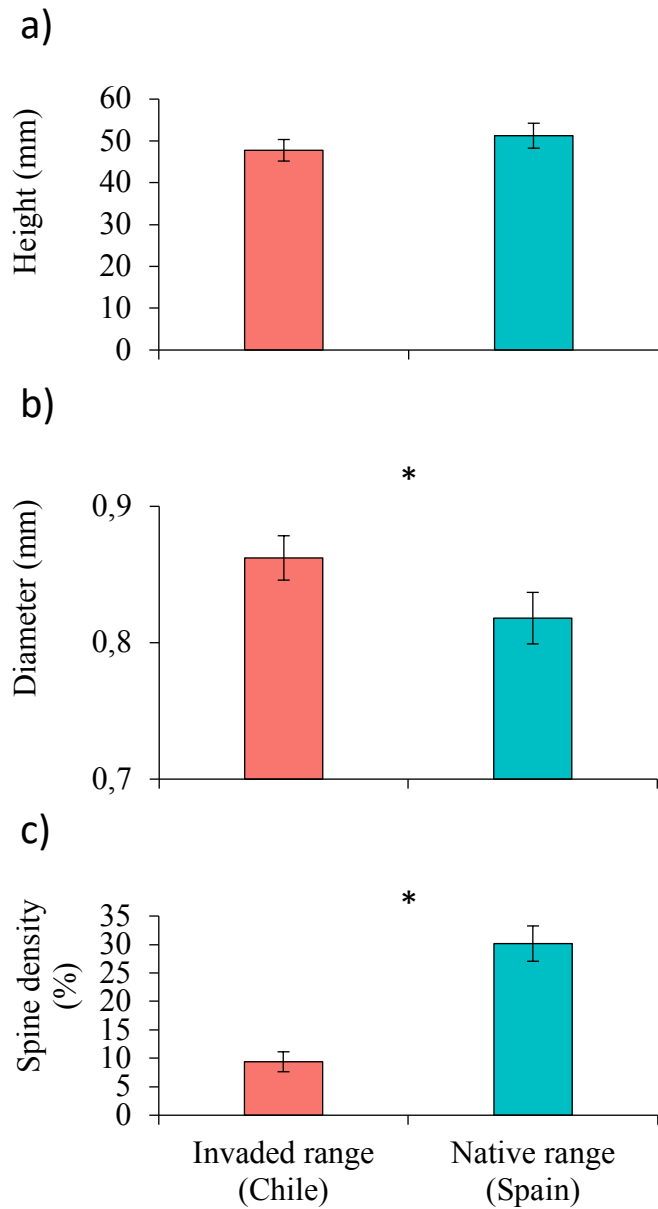


Fig. 5. Mean ( $\pm$  SE,  $n = 120$ ) of the variables: a) height (mm), b) stem diameter (mm) and c) spine density (%) of *Ulex europaeus* seedlings from the invaded and the native distribution range (Chile and Spain, respectively; Range). Asterisks indicate significant differences between ranges after a linear mixed model (population as random factor) for each variable.