



Article

# Prenatal Manganese Exposure and Long-Term Neuropsychological Development at 4 Years of Age in a Population-Based Birth Cohort

Ainara Andiarena <sup>1,2,\*</sup>, Amaia Irizar <sup>2</sup>, Amaia Molinuevo <sup>2</sup>, Nerea Urbietta <sup>2</sup>, Izaro Babarro <sup>2</sup>, Mikel Subiza-Pérez <sup>2</sup>, Loreto Santa-Marina <sup>2,3,4</sup>, Jesús Ibarluzea <sup>2,3,4</sup>  and Aitana Lertxundi <sup>2,3</sup>

<sup>1</sup> Faculty of Psychology of the University of the Basque Country, 20018 San Sebastian, Spain

<sup>2</sup> Biodonostia Health Research Institute, Group of Environmental Epidemiology and Child Development, Doctor Begiristain, s/n, 20014 San Sebastian, Spain; a-irizarloibide@euskadi.eus (A.I.); au-molinuevo@euskadi.eus (A.M.); n-urbietamacazaga@euskadi.eus (N.U.); i-babarro@euskadi.eus (I.B.); m-subiza@euskadi.eus (M.S.-P.); ambien4ss-san@euskadi.eus (L.S.-M.); mambien3-san@euskadi.eus (J.I.); aitana.lertxundi@ehu.eus (A.L.)

<sup>3</sup> Spanish Consortium for Research on Epidemiology and Public Health (CIBERESP), Instituto de Salud Carlos III, C/Monforte de Lemos 3-5, 28029 Madrid, Spain

<sup>4</sup> Health Department of the Basque Government, Sub-directorate of Public Health of Gipuzkoa, 20013 San Sebastian, Spain

\* Correspondence: ainara.andiarena@ehu.eus

Received: 22 January 2020; Accepted: 28 February 2020; Published: 4 March 2020



**Abstract:** *Background:* Manganese (Mn) is an essential micronutrient for humans, the diet being the main source of exposure. Some epidemiological studies describe a negative association between prenatal Mn and later neuropsychological development, but results are inconsistent. The aim of this study was to explore the association between prenatal Mn exposure and neuropsychological development assessed at 4 years of age. *Methods:* Study subjects were 304 mother-child pairs from the Gipuzkoa cohort of the INMA (Environment and Childhood) Project. Mn was measured in newborns' hair. Children's neuropsychological development was assessed at 4 years of age using the McCarthy Scales of Children's Abilities. Multivariate linear regression models were built. Stratified analysis by sex was performed. Generalized additive models were used to assess the shape of the relation. *Results:* The median Mn concentration in newborns' hair was 0.42 µg/g (95% CI = 0.38, 0.46). The association between Mn levels and the neuropsychological development was not statistically significant for the general cognitive scale ( $\beta$  [95% CI] = 0.36 [−5.23, 5.95]), motor scale ( $\beta$  [95% CI] = 1.9 [−3.74, 7.55]) or any of the other outcomes. No sex-specific pattern was found. The best shape describing the relationship was linear for all the scales. *Conclusion:* Our results suggest that prenatal Mn concentrations measured in newborns' hair do not affect cognitive or motor development at 4 years of age in boys or in girls at the observed Mn levels.

**Keywords:** hair Mn; in utero exposure; prenatal; neurodevelopment; prospective

## 1. Introduction

Manganese (Mn) is an essential trace element [1] and an essential micronutrient that plays a critical role in normal growth and development, having an impact on body growth, immune function, enzymatic regulation reactions, bone growth and metabolism. It is also needed for proper fetal development [2–4]. Nevertheless, overexposure to Mn can also be detrimental to health and accumulation of Mn in the brain may have neurotoxic effects [2,5,6].

Mn is naturally found in the environment [7], and for the general population, food is the main source of exposure, since it is present in many foods, especially legumes, cereals, nuts and tea [8]. Air is also a major environmental source of Mn exposure [6,9–12]. Its presence in the air is most closely linked to the iron and steel industry and the burning of fossil fuels but is also related to its use in the chemical industry. In addition, atmospheric Mn originates from traffic pollution, since it is employed as an additive to improve the octane rating of gasoline [13] and Mn derived from agricultural activities [14,15] should also be considered. The intake of water has been observed to be another source of exposure, especially in regions with high Mn levels in drinking water, such as Bangladesh [16], Japan and Australia [17]. Lastly, tobacco smoke also plays a role as a source of Mn [8], because of the high Mn concentrations in all green leaves [18], and the presumably high absorption of Mn in the lungs. Mothers who smoke have been shown to have increased Mn concentrations in umbilical cord blood [19].

Interactions between Fe and Mn are noteworthy when assessing the health effects of Mn intake. While plasma Fe overload significantly decreases the uptake of Mn across the blood-brain barrier, Fe deficiency is associated with increased central nervous system (CNS) burden of this element [20]. The individuals most vulnerable to Fe deficiency are women who are pregnant or of childbearing age, infants and small children [21].

Early research on Mn health effects focused on adult populations highly exposed to Mn mainly due to occupational exposure [8,22]. Reference levels exist for adult population [23] while age-specific exposure limits are not established. More recently, epidemiologic studies have moved toward the examination of potential health effects of lower-level Mn exposure in children through air and drinking water [23].

Many cross-sectional studies conducted in infant populations found negative effects of Mn on general cognitive development [10,24–27]. Negative associations have also been found for specific cognitive functions such as verbal abilities [25,26] or memory [12] and for psychomotor development [28]. Some studies have observed an inverted U-shape relationship between Mn and neurodevelopment [6]. Nevertheless, other studies have failed to find any associations [11,27,29,30].

Prospective studies have had inconsistent findings [23]. Regarding prenatal exposure, Lin et al. [31] describe a negative effect of umbilical cord Mn levels on overall, cognitive and language neurodevelopment at 2 years. Chung et al. [1] found an inverted U-shape association between maternal blood Mn levels at delivery and neurodevelopment at 6 months. Other studies have found a negative effect at 9 months of age but not at later stages [14,32] and Mora et al. [15] found that higher prenatal Mn levels, measured in dentine of deciduous teeth, were associated with poorer behavioral outcomes in school-age boys and girls, but also with better motor function, memory and/or cognitive abilities in school-age boys. Limited and inconsistent evidence of sex-specific neurological effects have been found, generally indicating greater effects in girls both negative [25,27] and positive [33], but also some in boys [34].

Despite many studies having analyzed the association between prenatal Mn levels and longer-term neurodevelopment, there is little consistency in the results [23]. Therefore, after having characterized intrauterine Mn exposure, using Mn levels in newborns' hair as a biomarker of prenatal exposure [19] and considering the high vulnerability of the developing brain, in the present study, we sought to assess whether newborns' hair Mn levels are associated with neuropsychological development at 4 years of age. Further, we investigated the linearity of the association and possible sex differences.

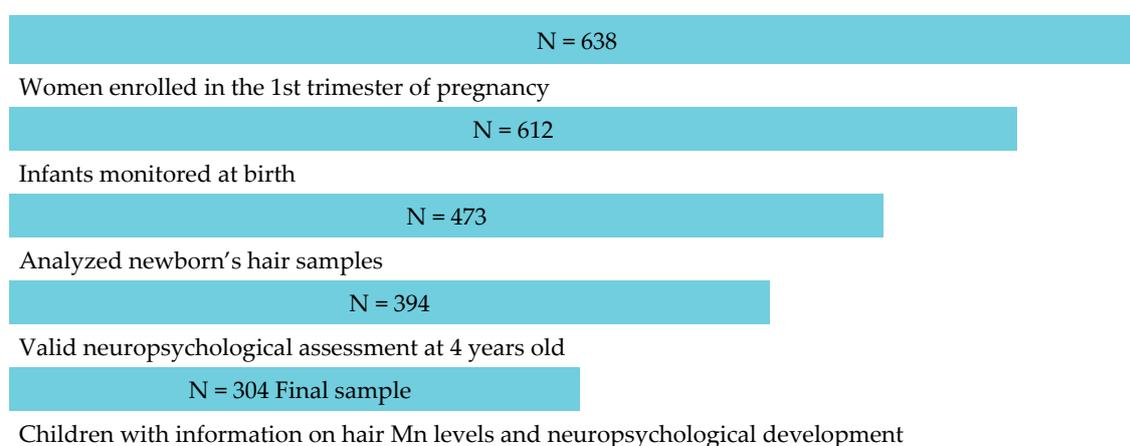
## 2. Materials and Methods

### 2.1. Study Design and Participants

The study participants were mother-children couples of the INMA-Gipuzkoa cohort. The INMA project (from the Spanish terms for environment and childhood, *Infancia y Medio Ambiente*) (<http://www.proyectoinma.org>) is a prospective multicenter study of mother and child birth cohorts [35].

The main objective of this project is to evaluate the relationship between prenatal and postnatal exposure to environmental pollutants and children's health from the fetal stage to adolescence [35].

In Gipuzkoa, in the 2006–2008 period, 638 women were recruited in the first trimester of pregnancy. The women were followed during pregnancy and their children were subsequently followed from birth to the fourth year of life. From recruitment to the 4 years old follow up stage there was a 10.65% of participants lost ( $n = 68$ ) due to: foetal death ( $n = 4$ ), miscarriage ( $n = 10$ ), death ( $n = 2$ ), withdrew ( $n = 44$ ) and contact lost ( $n = 8$ ). Of the 612 births, hair samples were collected in 437 newborns (Irizar et al., 2019). Of the 394 children who had a valid neuropsychological evaluation at 4 years of age, 304 also had information on Mn levels in their hair at birth. Therefore, this study included 304 participants for whom we had information on Mn levels in newborn hair and neuropsychological development at 4 years of age (Figure 1). The study was approved by the Ethics Committee of Donostia Hospital (Gipuzkoa) and all mothers signed informed consent forms before inclusion.



**Figure 1.** Flow of participants through the study and final sample.

## 2.2. Manganese in Hair

A 50 mg sample of newborn hair was taken from the area closest to the occipital region. The samples were washed twice by ultrasonic cleaning and rinsed with ultrapure water. Intra-batch and inter-batch accuracies of reactive targets were estimated at 0.95% and 6.79%, respectively. The accuracy in the concentration range of 4.75–5.68  $\mu\text{g/g}$  ( $n = 11$ ) was 101.0% (certified reference value of 5.2  $\mu\text{g/g}$ ). The analysis was carried out in the Department of Legal Medicine and Toxicology of the University of Granada [36–38]. The method has been reported in detail by Irizar et al. [19].

## 2.3. Children's Neuropsychological Development

A standardized version of the McCarthy Scales of Children's Abilities (MSCA) was used to assess the children's cognitive and psychomotor development at 4 years of age [39]. The Basque version of the instrument (MSCA-E) was administered to children whose first language was Basque [40]. The MSCA comprises 18 subtests that yield standardized test scores for six subscales: verbal, perceptual-performance, quantitative, memory and motor scales. The sum of scores on the first three scales provides a General Cognitive Index [40]. In addition, a previously validated Executive Function scale was analyzed [41]. MSCA raw scores were standardized to a mean of 100 and a standard deviation (SD) of 15. All testing was performed by a trained neuropsychologist under appropriate assessment conditions. The neuropsychologist was blinded to any information concerning the child, including hair Mn levels. A total of 4.7% ( $n = 7$ ) of the test results were excluded from the final analysis due to the poor quality of the child's assessment (e.g., lack of collaboration or evident tiredness) or to the diagnosis of neurodevelopmental disorders.

#### 2.4. Potential Confounding or Predictor Variables

Data on maternal characteristics, including socio-demographic, environmental and lifestyle factors were collected through two questionnaires administered in face to face by interviews in the first and third trimesters of pregnancy.

Information on maternal variables were: education level (up to primary, secondary, university), body mass index (BMI, kg/m<sup>2</sup>) before pregnancy (low weight [ $<18.5$ ], healthy [ $18.5- < 25$ ], overweight [ $25- < 30$ ], obese [ $\geq 30$ ]) age at conception (years), parity ( $0/\geq 1$ ), alcohol and smoking consumption during pregnancy (yes/no), and maternal ferritin levels analyzed in the first trimester of pregnancy [(mean (SD) 13.3 (1.5)] by fluoroimmunoassay [42]. Maternal exposure to PM<sub>2.5</sub> throughout pregnancy and in each trimester was also assessed [43].

Information related to the child's sex, birthweight (low birth weight  $< 2500$  g yes/no), season of birth, sibling order (first/other) and length of gestation (preterm  $< 37$  weeks yes/no) were retrieved from clinical records. In a subsequent interview when the child was 14 months information on main caregiver, duration of breastfeeding (weeks) nursery attendance, and child second-hand smoke (yes/no) exposure was collected. When the child was 2 years of age was assessed the quality of the family context using the Etxadi-Gangoiti Scale [44]. The subscale of Stimulation of Cognitive and Linguistic Development (SCLD) was also selected because it was strongly associated with cognitive development [45]. The child's age was also recorded at the time of the neuropsychological evaluation at 4 years of age.

#### 2.5. Statistical Analyses

First, a descriptive analysis of the population was performed. Qualitative variables were described in terms of percentages and quantitative variables using mean and standard deviation (SD). It was checked that the assumption of normality was met for the McCarthy scales (Supplementary, Figure S1).

In order to decide a priori which potential confounding or predictor variables needed to be included in our models, we drew a directed acyclic graphs (DAGs) according to current knowledge from the scientific literature (Supplementary, Figure S2). Results identified a minimally sufficient adjustment set that included maternal age, smoking during pregnancy, PM<sub>2.5</sub> and ferritin levels.

In consequence, minimally adjusted regression models were constructed. Then, the model with the best Akaike information criterion (AIC) fit was selected. Thus, final models were adjusted for age at the time of the neuropsychological assessment, sex, sibling order, nursery attendance and maternal education.

Linear regression models were used to analyze the association between hair Mn and all the neuropsychological outcomes. First, hair Mn was included in a continuous manner and later it was analyzed in tertiles (taking lower values as the reference). Subsequently, in order to analyze possible sex differences, analysis was stratified by sex.

Due to the missing data on the Etxadi-Gangoiti Scale scores (15.13%), the models were not adjusted for this variable. On the other hand, sensitivity analyses were performed.

Generalized additive models (GAMs) were fitted to check the linearity of the relationships between Mn and the scales by means of natural cubic splines with one internal knot. In all cases, the linear model was the best for describing the relationship between exposure and outcome.

Statistical analyses were carried out with R, version 3.6.1 (R Core Team, Vienna, Austria).

### 3. Results

An initial analysis was conducted to assess differences in the variables studied between our participants and the whole INMA Gipuzkoa cohort. The only statistically significant differences observed were in maternal age at conception, participating mothers being significantly older ( $p = 0.029$ ) (Supplementary, Table S1).

We analyzed data from the 304 children for whom hair samples taken at delivery were available and who had completed a valid neuropsychological assessment (Figure 1). The median Mn concentration in hair was 0.42  $\mu\text{g/g}$  (95% CI = 0.38, 0.46). The geometric mean (GM) was 0.27  $\mu\text{g/g}$  (95% CI = 0.24, 0.31).

Table 1 shows the characteristics of the study population. Regarding maternal characteristics, the women's mean age at conception was 31.7 years, and 48% of them had university education, 76% had a healthy BMI before pregnancy, 55% were primiparous, 6.91% consumed some alcohol during pregnancy and 10% smoked during pregnancy. Regarding children, 50% were girls, 3.6% had a low-birth weight, and 2.6% were born preterm. They were breastfed for a mean of 30 months. At 12 months of age, 46% attended nursery and 14% were exposed to secondhand smoke. At the 4-year follow up, the children's mean age was 4.48.

**Table 1.** Characteristics of the study population.

		N (%)	Mean (sd)	Number of Missing Data
<b>Maternal characteristics</b>				
Age			31.67 (3.3)	0(0%)
Educational level	Primary	38 (12.5%)		1 (0.33%)
	Secondary	120 (39.47%)		
	University	145 (47.7%)		
Body mass index, kg/m <sup>2</sup>	Underweight (<18.5)	13 (4.28%)		0 (0%)
	Normal weight (>18.5–25)	231 (75.99%)		
	Over weight (26–30)	45 (14.8%)		
	Obese ( $\geq 30$ )	15 (4.93%)		
Parity	0	167 (54.93%)		0 (0%)
	$\geq 1$	116 (45.07%)		
Alcohol consumed in pregnancy	No	271 (89.10%)		2 (0.66%)
	Yes	21 (6.91%)		
Smoking in pregnancy	No	266 (87.5%)		7 (2.3%)
	Yes	31 (10.20%)		
Ferritin level ( $\mu\text{g/L}$ )			34.54 (26.11)	1 (0.33%)
PM <sub>2.5</sub> in pregnancy (unit)			16.98 (2.43)	22 (7.24%)
<b>Child characteristics</b>				
Sex	Male	153 (50.33%)		0 (0%)
	Female	151 (49.67%)		
Low birth weight (<2500 g)	No	292 (96.05%)		1 (0.33%)
	Yes	11 (3.62%)		
Season of birth	Winter	87 (28.62%)		0 (0%)
	Autumn	50 (16.45%)		
	Spring	88 (28.95%)		
	Summer	79 (25.99%)		
Sibling order	First	169 (55.59%)		0 (0%)
	Other	135 (44.41%)		
Preterm (<37 weeks)	No	294 (96.71%)		8 (2.63%)
	Yes	8 (2.63%)		
Breastfed (weeks)			29.98 (20.20)	12 (3.95%)
Nursery (14 months)	No	150 (49.34%)		13 (4.28%)
	Yes	141 (46.38%)		
Passive smoking (14 months)	No	262 (86.18%)		0(0%)
	Yes	42 (13.82%)		
Etxadi-Gangoiti Scale (at 2 years)			73.71 (9.78)	46 (15.13%)
Age at assessment (at 4 years)			4.48 (0.12)	3 (0.99%)

Regression models did not identify any statistically significant associations between newborns' hair Mn levels and neurodevelopment at 4 years of age (Table 2). Regression models for cognitive outcomes showed that for each 1  $\mu\text{g/g}$  increase in Mn, scores increased on the general cognitive scale by 0.36 [95% CI = -5.23 to 5.95], on the perceptive manipulative scale by 2.24 [95% CI = -3.27 to 7.76], on the executive function scale by 0.1 [95% CI = -5.46 to 5.65], on the general motor scale by 1.9 [95% CI = -3.74 to 7.55], on the gross motor scale by 3.57 [95% CI = -2.1 to 9.23], and decreased on the verbal scale by 0.93 [95% CI = -6.71 to 4.85], on the quantitative scale by 0.43 [95% CI = -5.92 to 5.06], on the memory scale by 0.39 [95% CI = -6.16 to 5.38], and on the fine motor scale by 0.94 [95% CI = -6.23 to 4.36]. When analyzing the association in tertiles, regression models again showed no significant associations between hair Mn and the outcomes of interest (Table 2).

**Table 2.** Linear regression models for the association between Mn and the McCarthy scale scores. Change in McCarthy scale score for 1 µg/g increase in Mn and by tertiles of the Mn distribution.

		Beta (CI 95%)						
		For a 1-Point Increase in Mn Levels	R <sup>2</sup>	AIC	Tertile 1 [0.0065–0.2316]	Tertile 2 (0.2316–0.4093]	Tertile 3 (0.4093–2.1658]	p for Trend
General cognitive	Model 1	−1.05 (−6.72, 4.62)	0.011	2272.542	1 (Ref.)	0.18 (−4.26, 4.61)	−0.48 (−4.97, 4.01)	0.834
	Model 2	0.36 (−5.23, 5.95)	0.144	2128.497	1 (Ref.)	0.53 (−3.85, 4.9)	0.12 (−4.29, 4.54)	0.954
Verbal	Model 1	−1.78 (−7.45, 3.88)	0.014	2271.846	1 (Ref.)	0.86 (−3.58, 5.29)	0.48 (−4.01, 4.97)	0.833
	Model 2	−0.93 (−6.71, 4.85)	0.100	2145.329	1 (Ref.)	1.58 (−2.94, 6.1)	0.74 (−3.82, 5.3)	0.744
Perceptive–manip.	Model 1	0.77 (−4.75, 6.28)	0.016	2257.169	1 (Ref.)	−0.41 (−4.72, 3.9)	−0.48 (−4.84, 3.89)	0.829
	Model 2	2.24 (−3.27, 7.76)	0.120	2121.771	1 (Ref.)	−0.57 (−4.9, 3.75)	0.23 (−4.13, 4.6)	0.919
Quantitative	Model 1	−2.21 (−7.88, 3.47)	0.009	2272.823	1 (Ref.)	−0.56 (−5, 3.88)	−1.07 (−5.57, 3.42)	0.638
	Model 2	−0.43 (−5.92, 5.06)	0.160	2119.094	1 (Ref.)	−0.11 (−4.41, 4.19)	−0.09 (−4.43, 4.24)	0.966
Memory	Model 1	−1.3 (−7.19, 4.58)	0.006	2292.962	1 (Ref.)	1.26 (−3.35, 5.86)	0.83 (−3.83, 5.49)	0.726
	Model 2	−0.39 (−6.16, 5.38)	0.132	2144.533	1 (Ref.)	1.68 (−2.83, 6.19)	1.23 (−3.32, 5.78)	0.591
Executive function	Model 1	−1.43 (−7.05, 4.2)	0.014	2267.655	1 (Ref.)	−0.23 (−4.63, 4.17)	−0.27 (−4.73, 4.18)	0.904
	Model 2	0.1 (−5.46, 5.65)	0.120	2125.193	1 (Ref.)	0.57 (−3.78, 4.92)	0.61 (−3.78, 4.99)	0.784
Motor global	Model 1	0.33 (−5.18, 5.84)	0.023	2256.665	1 (Ref.)	0.26 (−4.05, 4.57)	−0.16 (−4.52, 4.2)	0.943
	Model 2	1.9 (−3.74, 7.55)	0.060	2133.311	1 (Ref.)	0.58 (−3.84, 5.01)	1.12 (−3.34, 5.57)	0.621
Gross motor	Model 1	3.25 (−2.44, 8.93)	0.038	2273.539	1 (Ref.)	2.21 (−2.23, 6.66)	2.24 (−2.26, 6.74)	0.326
	Model 2	3.57 (−2.1, 9.23)	0.097	2135.314	1 (Ref.)	3.39 (−1.03, 7.81)	3.42 (−1.04, 7.87)	0.130
Fine motor	Model 1	−2.99 (−8.36, 2.37)	0.013	2241.805	1 (Ref.)	−1.98 (−6.17, 2.21)	−2.66 (−6.9, 1.59)	0.218
	Model 2	−0.94 (−6.23, 4.36)	0.135	2100.425	1 (Ref.)	−2.76 (−6.89, 1.38)	−1.98 (−6.14, 2.19)	0.346

Model 1: adjusted for maternal age, smoking during pregnancy, PM<sub>2.5</sub> and Fe levels in pregnancy. Model 2: adjusted also for age at the time of the test, sex, sibling order, nursery at 12 months and maternal educational level.

Finally, the stratified analysis showed no statistically significant association in either boys or in girls, when analyzing hair Mn levels continuously or in tertiles (Table 3).

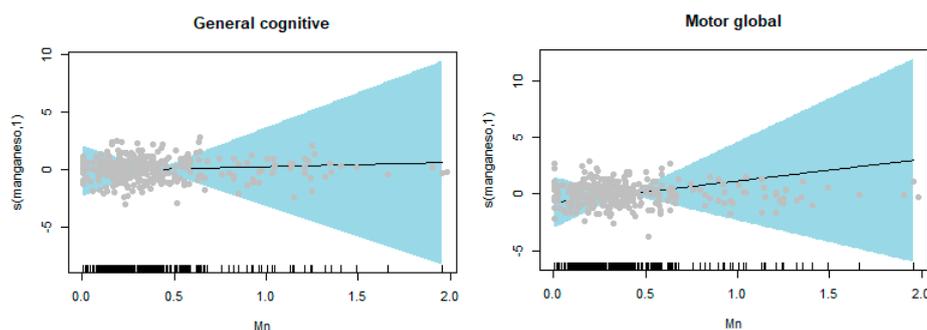
**Table 3.** Linear regression models for the association between Mn and McCarthy scale scores. Change in the McCarthy scale score for 1  $\mu\text{g/g}$  increase in Mn and by tertiles of the Mn distribution, stratified by sex of the child.

		Beta (CI 95%)				<i>p</i> for Trend
		Fora 1 Point Increase in Mn Levels	Tertile 1 [0.0065–0.2316]	Tertile 2 (0.2316–0.4093]	Tertile 3 (0.4093–2.1658]	
General cognitive	Boys	0.01 (−7.63, 7.65)	1 (Ref.)	−0.76 (−7.59, 6.08)	1.17 (−5.43, 7.77)	0.727
	Girls	4.56 (−4.01, 13.14)	1 (Ref.)	1.32 (−4.39, 7.03)	1.77 (−4.2, 7.75)	0.551
Verbal	Boys	−1.53 (−9.47, 6.42)	1 (Ref.)	1.61 (−5.49, 8.71)	3.01 (−3.85, 9.87)	0.384
	Girls	2.68 (−6.35, 11.71)	1 (Ref.)	1.14 (−4.86, 7.14)	0.65 (−5.63, 6.93)	0.826
Perceptive–manip.	Boys	3.86 (−3.85, 11.58)	1 (Ref.)	−2.7 (−9.61, 4.21)	0.48 (−6.19, 7.16)	0.892
	Girls	2.9 (−5.53, 11.33)	1 (Ref.)	0.86 (−4.74, 6.46)	1.76 (−4.1, 7.62)	0.552
Quantitative	Boys	−3.53 (−10.66, 3.59)	1 (Ref.)	−2.15 (−8.54, 4.24)	−2.66 (−8.83, 3.51)	0.392
	Girls	5.82 (−3.06, 14.69)	1 (Ref.)	1.8 (−4.08, 7.69)	4.33 (−1.83, 10.49)	0.166
Memory	Boys	−1.1 (−8.35, 6.14)	1 (Ref.)	0.96 (−5.52, 7.43)	2.56 (−3.69, 8.82)	0.417
	Girls	4.09 (−5.54, 13.72)	1 (Ref.)	1.84 (−4.56, 8.25)	2.37 (−4.32, 9.07)	0.475
Executive function	Boys	−2.73 (−10.62, 5.16)	1 (Ref.)	−1.93 (−9.01, 5.15)	−1.11 (−7.95, 5.72)	0.745
	Girls	6.12 (−2.08, 14.31)	1 (Ref.)	2.34 (−3.1, 7.79)	4.1 (−1.6, 9.79)	0.153
Motor global	Boys	4.06 (−3.61, 11.73)	1 (Ref.)	−0.27 (−7.16, 6.62)	2.03 (−4.62, 8.68)	0.547
	Girls	0.05 (−8.56, 8.67)	1 (Ref.)	2.18 (−3.52, 7.88)	2.29 (−3.67, 8.26)	0.436
Gross motor	Boys	4.05 (−3.85, 11.96)	1 (Ref.)	3.24 (−3.83, 10.31)	3.98 (−2.85, 10.81)	0.248
	Girls	2.3 (−6, 10.59)	1 (Ref.)	5.15 (−0.27, 10.58)	4.53 (−1.14, 10.21)	0.105
Fine motor	Boys	1.82 (−5.46, 9.1)	1 (Ref.)	−3.9 (−10.39, 2.59)	−1.19 (−7.46, 5.07)	0.702
	Girls	−2.39 (−10.78, 5.99)	1 (Ref.)	−2.22 (−7.78, 3.34)	−1.39 (−7.21, 4.43)	0.617

Model 2\*: also adjusted for age at the time of the test, sibling order, nursery at 12 months and maternal education.

In the sensitivity analysis, the results did not vary substantially. Specifically, the inclusion of quality of family context had very little influence, if any, on the regression coefficients for the association between hair Mn and child outcomes, suggesting that this variable was not a significant confounder in our sample (Supplementary Table S2).

Results for the GAM analysis showed linearity in all cases (Supplementary Figure S3). To illustrate this, the models for the General Cognitive and General motor scales are plotted in Figure 2.



**Figure 2.** Generalized additive models for the general cognitive and general motor scales.

#### 4. Discussion

In our study, no associations were found between children's hair Mn levels at birth and neurodevelopment at 4 years of age and. Linear associations were observed between children's hair Mn levels at birth and all the cognitive and psychomotor scales and no significant patterns of sex-specific associations were found.

Relatively few studies have prospectively examined potential associations between prenatal exposure to Mn and neurodevelopment [1,14,15,31,32,46]. Moreover, different matrices have been used to assess Mn concentrations. Our results are in agreement with those of Takser et al. [32], who found

no correlation between Mn measures at birth (from hair and umbilical cord blood samples) and the McCarthy scales general cognitive index at 3 or 6 years. Nonetheless, the concentration of Mn found in hair in their study (Mn GM of 0.75 µg/g [range 0.05–13.33]) was higher than that found in our cohort. To our knowledge, it is the only other study that has measured Mn in newborns' hair.

Other studies that have analyzed prenatal Mn using dentin as a biomarker of chronic exposure have also found no association with cognitive or psychomotor development during childhood [14,15]. On the other hand, while Erikson et al. [46] found no differences in cognitive ability or psychomotor development, they reported that prenatal Mn levels were positively associated with behavioral outcomes: higher levels of disinhibition (36 months), impulsivity (4.5 years), externalizing and internalizing problems (1st and 3rd grades) and disruptive behaviors (3rd grade).

Regarding studies that have used measures of short-term prenatal exposure, Chung et al. [1] found that associations between maternal blood manganese and neurodevelopment followed an inverted U-shape dose-response curve, but in a recent study conducted with our cohort, no significant association has been found between maternal blood Mn and child cognitive or psychomotor development at 14 months of age [47]. Further, as mentioned in the introduction, other authors have found a negative association between prenatal Mn levels (umbilical cord) and neurodevelopment [31].

Findings are inconsistent across biomarkers regarding associations with mental and psychomotor development [30]. In our study, Mn exposure was measured through hair Mn concentration which, among biomarkers of Mn exposure, is considered a good indicator of chronic environmental exposure [48]. In a recent systematic review, authors concluded that with the evidence available it is not possible to draw definitive conclusions concerning adverse effects of early-life Mn exposure on neuropsychological development, and highlighted the need for more prospective studies to establish a causal link [23].

Our results show linear associations with all the cognitive and psychomotor scales. Similarly, studies using chronic exposure measures (hair or dentine Mn levels) also indicate a linear association between prenatal Mn and long-term cognitive outcomes [15,49]. Gunier et al. [14] described a linear association with cognitive development but a non-linear association with psychomotor development, while other authors proposed an inverted U-shape association between maternal blood Mn and neurodevelopment [1] and some studies have suggested that there may be a threshold value above which there is an association with neurodevelopment [1,31].

In our study, we observed no significant sex-specific patterns of association. Our results suggest a trend in the association between Mn and cognitive scales that is negative for boys and positive for girls. In contrast, the association for the motor scale seems positive for boys but not for girls. Similarly, Gunier et al. [14] did not find an interaction by sex between prenatal Mn levels and mental or psychomotor development at 6, 12 or 24 months. On the other hand, Takser et al. [32] observed a negative association with hand skills in boys and Mora et al. [15] found that higher prenatal Mn levels were associated with better cognitive and motor outcomes in boys only. A possible explanation for the differences found in previous studies between boys and girls are the associated mechanisms of absorption and excretion and interactions with hormones and neurotransmitters [50]. Nevertheless, more epidemiological and animal studies are needed to characterize these potential biological differences between males and females.

The inconsistencies between our results and some previous longitudinal studies could be explained by various factors. Firstly, the use of different matrices to measure Mn exposure. Results are generally discordant between studies that have used blood Mn, hair Mn or tooth dentin Mn levels as a biomarker for Mn exposure [23]. Studies of infants and toddlers most frequently measure the Mn in hair, cord blood, tooth enamel, maternal or child blood or dentin and findings are inconsistent across these biomarkers in relation to associations with mental and psychomotor development [30]. Besides, differences in neurodevelopmental assessment instruments and ages of evaluation may affect the results. The timing of exposure, during pregnancy, may also be critical for Mn neurotoxicity, because the susceptibility of the brain to toxic insults is known to change during different phases of neurodevelopment [5].

Finally, the differences in the Mn concentrations found between studies hinder comparability. Our Mn concentrations are low and could be considered to be within the normal range, in which this compound behaves as an essential element.

This study has some limitations. On the one hand, the sample size, although it is adequate for this type of study, is relatively small and considering the low Mn levels, could be limiting our statistical power. On the other hand, although we were able to assess numerous confounding factors (such as family context, and ferritin or PM2.5 levels), there could be residual confounding in the relationships between Mn exposure and the results of neuropsychological development for which we have been unable to control. In addition, the outcomes studied have been cognitive and psychomotor development and not others possibly related to Mn exposure, such as behavioral outcomes.

Despite its limitations, the present study also has considerable strengths including its longitudinal design, use of comprehensive neurodevelopmental assessments, information on a wide variety of potential confounders, and use of a good indicator of chronic Mn environmental exposure. Considering family contexts have an important influence on children's psychological development [51] and that effects can be masked by this factor [32], we consider that having measured family context is a key strength of this study. As expected, as no association was found between Mn and neurodevelopment, the inclusion of family context did not change the results. Although it is important to consider all the possible socio-environmental confounders [52], as far as we know, no studies have analyzed the family context among the covariates in their analysis. Further, hair Mn level is considered a good indicator of chronic environmental exposure [48] and therefore a good measure for exploring possible long-term neurodevelopmental effects. Previous studies determining Mn levels have more often been in older children [30]. Some authors have suggested that Mn levels remain similar throughout childhood [28,32] given that the main Mn sources do not change, it could be reasonable to compare Mn hair levels during childhood with our data. Doing this, Mn concentrations in this study are seen to be similar to those reported in children's hair in numerous child cohort studies [9,24,27–29,37,53] this suggesting that our measure is reliable.

## 5. Conclusions

In conclusion, this study does not show any significant associations between neuropsychological development at 4 years of age and prenatal exposure to Mn. Further, no consistently different sex-specific patterns were detected. Finally, a linear association is observed between exposure and outcome.

It is possible that the prenatal Mn levels observed in our sample are lower than a threshold at which Mn causes neuropsychological adverse effects. These findings therefore provide information to the body of scientific knowledge about which prenatal Mn levels are likely not detrimental and about the window of susceptibility to Mn exposure. Nevertheless, similar epidemiological studies are necessary in order to improve our knowledge about the safe range of this compound in the prenatal period and the possible sex-specific effects.

**Supplementary Materials:** The following are available online at <http://www.mdpi.com/1660-4601/17/5/1665/s1>, Figure S1: McCarthy scale plots, Figure S2: Directed acyclic graphs to select the covariates included in the analyses, Figure S3: General additive models, Table S1: Sensitivity analysis: differences between our participants and the whole cohort, Table S2: Linear regression models for the association between Mn and the McCarthy scale scores. Change in the McCarthy scales for 1 µg/g increase in Mn and by tertiles of the Mn distribution including Family Context (Etxadi-Gangoiti SCLD subscale).

**Author Contributions:** A.A., A.I., A.M., N.U., I.B., M.S.-P., L.S.-M., J.I. and A.L. have read and agree to the published version of the manuscript. The co-authors had together contributed to the completion of this article. Specifically, it follows their individual contribution: Conceptualization, L.S.-M. and A.L.; Data collection, L.S.-M. and A.A.; Formal analysis, A.M. and A.I.; Methodology, N.U., A.L., A.I. and A.M.; Project administration, L.S.-M. and A.L.; Supervision, A.A.; Writing—original draft, A.A., N.U. and A.L.; Writing—review and editing, A.A., I.B. M.S.-P., and L.S.-M. All authors have read and agreed to the published version of the manuscript.

**Funding:** This study is part of the INMA (INfancia y MedioAmbiente) project. This research was funded by the Spanish Ministry of Health (FIS-FEDER PI060867 and PI09/00090), the Council of Gipuzkoa (DFG15/009), Health

Department of the Basque Government (2013111089), CIBERESP and annual agreements with the municipalities of the study area (Zumarraga, Urretxu, Legazpi, Azkoitia y Azpeitia y Beasain).

**Acknowledgments:** We want to acknowledge all the INMA families for their altruistic participation in all the phases of the follow-up, which has made this research possible.

**Conflicts of Interest:** The authors declare no conflicts of interest.

## References

1. Chung, S.E.; Cheong, H.K.; Ha, E.H.; Kim, B.N.; Ha, M.; Kim, Y.; Hong, Y.C.; Park, H.; Oh, S.Y. Maternal Blood Manganese and Early Neurodevelopment: The Mothers and Children’s Environmental Health (MOCEH) Study. *Environ. Health Perspect.* **2015**, *123*, 717–722. [[CrossRef](#)]
2. Wood, R.J. Manganese and birth outcome. *Nutr. Rev.* **2009**, *67*, 416–420. [[CrossRef](#)]
3. Zota, A.R.; Ettinger, A.S.; Bouchard, M.; Amarasiriwardena, C.J.; Schwartz, J.; Hu, H.; Wright, R.O. Maternal blood manganese levels and infant birth weight. *Epidemiology* **2009**, *20*, 367–373. [[CrossRef](#)]
4. Mistry, H.D.; Williams, P.J. The importance of antioxidant micronutrients in pregnancy. *Oxid. Med. Cell Longev.* **2011**, *2011*, 841749. [[CrossRef](#)]
5. Grandjean, P.; Landrigan, P.J. Neurobehavioural effects of developmental toxicity. *Lancet Neurol.* **2014**, *13*, 330–338. [[CrossRef](#)]
6. Haynes, E.N.; Sucharew, H.; Kuhnell, P.; Alden, J.; Barnas, M.; Wright, R.O.; Parsons, P.J.; Aldous, K.M.; Praamsma, M.L.; Beidler, C.; et al. Manganese Exposure and Neurocognitive Outcomes in Rural School-Age Children: The Communities Actively Researching Exposure Study (Ohio, USA). *Environ. Health Perspect.* **2015**, *123*, 1066–1071. [[CrossRef](#)]
7. Nadaska, A.G.; Lesny, J.; Michalik, I. Environmental aspect of manganese chemistry. *Hung. J. Sci.* **2010**, *100702*, 1–16.
8. ATSDR. Toxicological Profile for Manganese. U.S. Department of Health and Human Services, Public Health Service, ATSDR, Atlanta, GA. Available online: <https://www.atsdr.cdc.gov/toxprofiles/tp151.pdf> (accessed on 6 June 2019).
9. Bouchard, M.; Laforest, F.; Vandelac, L.; Bellinger, D.; Mergler, D. Hair manganese and hyperactive behaviors: Pilot study of school-age children exposed through tap water. *Environ. Health Perspect.* **2007**, *115*, 122–127. [[CrossRef](#)] [[PubMed](#)]
10. Carvalho, C.F.; Menezes-Filho, J.A.; de Matos, V.P.; Bessa, J.R.; Coelho-Santos, J.; Viana, G.F.; Argollo, N.; Abreu, N. Elevated airborne manganese and low executive function in school-aged children in Brazil. *Neurotoxicology* **2014**, *45*, 301–308. [[CrossRef](#)] [[PubMed](#)]
11. Khan, K.; Wasserman, G.A.; Liu, X.; Ahmed, E.; Parvez, F.; Slavkovich, V.; Levy, D.; Mey, J.; van Geen, A.; Graziano, J.H.; et al. Manganese exposure from drinking water and children’s academic achievement. *Neurotoxicology* **2012**, *33*, 91–97. [[CrossRef](#)] [[PubMed](#)]
12. Torres-Agustín, R.; Rodríguez-Agudelo, Y.; Schilman, A.; Solís-Vivanco, R.; Montes, S.; Riojas-Rodríguez, H.; Cortez-Lugo, M.; Ríos, C. Effect of environmental manganese exposure on verbal learning and memory in Mexican children. *Environ. Res.* **2013**, *121*, 39–44. [[CrossRef](#)] [[PubMed](#)]
13. Gulson, B.; Mizon, K.; Taylor, A.; Korsch, M.; Stauber, J.; Davis, J.M.; Louie, H.; Wu, M.; Swan, H. Changes in manganese and lead in the environment and young children associated with the introduction of methylcyclopentadienyl manganese tricarbonyl in gasoline—Preliminary results. *Environ. Res.* **2006**, *100*, 100–114. [[CrossRef](#)] [[PubMed](#)]
14. Gunier, R.B.; Arora, M.; Jerrett, M.; Bradman, A.; Harley, K.G.; Mora, A.M.; Kogut, K.; Hubbard, A.; Austin, C.; Holland, N.; et al. Manganese in teeth and neurodevelopment in young Mexican-American children. *Environ. Res.* **2015**, *142*, 688–695. [[CrossRef](#)] [[PubMed](#)]
15. Mora, A.M.; Arora, M.; Harley, K.G.; Kogut, K.; Parra, K.; Hernández-Bonilla, D.; Gunier, R.B.; Bradman, A.; Smith, D.R.; Eskenazi, B. Prenatal and postnatal manganese teeth levels and neurodevelopment at 7, 9, and 10.5 years in the CHAMACOS cohort. *Environ. Int.* **2015**, *84*, 39–54. [[CrossRef](#)] [[PubMed](#)]
16. Wasserman, G.A.; Liu, X.; Parvez, F.; Ahsan, H.; Levy, D.; Factor-Litvak, P.; Kline, J.; van Geen, A.; Slavkovich, V.; Lolocono, N.J.; et al. Water manganese exposure and children’s intellectual function in Araihaazar, Bangladesh. *Environ. Health Perspect.* **2006**, *114*, 124–129. [[CrossRef](#)]

17. Freeland-Graves, J.H.; Mousa, T.Y.; Kim, S. International variability in diet and requirements of manganese: Causes and consequences. *J. Trace Elem. Med. Biol.* **2016**, *38*, 24–32. [[CrossRef](#)]
18. Koivistoinen, P. (Ed.) Mineral element composition of Finnish foods: N, K, Ca, Mg, P, S, Fe, Cu, Mn, Zn, Mo, Co, Ni, Cr, F, Se, Si, Rb, Al, B, Br, Hg, As, Cd, Pb, and ash. In *Acta Agriculturae Scandinavica Supplementum 22*; Scandinavian Association of Agricultural Scientists and the Royal Swedish Academy of Agriculture and Forestry: Stockholm, Sweden, 1980.
19. Irizar, A.; Gil, F.; Lertxundi, A.; Martín-Domingo, M.C.; Urbieto, N.; Molinuevo, A.; Ibarluzea, J.; Basterrechea, M.; Aurrekoetxea, J.J.; Jiménez-Zabala, A.; et al. Manganese levels in newborns' hair by maternal sociodemographic, dietary and environmental factors. *Environ. Res.* **2019**, *170*, 92–100. [[CrossRef](#)]
20. Mena, I.; Horiuchi, K.; Lopez, G. Factors enhancing entrance of manganese into brain: Iron deficiency and age. *J. Nucl. Med.* **1974**, *15*, 516.
21. Aschner, M. Manganese: Brain transport and emerging research needs. *Environ. Health Perspect.* **2000**, *108* (Suppl. 3), S429–S432.
22. U.S. EPA. *Report on the 2015 U.S. Environmental Protection Agency (EPA) International Decontamination Research and Development Conference*; Environmental Protection Agency: Washington, DC, USA, 2015.
23. Leonhard, M.J.; Chang, E.T.; Loccisano, A.E.; Garry, M.R. A systematic literature review of epidemiologic studies of developmental manganese exposure and neurodevelopmental outcomes. *Toxicology* **2019**, *420*, 46–65. [[CrossRef](#)]
24. Menezes-Filho, J.A.; Novaes, C.D.O.; Moreira, J.C.; Sarcinelli, P.N.; Mergler, D. Elevated manganese and cognitive performance in school-aged children and their mothers. *Environ. Res.* **2011**, *111*, 156–163. [[CrossRef](#)] [[PubMed](#)]
25. Bouchard, M.F.; Sauvé, S.; Barbeau, B.; Legrand, M.; Brodeur, M.È.; Bouffard, T.; Limoges, E.; Bellinger, D.C.; Mergler, D. Intellectual impairment in school-age children exposed to manganese from drinking water. *Environ. Health Perspect.* **2011**, *119*, 138–143. [[CrossRef](#)] [[PubMed](#)]
26. Wright, R.O.; Amarasiriwardena, C.; Woolf, A.D.; Jim, R.; Bellinger, D.C. Neuropsychological correlates of hair arsenic, manganese, and cadmium levels in school-age children residing near a hazardous waste site. *Neurotoxicology* **2006**, *27*, 210–216. [[CrossRef](#)] [[PubMed](#)]
27. Riojas-Rodríguez, H.; Solís-Vivanco, R.; Schilmann, A.; Montes, S.; Rodríguez, S.; Ríos, C.; Rodríguez-Agudelo, Y. Intellectual function in Mexican children living in a mining area and environmentally exposed to manganese. *Environ. Health Perspect.* **2010**, *118*, 1465–1470. [[CrossRef](#)] [[PubMed](#)]
28. Oulhote, Y.; Mergler, D.; Barbeau, B.; Bellinger, D.C.; Bouffard, T.; Brodeur, M.È.; Saint-Amour, D.; Legrand, M.; Sauvé, S.; Bouchard, M.F. Neurobehavioral function in school-age children exposed to manganese in drinking water. *Environ. Health Perspect.* **2014**, *122*, 1343–1350. [[CrossRef](#)] [[PubMed](#)]
29. Rink, S.M.; Ardoino, G.; Queirolo, E.I.; Cicariello, D.; Mañay, N.; Kordas, K. Associations between hair manganese levels and cognitive, language, and motor development in preschool children from Montevideo, Uruguay. *Arch. Environ. Occup. Health* **2014**, *69*, 46–54. [[CrossRef](#)]
30. Coetzee, D.J.; McGovern, P.M.; Rao, R.; Harnack, L.J.; Georgieff, M.K.; Stepanov, I. Measuring the impact of manganese exposure on children's neurodevelopment: Advances and research gaps in biomarker-based approaches. *Environ. Health* **2016**, *15*, 91. [[CrossRef](#)]
31. Lin, C.C.; Chen, Y.C.; Su, F.C.; Lin, C.M.; Liao, H.F.; Hwang, Y.H.; Hsieh, W.S.; Jeng, S.F.; Su, Y.N.; Chen, P.C. In utero exposure to environmental lead and manganese and neurodevelopment at 2 years of age. *Environ. Res.* **2013**, *123*, 52–57. [[CrossRef](#)]
32. Takser, L.; Mergler, D.; Hellier, G.; Sahuquillo, J.; Huel, G. Manganese, monoamine metabolite levels at birth, and child psychomotor development. *Neurotoxicology* **2003**, *24*, 667–674. [[CrossRef](#)]
33. Rahman, S.M.; Kippler, M.; Tofail, F.; Bölte, S.; Derakhshani Hamadani, J.; Vahter, M. Manganese in Drinking Water and Cognitive Abilities and Behavior at 10 Years of Age: A Prospective Cohort Study. *Environ. Health Perspect.* **2017**, *125*, 057003. [[CrossRef](#)]
34. Lucchini, R.G.; Guazzetti, S.; Zoni, S.; Donna, F.; Peter, S.; Zacco, A.; Salmistraro, M.; Bontempi, E.; Zimmerman, N.J.; Smith, D.R. Tremor, olfactory and motor changes in Italian adolescents exposed to historical ferro-manganese emission. *Neurotoxicology* **2012**, *33*, 687–696. [[CrossRef](#)] [[PubMed](#)]
35. Guxens, M.; Ballester, F.; Espada, M.; Fernández, M.F.; Grimalt, J.O.; Ibarluzea, J.; Olea, N.; Rebagliato, M.; Tardón, A.; Torrent, M.; et al. Cohort Profile: The INMA—Infancia y Medio Ambiente—(Environment and Childhood) Project. *Int. J. Epidemiol.* **2012**, *41*, 930–940. [[CrossRef](#)] [[PubMed](#)]

36. Gil, F.; Hernández, A.F. Toxicological importance of human biomonitoring of metallic and metalloid elements in different biological samples. *Food Chem. Toxicol.* **2015**, *80*, 287–297. [[CrossRef](#)] [[PubMed](#)]
37. Molina-Villalba, I.; Lacasaña, M.; Rodriguez-Barranco, M.; Hernandez, A.; Gonzalez-Alzaga, B.; Aguilar-Garduño, C.; Gil, F. Biomonitoring of arsenic, cadmium, lead, manganese and mercury in urine and hair of children living near mining and industrial areas. *Chemosphere* **2015**, *24*, 83–91. [[CrossRef](#)] [[PubMed](#)]
38. Olmedo, P.; Pla, A.; Hernández, A.F.; López-Guarnido, O.; Rodrigo, L.; Gil, F. Validation of a method to quantify chromium, cadmium, manganese, nickel and lead in human whole blood, urine, saliva and hair samples by electrothermal atomic absorption spectrometry. *Anal. Chim. Acta* **2010**, *659*, 60–67. [[CrossRef](#)]
39. McCarthy, D. *MSCA. Escalas McCarthy de Aptitudesy Psicomotricidad para Niños*; TEA Ediciones: Madrid, Spain, 2009.
40. Andiarena, A.; Balluerka, N.; Gorostiaga, A.; Ibarluzea, J. Neuropsychological Assessment at Preschool Age: Adaptation and Validation of the McCarthy Scales of Children’s Abilities to 4 Year-old Basque-speaking Children. *Span. J. Psychol.* **2017**, *20*, E49. [[CrossRef](#)]
41. Julvez, J.; Forns, M.; Ribas-Fitó, N.; Torrent, M.; Sunyer, J. Attention behavior and hyperactivity and concurrent neurocognitive and social competence functioning in 4-year-olds from two population-based birth cohorts. *Eur. Psychiatry* **2011**, *26*, 381–389. [[CrossRef](#)]
42. Arija, V.; Hernández-Martínez, C.; Tous, M.; Canals, J.; Guxens, M.; Fernández-Barrés, S.; Ibarluzea, J.; Babarro, I.; Soler-Blasco, R.; Llop, S.; et al. Association of Iron Status and Intake During Pregnancy with Neuropsychological Outcomes in Children Aged 7 Years: The Prospective Birth Cohort Infancia y Medio Ambiente (INMA) Study. *Nutrients* **2019**, *11*, 2999. [[CrossRef](#)]
43. Lertxundi, A.; Martínez, M.D.; Ayerdi, M.; Alvarez, J.; Ibarluzea, J.M. Air quality assessment in urban areas of Gipuzkoa (Spain). *Gac. Sanit.* **2010**, *24*, 187–192. [[CrossRef](#)]
44. Velasco, D.; Sánchez de Miguel, M.; Egurza, M.; Arranz, E.; Aranbarri, A.; Fano, E.; Ibarluzea, J. Family context assessment in a public health study. *Gac. Sanit.* **2014**, *28*, 356–362. [[CrossRef](#)]
45. Barreto, F.B.; Sánchez de Miguel, M.; Ibarluzea, J.; Andiarena, A.; Arranz, E. Family context and cognitive development in early childhood: A longitudinal study. *Intelligence* **2017**, *65*, 11–22. [[CrossRef](#)]
46. Erikson, K.M.; Thompson, K.; Aschner, J.; Aschner, M. Manganese neurotoxicity: A focus on the neonate. *Pharmacol. Ther.* **2007**, *113*, 369–377. [[CrossRef](#)] [[PubMed](#)]
47. Soler-Blasco, R.; Murcia, M.; Lozano, M.; González-Safont, L.; Amorós, R.; Ibarluzea, J.; Broberg, K.; Irizar, A.; Lopez-Espinosa, M.J.; Lertxundi, N.; et al. Prenatal manganese exposure and effects on neuropsychological development in early childhood in the INMA cohort. *Int. J. Hygiene Environ. Health* **2020**, *224*, 113443. [[CrossRef](#)]
48. Eastman, R.R.; Jursa, T.P.; Benedetti, C.; Lucchini, R.G.; Smith, D.R. Hair as a biomarker of environmental manganese exposure. *Environ. Sci. Technol.* **2013**, *47*, 1629–1637. [[CrossRef](#)] [[PubMed](#)]
49. Mora, A.M.; Córdoba, L.; Cano, J.C.; Hernandez-Bonilla, D.; Pardo, L.; Schnaas, L.; Smith, D.R.; Menezes-Filho, J.A.; Mergler, D.; Lindh, C.H.; et al. Prenatal Manganese Exposure, Excess Manganese, and Neurodevelopment at 1 Year of Age in the Infants’ Environmental Health (ISA) Study. *Environ. Health Perspect.* **2018**, *126*, 057007. [[CrossRef](#)] [[PubMed](#)]
50. Llop, S.; Lopez-Espinosa, M.J.; Rebagliato, M.; Ballester, F. Gender differences in the neurotoxicity of metals in children. *Toxicology* **2013**, *311*, 3–12. [[CrossRef](#)]
51. Arranz Freijo, E.B.; Olabarrieta Artetxe, F.; Manzano Fernández, A.; Martín Ayala, J.L.; Galende Pérez, N. Etxadi-Gangoiti scale: A proposal to evaluate the family contexts of two-year-old children. *Early Child Dev. Care* **2014**, *184*, 933–948. [[CrossRef](#)]
52. Boucher, O.; Julvez, J.; Guxens, M.; Arranz, E.; Ibarluzea, J.; Sánchez de Miguel, M.; Fernández-Somoano, A.; Tardon, A.; Rebagliato, M.; Garcia-Esteban, R.; et al. Association between breastfeeding duration and cognitive development, autistic traits and ADHD symptoms: A multicenter study in Spain. *Pediatr. Res.* **2017**, *81*, 434–442. [[CrossRef](#)]
53. Do Nascimento, S.N.; Barth, A.; Göethel, G.; Baierle, M.; Charao, M.F.; Brucker, N.; Moro, A.M.; Bubols, G.B.; Sobreira, J.S.; Sauer, E.; et al. Cognitive deficits and ALA-D-inhibition in children exposed to multiple metals. *Environ. Res.* **2015**, *136*, 387–395. [[CrossRef](#)]

