

1 **Applying Fault Tree Analysis to biological invasions**
2 **identifies optimal targets for effective biosecurity**

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17 **Abstract**

- 18 1. The management of invasive species requires analytical tools that can synthesise the
19 increasing and complex information generated through risk assessment protocols. To
20 that end, Fault Tree Analysis (FTA) provides a means to conceptually map all of the
21 events leading to a particular undesired scenario with associated probabilities and
22 uncertainty.
- 23 2. We used a peer-reviewed dataset (the GB Non-Native Species Risk Assessments) to
24 build and quantify a FT of all the events leading to the transport, introduction,
25 establishment and spread of harmful aquatic invasive species in Great Britain. We also
26 simulated management scenarios.
- 27 3. Individual barriers to invasion, either natural or human, were largely unsuccessful in
28 hindering invasion (42-91% probability of failure in a 5-year period); yet the high
29 interdependence of events in the tree resulted in an overall probability of harmful
30 invasion of about 3%. This figure is much greater than that estimated by the tens rule,
31 which posits that 10% of non-native species manage to colonise a new area, and only
32 10% of those become invasive, resulting in a 1% overall probability of harmful
33 invasion.
- 34 4. We used the FTA to explore different management intervention scenarios and found
35 that pre-border management reduced the overall risk of invasion by 86%, followed in
36 importance by early action after introduction (85%), and detection at the border
37 (81%). In contrast, post-establishment management techniques, such as eradication
38 and containment, had a limited impact on the probability of widespread invasion (18-
39 24%).
- 40 5. *Synthesis and applications.* While prevention has been long recognized as the most
41 cost-effective action against biological invasions, here we were able to quantify the
42 reduction in invasion risk under a range of management scenarios. Optimising all

43 management elements included in the FT reduced the overall probability of invasion
44 by three orders of magnitude.

45 6. We conclude that FTA provides a baseline to capitalise on a growing source of peer-
46 reviewed risk assessments, which allows systematic assessment of the effectiveness of
47 future actions to prevent and manage invasive species at the national and
48 international levels. The analytical framework can be extended to other biological
49 threats (e.g. pests, pathogens, diseases) and scenarios (e.g. climate change, war), so
50 that breach and leverage points in biosecurity can be identified.

51 **Keywords:** aquatic invasive species, biosecurity, cost-effectiveness, failure, fault-tree, invasion
52 scenario, risk assessment, prevention.

53 **Resumen**

54 1. La gestión de las invasiones biológicas requiere de herramientas analíticas que permitan
55 sintetizar la creciente y compleja información generada a través de protocolos de análisis de
56 riesgos. Para ello, los árboles de fallas (*Fault Tree Analysis, FTA*) permiten mapear
57 conceptualmente todos los eventos que tienen que suceder para alcanzar un determinado
58 escenario de catástrofe, incluyendo sus probabilidades e incertidumbre.

59 2. Utilizamos una base de datos revisada por expertos (*GB No-Native Species Risk*
60 *Assessments*) para construir y cuantificar un árbol de fallas con todos los eventos que
61 conducen al transporte, introducción, establecimiento y dispersión de especies invasoras
62 acuáticas con impacto en Gran Bretaña. También simulamos escenarios de manejo.

63 3. Las barreras individuales frente a la invasión, tanto las naturales como las artificiales, se
64 revelaron como muy ineficaces a la hora de frenar el riesgo de invasión (42-91% probabilidad
65 de fallo en un rango de 5 años); y sin embargo la elevada interdependencia de los eventos en
66 el árbol resultó en una probabilidad total de invasión de tan solo el 3%. Este dato es mucho

67 mayor que el predicho por la regla de los 10, que anticipa que solo el 10% de las especies no
68 nativas introducidas llegan a colonizar un nuevo área, y de ellas tan solo el 10% se vuelven
69 invasoras, por lo que la probabilidad de una invasión biológica dañina quedaría en un 1%.

70 4. Utilizamos el árbol de fallas para explorar escenarios alternativos de gestión y
71 confirmamos que el manejo en origen reduce el riesgo total de invasión en un 86%, seguido en
72 importancia por la erradicación inmediata tras la introducción (85%) y la detección en la
73 frontera (81%). En comparación, la gestión reactiva, como la erradicación y contención a largo
74 plazo, tienen un impacto muy bajo en la probabilidad de invasión (18-24%).

75 5. Síntesis y aplicaciones. Siempre hemos considerado la prevención como la actuación
76 más coste-eficiente frente a las invasiones biológicas. En este estudio hemos sido capaces de
77 cuantificar la reducción en la probabilidad de invasión dañina que supone la prevención
78 proactiva bajo un amplio rango de escenarios de manejo. Optimizar todas las barreras de
79 manejo en el árbol de fallas redujo la probabilidad de invasión dañina en tres órdenes de
80 magnitud.

81 6. Concluimos que los árboles de fallas proporcionan una herramienta óptima para
82 aprovechar la información de experto recogida de forma sistemática en los análisis de riesgo,
83 permitiendo estudiar la efectividad de actuaciones de prevención y gestión a escala nacional e
84 internacional. El marco de trabajo se podría extender a otros riesgos biológicos (por ej. pestes,
85 patógenos, enfermedades) y escenarios (por ej. cambio climático, guerra), de modo que
86 podamos identificar los puntos débiles y fuertes de cada estrategia en bioseguridad.

87 Palabras clave: especies invasoras acuáticas, bioseguridad, coste-efectividad, fallo, árbol
88 de fallas, escenario de invasión, prevención,

89 **Introduction**

90 The worldwide spread of invasive alien species is accelerating (Seebens et al., 2017),
91 negatively affecting human, animal, and plant health (Vilà et al., 2021), damaging
92 infrastructure (Pimentel et al., 2005), and threatening native biodiversity (Brondizio et al.,
93 2019; Pyšek et al., 2020), costing the global economy US\$26.8 billion each year (Diagne et al.,
94 2021). As a result of these growing threats, strategies to deliver biological security have
95 become a global priority in order to protect economies, society and the environment (Kemp et
96 al., 2021). At a national scale, the UK's biological security strategy (2018) incorporates the
97 need to reduce the risk of spreading invasive species, pests, pathogens and diseases. In
98 particular, it highlights the need to learn from previous incidents and build scientific
99 capabilities to respond to biological risks.

100 One method of learning from previous incidents is through Fault Tree Analysis (FTA). This is a
101 systematic and deductive approach that allows for estimation of the likelihood of an undesired
102 "Top Event" happening based upon the probability of failure of natural and management
103 barriers, thereby identifying the events and pathways that are most likely to cause breaches in
104 biosecurity (Ruijters & Stoelinga, 2015). Invasive alien species go through five major stages –
105 transport, introduction, establishment, spread and impact—that result in severe impacts on
106 biodiversity such as the displacement and ultimately extinction of native species, as well as
107 impacts on socio-economic interests such as agriculture yields, tourism or human health; this is
108 what we call the "Top Event". As represented in Figure 1, a series of barriers hinder whether a
109 species becomes invasive. Some barriers are natural and therefore intrinsic to the system such
110 as biogeographic barriers (mountain ranges, oceans), survival during transportation from
111 native to invaded ranges, founder effects, environmental filtering or physical barriers to
112 dispersal (Blackburn et al., 2011). Other barriers are related to the management of either
113 species or the environment they invade, such as pre-border pathway management,
114 disinfection in transit, secure keeping, early eradication or long-term containment (Robertson
115 et al., 2020).

116 The application of FTA to complex ecological systems such as biological invasions has so far
117 been limited to a few examples. Two decades ago, Hayes (2002a) built a comprehensive, yet
118 descriptive, fault-tree for ballast water introductions; a similar exercise was later applied to
119 recreational boating (Acosta & Forrest, 2009). These two conceptual trees demonstrated the
120 value of FTA to guide decision-making by detailing all of the elements leading to the
121 introduction of unwanted organisms through very specific pathways. Subsequent exercises
122 populated the different elements in the tree with probabilities of failure obtained through
123 expert elicitation. For instance Hayes (2002b) identified 286 combinations of vessel
124 components and infection modes that were then scored through a series of nine workshops
125 with different water users and authorities. This extraordinary effort allowed identifying water
126 retention and internal fouling as the most hazardous for the introduction of marine alien
127 invaders. Similarly, Acosta et al. (2010) used expert elicitation and a fuzzy logic system to
128 quantify the risk associated with 72 different scenarios of vessel infection; they identified hull
129 fouling, deck, internal space, anchor and fishing gear as high-risk invasion mechanisms. These
130 initial works focused on the first steps of the invasion process (i.e. transport and introduction
131 into the invaded area), but, to date, no FTA has covered the full invasion curve (Fig. 1),
132 including establishment and secondary spread. In addition, expert elicitation is time and
133 resources intensive and provides a snapshot of the scientific evidence that is difficult to update
134 unless the exercise is repeated periodically.

135 Recently, a variety of risk assessment protocols have been developed to systematically
136 evaluate and synthesise the scientific evidence available from other invaded regions about the
137 likelihood of transport, introduction, establishment, spread and impact of invaders (González-
138 Moreno et al., 2019). The results of risk assessments are commonly used to rank established or
139 prospective invasive species in a given area so that management resources can be focused on
140 the biggest risks (e.g. Gallardo et al., 2016; Peyton et al., 2019; Roy, Bacher, et al., 2018).
141 However, simply ranking invasive species is of limited use for management because it fails to

142 identify the strategic management actions that are more likely to reduce the probability of the
143 worst invasion scenario (Booy et al., 2020). Since their initial development in the early 2010s,
144 hundreds of invasive alien species have been systematically evaluated using risk assessment
145 protocols (e.g. 116 in Great Britain, 135 by the European Commission, 2614 by the COST Action
146 AlienChallenge (González-Moreno et al., 2019; Roy, Bacher, et al., 2018)), and will continue to
147 be evaluated each year, offering a unique opportunity to use hazard identification techniques,
148 such as FTA, to analyse the natural barriers and failures in management that have contributed
149 to their widespread invasion.

150 Here we build an FTA to investigate the chain of events that leads to biological invasions in
151 Great Britain. We assembled a descriptive FTA and then use data from the GB-Non Native
152 Species Risk Assessment (GB-NNRA, www.nonnativespecies.org) to calculate the probability
153 and uncertainty associated with the basic events leading to an invasion scenario. We further
154 use FTA to simulate management scenarios, under the basic hypothesis that proactive
155 intervention before or at the border is more effective at reducing the overall risk of invasion
156 than reactive post-border interventions. This can be tested by simulating a reduction in the
157 probability of failure of the different management barriers. The novelty of our FTA relies in
158 that we build a fully quantitative tree that covers the initial four stages of the invasion process
159 leading to widespread negative impacts (transport, introduction, establishment and spread),
160 and simulate management scenarios. The result is a powerful analytical tool that can
161 synthesise the increasing and complex information generated through risk assessment
162 protocols and can be easily updated as new species and new information are evaluated. The
163 FTA framework can be extended and stress-tested for other biological threats (e.g. pests,
164 pathogens, diseases), scenarios (e.g. climate change) and contexts, so that the riskiest points in
165 the Fault Tree can be identified and the cost-effectiveness of improved biosecurity at different
166 stages quantified accordingly. More generally, we discuss the application of FTA to ecological
167 systems and suggest ways of improvement.

168 **Material and Methods**

169 We developed a Fault Tree to analyse the risk of biological invasions in Great Britain
170 following six steps.

171 ***Step 1. Constructing the Fault Tree***

172 FTA is a deductive technique where we start from the Top Event (i.e. invasion scenario)
173 and work top-down to identify the specific chain of events that caused failure of the system. A
174 Fault Tree is, therefore, a graphical model of all the parallel and sequential combinations of
175 events that lead to the Top Event (Hayes, 2002a). A FTA consists of two types of nodes: events
176 and gates. *Intermediate events* are caused by the combination of several *basic events*, which
177 do not need or can't be subdivided further, because an appropriate level of resolution has
178 been reached. We connected basic events into intermediate events using logical gates. Gates
179 determine how failures propagate through the system: "AND" gates are used when all of the
180 basic events have to occur for the intermediate event to take place. In the case of "OR" gates,
181 the intermediate event occurs if at least one (or all) of the basic events occur.

182 The GB-NNRA was established in 2006 and up to 2020 has been used to systematically
183 evaluate the risks associated with 116 invasive species. The scheme provides a structured
184 framework with four modules for evaluating the potential for non-native organisms to Enter,
185 Establish, Spread and cause significant Impacts in all or part of Great Britain. Assessments are
186 carried out by independent specialists, peer-reviewed by both an external expert and a central
187 panel of risk assessment professionals, and finally open to public feedback at the GB-Non-
188 native Species Secretariat webpage (GB-NNSS, <http://www.nonnativespecies.org>). To build our
189 FTA, we identified the basic events that can lead to the successful establishment and spread of
190 invasive species using the 34 questions in the Entry, Establishment and Spread modules (Table
191 S1). Intermediate events are not directly evaluated in the protocol, but combine multiple basic
192 events that can be linked together. An example of this is the intermediate event "biotic filter"

193 (Event 10 in Table 1, Fig. 2) that combines biological aspects that may prevent establishment
194 related to the invasive species (e.g. genetic bottlenecks, lack of adaptability to local conditions)
195 and the invaded community (e.g. presence of competitors or natural enemies). These
196 biological barriers are combined with an AND gate because all of them must fail for the species
197 to continue with the invasion process. In contrast, the intermediate event “management fails”
198 (Event 6 in Table 1, Fig 2) combines containment and eradication barriers connected in this
199 case with an OR gate, because failure in just one of them would allow the species to continue
200 spreading. We ignored the Impact module of the GB-NNRA because it describes the
201 consequences rather than the causes of the “Top Event” i.e. the harmful and widespread
202 colonisation of the study area by a non-native species.

203 ***Step 2. Obtaining probabilities from expert judgement***

204 From the 116 species evaluated using GB-NNRA, we downloaded risk assessments for 28
205 aquatic species (Table S2) from the GB-NNSS webpage (<http://www.nonnativespecies.org>, last
206 accessed 1st September 2019). We focused on the aquatic environment because, in
207 comparison with terrestrial habitats, it is i) more *invasible* because of multiple vectors and
208 pathways of invasions that are difficult to control, ii) particularly susceptible to the negative
209 impacts of invasion, and iii) challenging to manage because of the low detectability of species,
210 and high connectivity inherent to the aquatic environment (Moorhouse & Macdonald, 2015).
211 The list included 19 freshwater and 9 marine organisms introduced through three major
212 gateways: stowaway (N= 10), aquaculture (N= 10) and ornamental (N= 8), which correspond to
213 the most important pathways of aquatic introduction in Europe (Nunes et al., 2015). These 28
214 species have been identified by the GB-NNSS as representing the greatest immediate risks to
215 biosecurity in Great Britain and include high-profile invaders such as the water hyacinth
216 (*Eichhornia crassipes*), the zebra and quagga mussels (*Dreissena polymorpha* and *D. r.*
217 *bugensis*), the signal crayfish (*Pacifastacus leniusculus*) and the topmouth gudgeon
218 (*Pseudorasbora parva*).

219 In traditional FTA, failure probabilities of system components are used to investigate the
220 probability of the Top Event. This is because original Fault Trees were used in the chemical and
221 safety industries where the probability of operational faults in valves and pumps can be
222 quantitatively measured. In ecological systems, however, this kind of information is rarely
223 available. In the case of the GB-NNRA, the assessor is required to answer questions choosing
224 one of five levels that represent the probability of the event occurring over a 5-year period
225 (very low= 0-10%, low= 11-33%, medium= 34-66%, high= 67-90%, and very high= 91-100%);
226 justifying these with a written, referenced comment (Mumford et al., 2010). The scale is
227 narrower at the extremes to allow more positive discrimination of both very rare and highly
228 likely events. Histograms showing the frequency of each response are given in Figure S1. Here
229 we used the mean value to represent each category (5, 22, 50, 78 and 95%). We replicated
230 analyses using alternative transformations (e.g. 10-33-66-90-100%; 0-11-34-67-91%) but given
231 that general patterns were similar, they are not shown for simplicity.

232 We calculated the mean probability of failure (Q) across the 28 species for each of the 34
233 questions in the GB-NNRA and used these values to populate the Fault Tree. Results using the
234 median probability did not differ significantly, and are therefore not reported here.

235 ***Step 3. Quantification of the Fault Tree: probability of the Top Event***

236 To calculate the probability of the Top Event (i.e. the harmful and widespread colonisation
237 of the study area by a non-native species that results in severe impacts on biodiversity and/or
238 socio-economic interests), numerical probabilities (Q) assigned to the basic events were
239 combined using boolean statistics depending on the gates that connect them: the probability
240 of events linked with an OR gate is summed, whereas the probability of events linked with
241 AND gates is multiplied. We calculated the probability of the Top Event for individual species
242 using the R package FaultTree (Silkworth, 2017a), and also grouped them by pathway and
243 region of origin.

244 ***Step 4. Quantification of the Fault Tree: importance of basic events***

245 The importance of basic events provides information about their impact on the system.
246 This analysis can be used to prioritise those management actions that could reduce the overall
247 risk. There are multiple measures of importance, here we employed the following indicators,
248 already implemented in the R package FaultTree.SCRAM (Silkworth, 2017b):

- 249 - *Marginal Importance Factor (MIF)* gives the increase in risk due to the failure of the
250 event by measuring the difference between failed-event ($Q_{\text{event}=1}$) and non-failed
251 event ($Q_{\text{event}=0}$).
- 252 - *Risk Reduction Worth (RRW)* indicates the maximum decrease in risk of the system if
253 the event never fails; it measures the difference between the current level of risk
254 (Q_{event}) and the risk if the component never fails ($Q_{\text{event}=0}$). RRW is useful for
255 prioritising improvements that can most reduce the risk in the future.
- 256 - *Minimal Cut Sets (MCS)* is the minimum collection of basic events that would lead to
257 the Top Event. MCS are used to understand the structural vulnerability of the system
258 and identify weak points.

259 Importance measures take into account the position of the event in the FT, which means
260 that those events with high MIF or RRW do not necessarily show the highest probability of
261 failure (Q).

262 ***Step 5. Uncertainty of the Fault Tree***

263 Confidence in question responses, registered in the GB-NNRA protocol, was transformed
264 into a numerical measure of uncertainty (U) using the following scale: “Certain”= 0, “Low
265 uncertainty”= 0.2, “Moderate uncertainty”= 0.4 and “High uncertainty”= 0.6. The maximum
266 confidence level (Certain) is used for events that refer to intrinsic characteristics of the species
267 that are well known, such as the ability to establish and reproduce. We used the package

268 “propagate” in R (Spiess et al., 2018) to propagate the error of basic events towards the Top
269 Event.

270 ***Step 6. Simulate management scenarios***

271 We simulated a range of management scenarios by modifying the probability of failure of
272 all management events, individually and collectively, from 5 to 100%. This allowed exploration
273 of the effectiveness of investing in one single strategy, *versus* a more holistic approach to
274 prevent invasions. Our options therefore range from an “ideal scenario”, where the probability
275 of failure of all management interventions is set to 5%, to the “worst case scenario” where all
276 probabilities of failure are set to 100%.

277 **Results**

278 Our Fault Tree of aquatic invasions in Great Britain was composed of 26 basic and 12
279 intermediate events that correspond to the sequential and parallel barriers, both natural and
280 related to management, which hinders biological invasions. Steps in the FT are fully described
281 in Supplementary Information 2, and synthesised in Figure 2.

282 ***Quantification of the Fault Tree: probability of the Top Event***

283 The fault probability of the invasion scenario, averaged across 28 aquatic species, was
284 calculated as $3.14 \pm 0.05\%$ for a 5-year period. Because the probability of the Top Event strongly
285 depends on the number of events included in the tree, and the idiosyncrasies of the 28 species
286 that have been assessed by the GB-NNRA, we must use this value with caution to compare
287 across species and scenarios rather than as an absolute probability of invasion.

288 Barriers showing very high probability of failure (darker shading in Fig. 2) include
289 propagule pressure (Event 38, $Q=98 \pm 13\%$), transference from the pathway to the natural
290 environment (Event 19, $Q=95 \pm 6\%$), biotic filters (Events 26 and 30, $Q= 99 \pm 1\%$ and $92 \pm 6\%$
291 respectively), and management post-establishment (Event 6, $Q=97 \pm 5\%$). The probability of

292 failure of the propagule pressure barrier is particularly high because basic events are
293 connected with an OR gate and thus probabilities of failure are combined. This means that the
294 propagule pressure barrier is expected to fail if any of the initial premises (concentration
295 transported, frequency transported or volume transported) is high enough. This could occur
296 through one single transportation event, albeit with a very high concentration of the invader,
297 or through repeated events with a low concentration of the invader.

298 Species differed in their probability of reaching the Top Event (Fig. 3), and across pathways
299 of introduction (ANOVA, $F_{2,25}=3.77$, $P=0.04$) (Fig. 4). Organisms that were introduced as
300 stowaways were seven times more likely to reach the Top Event than organisms introduced by
301 aquaculture, and four times more than organisms introduced through the ornamental trade.
302 No significant differences in the probability of the Top Event were found among regions of
303 origin or types of habitat. The probability of the Top Event was distinctively high for marine
304 organisms originating in Australia and Asia, but differences were not statistically significant
305 (Fig. 4).

306 ***Quantification of the Fault Tree: importance of Basic Events***

307 Minimal Cut Sets were found with 19 degrees of freedom; this means that 19 (68%) of the
308 28 Basic Events included in the tree should happen for the system to fail completely (Table 1).
309 Early detection at the border (Event 33), adaptability of the species to local conditions (Event
310 13), and the degree to which the pathway (as commodity, contaminant or stowaway; Event
311 34) is associated with source populations of potential invaders, were highlighted as the most
312 important events in terms of Marginal Importance Factor, which calculates the increase in risk
313 due to the failure of the event. Other events in order of importance were: suitable climate and
314 habitat conditions for establishment (Events 23 and 24), and the absence of genetic
315 bottlenecks preventing establishment (Event 14). Events achieving highest Risk Reduction
316 Worth values (RRW, the maximum decrease in risk of the system if the event never fails)

317 include pre-border management (Event 42), capacity of the species to spread (Event 16) and
318 capacity to survive in transit (Event 31).

319 ***Simulation of management scenarios***

320 The average probability of failure across the five management barriers included in the FT is
321 $79\pm 8\%$. This means that management is typically considered highly probable to fail by
322 assessors. Here, we simulated values of management failure between 5 and 100% to find the
323 ideal allocation of resources to avert widespread harmful invasions (Fig. 5). In accordance with
324 our expectations, proactive management, such as reducing the failure of pre-border
325 management (Event 42), early detection (Event 33) or rapid response (Event 17), were able to
326 reduce 17-fold the probability of the Top Event in comparison with the current scenario (Fig.
327 5A). In contrast, reactive measures such as eradication (Event 8) and long-term containment
328 (Event 7) had little to no impact on the overall probability of invasion (Fig. 5A). Despite these
329 figures, it is important to note that control and eradication can still be important to mitigate
330 the harmful effects of the invaders at local or regional scales.

331 Under an ideal management scenario, strong support of management interventions across
332 all stages of the invasion process (probability of failure $Q= 5\%$ for all management events) is
333 able to reduce the probability of the top event considerably with respect to the current
334 scenario (from $Q=3.14\%$ to 0.000087% , Fig. 5B). Simply decreasing the probability of failure of
335 management actions currently calculated at 79% (High) to a modest target of 50% (Medium),
336 would achieve a 5-fold decrease in the chances of the invasion scenario (from $Q=3.14\%$ to
337 $Q=0.67\%$).

338 In contrast, under the worst case scenario of no management (probability of failure
339 $Q=100\%$ for all management events), the probability of harmful widespread invasion doubles
340 in comparison with the current situation (from $Q=3.14\%$ to $Q= 7.12\%$), further evidencing the
341 benefits of the ideal management scenario.

342 **Discussion**

343 ***Applying FTA to assess the risk of biological invasions***

344 In this study, we built the first quantitative Fault Tree Analysis that incorporates the four
345 major stages of the invasion process that lead to widespread harmful invasions: transport,
346 introduction, establishment and spread. We were able to do so by capitalising on the vast
347 amount of information generated by the GB-NNSS after 15 years of systematic risk assessment
348 of invasive species. Across the 28 species we investigated, the probability of failure of natural
349 and management barriers to invasion was high, ranging from 42 to 91% in a 5-year period. As
350 such, most barriers were moderately to very unlikely to avert invasion, at least individually.
351 Yet, because many of them need to fail simultaneously at each major invasion stage and in
352 subsequent stages, the overall probability of widespread and harmful invasion drops to 3% in a
353 5-year period.

354 The probability of the top event can be compared with the “tens-rule”, which states that
355 about 10% of introduced species take all the consecutive steps of the invasion process, and
356 that 10% of them (that is, 1% of all initial introductions) become invasive and cause “significant
357 detrimental impacts” (sensu Williamson & Brown, 1986); a concept that is equivalent to our
358 Top Event. The tens rule gained popularity in the 1990s even if it had little empirical or
359 theoretical basis, but later evidence suggested that the actual rate of successful introduction
360 may be 4-times higher, with large variations across taxonomic groups (Jeschke & Heger, 2018).
361 This is more in line with the rough 3% that we obtained through a quantitative FTA. This
362 apparently low figure should not be underestimated since the propagule and colonisation
363 pressure of aquatic habitats is enormous: one study estimated that 7–10,000 aquatic species
364 are being transported with ballast water globally at any one time (Endresen et al., 2004);
365 aquaculture is responsible of >5,600 introductions globally (FAO 2019); focusing in Great

366 Britain, there are approximately 560,000 registered boats that can transport invasive species
367 within and towards the country (Ashton et al., 2006).

368 Top event probabilities varied considerably among species, and were highest for those
369 transported as stowaways, one of the most important gateways of aquatic invasion into Great
370 Britain (Gallardo et al., 2016; Gallardo & Aldridge, 2015). The Japanese skeleton shrimp
371 (*Caprella mutica*) and New Zealand Mudsnail (*Potamopyrgus antipodarum*) stand out as the
372 invasive species with the highest probability of widespread invasion in Great Britain. Indeed, *C.*
373 *mutica* (probability of Top Event, Q=39%) has been present in Great Britain since 2000 and is
374 distributed from the English Channel to the Celtic Sea coast (Ashton et al., 2006, 2007; Cook et
375 al., 2007). Likewise, *P. antipodarum* (Q= 24%) is an extremely tolerant species that is common
376 and widespread across freshwater and brackish habitats in Great Britain with a very small body
377 size that is difficult to detect during biosecurity screening (Alonso & Castro-Diez, 2008;
378 Gallardo et al., 2020). Only one of the species investigated, the marbled crayfish (*Procambarus*
379 *fallax f. virginialis*), is not yet established in Great Britain. The probability of the Top Event
380 (Q=0.004%) is one of the lowest, which makes sense because its sale and keeping is regulated
381 in Great Britain, it cannot be imported from non-EU countries, it is adapted to warm
382 subtropical climates, subject to a high competition and predation by resident species, and not
383 known to be able to spread naturally, only through human intervention (Holdich, 2011).

384 In this study, FTA was applied to an ecological system, and this application differs
385 substantially from their typical use in industrial settings. First, basic events are not necessarily
386 accidents or failures but rather aspects related to the species' biology or ecology (e.g.
387 adaptability, traits, survival capacity), or to the environment (e.g. climate and habitat
388 suitability) that cannot be modified but definitely contribute to the process of invasion. The
389 probability of failure of such ecological basic events is difficult to quantify and subject to high
390 uncertainty. Another important difference is the high interdependence between events in

391 ecological systems, which is reflected in the prevalence of AND gates, as opposed to OR gates.
392 This makes it more difficult to find alternative cut sets (the minimum collection of basic events
393 that would lead to the Top Event) but has the practical advantage that minimising the
394 probability of failure of key events can dramatically affect the overall risk of invasion. This can
395 be clearly seen in the management scenarios, where reducing to 10% the probability of failure
396 of just one basic event “detection at the border” resulted in a reduction of >80% in the
397 probability of the Top Event.

398 ***How can we optimise the allocation of resources for management?***

399 Beyond the Top Event, FTA allowed quantification of the maximum reduction in total risk
400 that could be achieved focusing resources on different stages of the invasion process. Based on
401 the evidence collected, we confirm that even small improvements in proactive management
402 can have major repercussion in the overall probability of invasion; whereas the effectiveness
403 of reactive management is rather limited. According to our analysis, eradication programs for
404 aquatic invasive species of established, reproducing populations are very likely (for 68% of
405 species) or likely (25%) to fail (Fig S1A). To illustrate the enormous challenge posed by
406 biological invasions, only nine of the 3,163 non-native species known to have established in
407 Great Britain (0.002%) have been eradicated, and even then only locally (NNSIP scorecard
408 2017). Our study therefore adds to the mounting calls that prevention is far more cost-
409 effective than control for invasive species (Leung et al., 2002). According to Cuthbert et al.
410 (2022), prioritising investment at early invasion stages for prevention and rapid eradication
411 could save trillions in economic costs over the long-term.

412 We can also confirm that investing in one single strategy has limited capacity to reduce the
413 overall risk of invasion (Robertson et al., 2020; Vander Zanden et al., 2010). Considering the
414 current state of invasion of aquatic habitats in Great Britain, with the south-east of England
415 considered one of Europe’s hotspots (Gallardo & Aldridge, 2015; Jackson & Grey, 2013; Zieritz

416 et al., 2014), biosecurity strategies should tackle all stages of the invasion process
417 simultaneously, with a focus on proactive prevention and rapid eradication. National
418 biosecurity campaigns, like Check Clean Dry (www.nonnativespecies.org/checkcleandry) and
419 Be Plant Wise (www.nonnativespecies.org/beplantwise) share such a holistic vision and are
420 fundamental to protect aquatic ecosystems, but their actual efficacy is difficult to quantify
421 (Sutcliffe et al., 2018). An FTA focused on the cascade of human actions and decisions that
422 facilitate the Top Event would best reveal breach and leverage points, and help optimise
423 management of aquatic invasions. For instance, reducing the probability of failure of
424 management barriers affected by the check-clean-dry strategy to a mere 50/50 chance
425 reduces the probability of widespread invasion of *P. antipodarum* from 24% to 0.03%. The GB
426 Risk Management scheme (GB-NNRM) can be further used to qualify basic events in the FT by
427 addressing the feasibility of eradicating invasive species according to their effectiveness,
428 practicality, cost, impact, acceptability, window of opportunity and likelihood of re-invasion
429 (Booy et al., 2017). To that end, it is important that risk assessment and risk management
430 protocols are applied to the same species (which was not always the case), so that we can
431 integrate information about how an invasion may unfold in Great Britain with information
432 about how best to respond to it.

433 Once the FT is built, is it easy to modify and run multiple scenarios, allowing a rapid yet
434 rigorous evaluation of the effectiveness of intervention scenarios that is fundamental to
435 prioritise investment. Another possibility, not explored here, is to simulate climate change
436 scenarios by modifying the probability of events likely to be affected, such as propagule
437 pressure, climate suitability, natural spread (e.g. because of a higher frequency of climatic
438 events such as storms and floods), and the effectiveness of control measures. As such, FTA can
439 be used to set up a baseline and track progress towards reducing the probability of the Top
440 Event.

441 ***Limitations and suggestion for improvement***

442 The FTA has several advantages over other traditional statistical techniques that call for
443 further application to biological invasions. Failure space, albeit initially counterintuitive for an
444 invasion ecologist, was particularly useful to identify all of the management interventions that
445 must fail for the worst case scenario to materialise. This process should be ideally undertaken
446 by a multidisciplinary group of researchers and stakeholders, in order to incorporate different
447 perceptions on what constitutes failure as well as their respective responsibilities. However,
448 this approach is very time and resource intensive (Acosta & Forrest, 2009; as in Hayes, 2002a),
449 which is why we limited the construction of the FT to a predefined set of questions from the
450 GB-NNRA. We considered three major pathways of introduction: aquaculture, ornamental
451 trade and stowaway; but other secondary pathways such as intentional release (as fish bait, or
452 as food source) and natural spread from invaded locations, should be ideally incorporated. For
453 instance, waterfowl can be a vector for a broad spectrum of invertebrates that attach to
454 feathers or are eaten (Frisch et al., 2007). Pre-border intervention could be broken down to
455 the different treatments that may fail such as quarantine, ballast water treatment, hull
456 cleaning or disinfection. Likewise, rapid response could distinguish between rapid detection
457 and immediate eradication. However, it must be noted that to build a FT, it is crucial to obtain
458 quantitative information that enables calculation of the probability of failure of these
459 additional barriers. Otherwise, the application of FTA to biological invasions would continue to
460 be limited by the evidence and data available.

461 One important limitation of FTA is the lack of temporal variation in the probability of
462 failure. In FTA, the basic components either occur or they do not (Hayes, 2002a), yet the
463 window of opportunity for controlling invasive species may vary between months to decades
464 (Booy et al., 2017), which means that a particular management measure applied at the wrong
465 time may increase the likelihood of failure considerably. In addition, the tree is based on data
466 from actual aquatic invasions (equivalent to past accidents in the industry), which means that

467 it is biased towards successful invaders. The difficulty of finding equivalent information for
468 failed introductions is a pervasive problem for invasion studies (Zenni & Nuñez, 2013), which in
469 this case means that the probability of failure of the various filters may be overestimated.

470 **Conclusions**

471 Beyond Great Britain, hundreds if not thousands of species have been evaluated using
472 over 30 different risk assessment protocols available in the literature (González-Moreno et al.,
473 2019; Roy, Rabitsch, et al., 2018). The information generated therein has been primarily used
474 to rank current and future invaders so that funds are focused on the most immediate risks
475 (Gallardo et al., 2016; McGeoch et al., 2012; Peyton et al., 2019; Roy, Bacher, et al., 2018).
476 Beyond rankings, FTA provides a systematic and rigorous tool to synthesise this increasing
477 body of information, thereby allowing practitioners to optimise resources on the events
478 offering the biggest wins to risk reduction. FTA could further allow identifying breaches in
479 biosecurity across different biological threats (e.g. GMO, plant pests, and human diseases) so
480 that common challenges are approached jointly.

481 FTA has had little application to ecological systems to date, but could be useful to tackle
482 other problems such as species extinction, pollution events (e.g. oil spill), food security or the
483 spread of human diseases. Results of this study clearly show the broad applicability of FTA in
484 risk analysis to identify breach points in biological invasions. Given the current understanding
485 and awareness of biosecurity issues globally, we conclude that FTA provides a baseline to
486 systematically assess the effectiveness of future actions to prevent and manage biosecurity
487 threats at the national and international levels.

488 **Authors' contributions**

489 Belinda Gallardo, William Sutherland and David C. Aldridge conceived the ideas and
490 designed methodology; Belinda Gallardo collected and analysed the data with the support of

491 Phillip Martin. All authors contributed critically to the drafts and gave final approval for
492 publication.

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501 **Conflict of Interest**

502 The authors declare no conflicts of interest in relation with this publication.

503 **Data availability statement**

504 Data used to construct the Fault Tree is available through Figshare (DOI:
505 10.6084/m9.figshare.20283339).

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- 664

665 Tables

666 **Table1.** Characteristics of events considered in the Fault Tree of invasion in Great Britain (GB). A description of events can be found in Table 1 and Supplementary
667 Material 2. Type: Top Event (TP), Intermediate Event (IE) or Basic Event (BE). MCS: inclusion in Minimal Cut Sets, the minimum collection of basic events that would lead to
668 the Top Event (yes/no). Q: probability of failure in 0-100% scale. Uncertainty: standard deviation of the probability of failure. The importance of events, measured through
669 the Marginal Importance Factor (MIF, increase in risk due to the failure of the event) and the Risk Reduction Worth (RRW, maximum decrease in risk of the system if the
670 event never fails), is only calculated by the program for Basic Events.

Event Num.	Name	Description	Type	MCS	Q (%)	Uncertainty (%)	MIF (x1000)	RRW
1	INVASION SCENARIO	Harmful and widespread colonisation of the study area by a non-native species, also called Top Event	TP	no	3.14	0.05		
2	ESTABLISHED	Non-native species is established in GB	IE	no	4.09	0.14		
3	Spreads	Secondary spread of the non-native species in GB	IE	no	79	9		
4	By Natural Means	Likelihood the non-native species is not able to spread naturally	BE	no	50	6	0.52	1.85
5	By Human Means	Likelihood the non-native species is not able to spread by human assistance	BE	yes	59	6	0.52	2.18
6	Management Fails	Reactive measures to halt the secondary spread of the non-native species fail	IE	no	97	5		
7	Containment	Likelihood the non-native species cannot be contained within the invaded area	BE	no	85	4	0.35	2.08
8	Eradication	Likelihood the non-native species survives eradication campaigns	BE	yes	79	3	0.35	1.92
9	INTRODUCED	Non-native species is introduced in GB	IE	no	18	1		
10	Biotic Filter	Biotic characteristics of the non-native species or the invaded habitat that prevent establishment	IE	no	28	4		
11	Competitors	Likelihood that competition will prevent establishment of the non-native species	BE	yes	81	5	0.71	- 1.02x10 ¹⁵
12	Natural Enemies	Likelihood that natural enemies will prevent establishment of the non-native species	BE	yes	81	5	0.71	- 1.02x10 ¹⁵
13	Adaptability	Likelihood the non-native species is not able to adapt to local conditions	BE	yes	73	7	0.78	0

14	Genetic bottlenecks	Likelihood genetic diversity in the founder populations prevent establishment	BE	yes	77	6	0.74	-	1.71×10^{15}
15	Traits	Likelihood the non-native species lacks biological characteristics that facilitate its establishment	BE	yes	90	2	0.63	-	5.14×10^{15}
16	Spread	Likelihood the capacity of the non-native species to spread limits establishment	BE	yes	85	4	0.67	-	2.57×10^{15}
17	Management Fails: Rapid Response	Likelihood that the non-native species establishes despite existing management practices	BE	yes	83	3	0.69	-	5.14×10^{15}
18	ENTERS GREAT BRITAIN	Non-native species enters GB	IE	no	35	4			
19	Transferred	Non-native species is transferred from pathway to natural habitat in GB	IE	no	95	6			
20	By Natural Means	Likelihood the non-native species is able to transfer from pathway to suitable habitat or host naturally	BE	yes	78	4	0.37	2.00	
21	By Human Means	Likelihood the intended use of the commodity or other material with which the non-native species is associated aids transfer to a suitable habitat	BE	no	78	4	0.37	2.00	
22	Abiotic Filter	Abiotic characteristics of the habitat preventing establishment in GB	IE	no	53	2			
23	Suitable Climate	Likelihood climate conditions will allow the establishment of the non-native species	BE	yes	77	3	0.74	-	1.71×10^{15}
24	Suitable Habitat	Likelihood the abiotic conditions will prevent the establishment of the non-native species	BE	yes	81	3	0.71	-1.03	$\times 10^{15}$
25	Timing	Likelihood the non-native species arrives during months of the year most appropriate for establishment	BE	yes	86	2	0.66	-5.14	$\times 10^{15}$
26	Biotic Filter	Biotic characteristics of the invaded habitat that prevent introduction into GB	IE	no	99	1			
27	Key species	Likelihood that species or habitats key for the survival, development and multiplication of the non-native species are absent	BE	no	87	0.38	0.32	1.96	
28	Widespread	Likelihood that species or habitats key for the survival, development and multiplication of the non-native	BE	no	91	1	0.32	2.05	

		species are rare							
29	TRANSPORTED TO GB	Non-native species is transported to GB	IE	no	58	7			
30	Biotic Filter	Biotic characteristics of the non-native species that prevents successful transportation into GB	IE	no	92	6			
31	Survives Transit	Likelihood to survive during transport/storage	BE	yes	86	3	0.44	2.91	
32	Reproduces in Transit	Likelihood that the non-native species multiplies or increases in prevalence during transport/storage	BE	no	45	5	0.44	1.52	
33	Management Fails: Detection at border	Likelihood that the non-native species enters GB undetected	BE	yes	66	3	0.87	0	
34	Associated to Pathway	Likelihood that the non-native species is associated to the pathway at origin	IE	no	76	6	0.75	-	5.10x10 ¹⁵
35	Aquaculture	Likelihood that the non-native species is associated to aquaculture	BE	yes	57	6	0.28	1.08	
36	Ornamental	Likelihood that the non-native species is associated to ornamental trade	BE	no	62	2	0.32	1.11	
37	Stowaway	Likelihood that the non-native species is transported accidentally	BE	no	66	6	0.36	1.13	
38	Propagule Pressure	Characteristics that increase propagule pressure	IE	no	92	13			
39	Frequency Transported	Likelihood the frequency of the pathway allows transportation	BE	yes	62	7	0.34	1.60	
40	Volume Transported	Likelihood the volume of movement along the pathway allows transportation	BE	no	42	10	0.34	1.34	
41	High Concentration	Concentration of the non-native species on the pathway at origin	BE	no	62	6	0.34	1.60	
42	Management Fails: pre-border intervention	Likelihood that the non-native species survives existing pre-management practices during passage along the pathway	BE	yes	84	4	0.68	2.57x10 ¹⁵	

Figures

Figure 1. Barriers to invasion along the invasion curve. Invasive alien species go through five different stages of invasion from being absent to widespread and causing negative ecological and socioeconomic impacts in a particular area. Stages are separated by natural (green) and management (red) barriers that must fail for the species to pass on to the next stage. The costs of managing the invasion accumulate over time, while the probabilities of controlling it decline. The invasion curve sets the conceptual failure space for the development of a Fault Tree Analysis. Curve adapted from Harvey and Mazzotti (2014).

Figure 2. Fault-tree of aquatic invasions in Great Britain. Probabilities of failure (Q) refer to a 5-year period. Events are colour-coded according to their probability of failure (see legend at the top-right corner). In brackets, the uncertainty associated with the event. Stages of the invasion process are represented along the left side of the tree. A description of events can be found in Table 1 and Supplementary Material 2.

Figure 3. Probability of harmful widespread invasion (Top Event) of 28 aquatic species calculated using Fault Tree Analysis (FTA). The vertical line represents the probability of the Top Event using data from all species (3.14%). Values represent the probability that all natural and management barriers against the colonisation of the species fail over a 5-year period.

Figure 4: Differences in the probability of harmful widespread invasion (Top Event) of aquatic species in Great Britain across major habitats (ANOVA, $F_{1,26}=1.72$, $P>0.05$), origins (ANOVA, $F_{4,23}=1.88$, $P>0.05$) and pathways of invasion (ANOVA, $F_{2,25}=3.77$, $P=0.04$).

Figure 4. Simulated changes in the likelihood of invasion at increasing probabilities of failure of individual management actions (A), or all management actions simultaneously (B). The probability of the Top Event refers to the probability of a non-native species passing through all of the natural and management barriers to invasion included in the fault-tree over a 5-year period. A probability of failure of $Q=5\%$ for all management interventions simultaneously represents an “ideal scenario”, whereas the “worst scenario” represents a situation where no management of invaders is implemented at all ($Q=100\%$).