This document is the Accepted Manuscript version of a Published Work that appeared in final form in: *Costantini, E.A.C.; Antichi, D.; Almagro, M.; Hedlund, K.; Sarno, G.; Virto, I.* 2020. Local adaptation strategies to
 *increase or maintain soil organic carbon content under arable farming in Europe: Inspirational ideas for setting operational groups within the European innovation partnership.* JOURNAL OF RURAL STUDIES.79. ©
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Local adaptation strategies to increase or maintain soil organic carbon content under
arable farming in Europe: inspirational ideas for setting Operational Groups within the
European Innovation Partnership

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#### 11 Abstract

12 In the European Union, the setting of Operational Groups (OG) is supported by the 13 European Innovation Partnership to tackle specific problems and favor innovation in 14 agriculture. They constitute an important aspect of the current Common Agricultural Policy. 15 Increasing or maintaining soil organic carbon (SOC) content under arable farming has been 16 acknowledged as a primary target of European agriculture. SOC-preserving agriculture needs 17 its techniques to be tailored to local conditions, namely, the combination of factors related to 18 the environment (climate and soil characteristics), to the farming system (land use type, farm 19 specialization, crop management), but also to the social and cultural context (market and 20 availability of production means, subsidies, farmers' education, propensity for innovation and 21 change), too. In this paper we present inspirational ideas and show success examples of local 22 adaptations strategies to increase or maintain SOC content in soils under arable farming in 23 Europe. They include:

24 • Adoption of soil management strategies to improve SOC storage in irrigated systems.

25	· Precision farming and other high-tech solutions able to generate local diagnosis and						
26	adaptive strategies for increasing SOC and reducing greenhouse gasses emissions.						
27	· Innovative strategies for extending soil cover periods and introducing cover crops in						
28	rotations in areas with limited water availability or prone to harsh weather conditions.						
29	• Management of rainfed and low input crops to maintain and increase SOC in dry						
30	climates and erosive prone soils.						
31	These case studies could facilitate the setting up of OGs and the application of innovative						
32	practices in different European countries.						
33							
34	Keywords: SOC; soil fertility; sustainable land management; conservation agriculture;						
35	cover crops						
36							
37							
38	Highlights						
39	- Rationale is given for considering local soil and climate in farming strategies						
40	- Limitations to increase SOC in arable land are described						
41	- Case studies illustrate possible references for Operational Groups						
42	- Diversification of cropping system as key factor to increase SOC						
43							

44 **1. Introduction** 

45 The Common Strategic Framework (EU Regulation 2013/1303) outlines the strategic guidelines and recommendations to be achieved by the European Union by 2020. Among the 46 47 Primary Objectives (Themes) of the Cohesion Policy there is promoting the adaptation to 48 climate change and the efficient use of resources. To this aim, measures to promote soil 49 organic carbon (SOC) sequestration in agriculture and forestry are supported by the pillars of 50 the Common Agricultural Policy (CAP). However, the Communication on the Future of Food 51 and Farming, recently delivered by the European Commission (COM (2017) 713 final), 52 clearly states that a one-size-fits-all approach to the CAP simply does not work in a Europe 53 where farms and farming conditions are so diverse. Despite intensive mechanization and 54 large supply of other technological inputs, the management of agricultural lands in Europe is 55 still very much differentiated. One of the main reasons for this differentiation is constituted 56 by the presence of large climatic and pedological differences, and varying environmental 57 limiting conditions. The main separations reflect the North-South gradient of moist-cold 58 Boreal, moist-temperate Oceanic, and dry-warm Mediterranean climates, with the 59 corresponding different limitations expressed by the length of plant growing period in the 60 North and the amount and length of plant water deficit in the South. Soil features and 61 constraints instead differ at a more detailed scale and form the basis for the success of adopted strategies to increase or maintain SOC at the farm and field All-regional levels. 62 63 The basic strategy to increase SOC stocks is through restitution of endogenous (e.g., crop 64 residues, wood litter, weeds) or incorporation of exogenous (e.g. animal manure, 65 depurationsewage municipal wastessludge, compost) organic matter to the soil. In this context, the recycling of organic wastes from domestic activities and urban areas as organic 66 67 fertilizers is an opportunity to transfer organic carbon in ways that enhance SOC storage, 68 ameliorate the nutrient content of soils and close nitrogen and phosphorus cycles at regional

69 scales (Rumpel et al., 2019). In addition to exogenous organic amendments such as compost 70 and manure, reduced or no tillage, improved management of crop residues and agro-industry 71 by-products, crop rotation, green manuring, and cover cropping are the most suitable 72 interventions to enhance SOC stocks in agricultural soils (FAO, 2017a). However, few 73 generalizations can be made of findings about sustainable agricultural practices, because their 74 effectiveness is inherently dependent on the local socio-economic, environmental and cultural 75 context (Henry et al., 2018; Rumpel et al., 2019; Sanz et al., 2017; Schoonhoven and 76 Runhaar, 2018). Local limitations due to climate and soil conditions can make standard 77 strategies to be ineffective when applied at these sites, or local growing conditions can 78 interfere with the efficiency of these strategies, suggesting the need of developing locally-79 adapted innovative strategies of adaptation.

80 In their analysis of the pros and cons of the 4p1000 Initiative, Rumpel et al. (2019) 81 highlighted how local conditions, and above all pedoclimate, land use and management 82 practices, may hamper reaching the targeted SOC stock gain. The same authors also indicated 83 that the levels of SOC inputs, soil-inherent pedologic characteristics and the state of soil 84 development are the major factors affecting the SOC stock potential of each single soil. It is 85 well known that this potential is the result of a new steady state reached by a soil when 86 management practices aimed at increasing SOC stocks are being applied. The level of this 87 new equilibrium conditions varies widely upon different soil and climatic conditions. 88 Normally, this process takes years or even decades, with differences among soils and 89 climates, and implies a decreasing trend of the SOC sequestration rate, meaning that long 90 term observations are needed to substantiate SOC storage potential of each single soil-climate 91 combination (Lal, 2008ref).

93 The Agricultural European Innovation Partnership (EIP-AGRI) was launched in 2012 to 94 contribute to the European Union's strategy 'Europe 2020' for smart, sustainable and inclusive 95 growth (https://ec.europa.eu/eip/agriculture/en/about). This strategy sets the strengthening of 96 research and innovation as one of its main objectives and supports a new interactive, multi-97 actornew interactive, multi-actorial approach to innovation. Farmers, advisers, researchers, 98 companies, NGOs, and other stakeholders are supported by the National and Regional 99 Rural Development Programs (RDPs) to form Operational Groups (OGs) aimed to create 100 innovations by tackling specific practical problems or opening new opportunities. The OG 101 approach makes the best use of different types of knowledge (practical, scientific, technical, 102 organizational, etc.), in an interactive and collaborative way. 103 The objective of this paper is to give the rationale for setting new OGs, present 104 inspirational ideas, and show success examples of local adaptations strategies to increase or 105 maintain SOC content in soils under arable farming in harshdifficult conditions. These case 106 studies could facilitate the adoption of innovative practices in other European regions, 107 showing similar environmental constraints. 108

**2. Diversity of environmental and management conditions in Europe** 

110 In Table 1 we summarized the results of the literature search performed to address the

111 <u>issue of the importance of adapting strategies to increase SOC storage in peculiar local</u>

112 conditions. Each management practice has been reported with related results in terms of SOC

113 <u>concentration, SOC storage and SOC storage rate.</u>

114

<u>Table 1 - Results from scientific literature and projects dealing with management practices</u>
 and soil organic carbon (SOC) concentration (g 100 g<sup>-1</sup>), stock (Mg C ha<sup>-1</sup>) and storage rate

- (Mg C ha<sup>-1</sup> y<sup>-1</sup>)Examples of successful case studies on the use of cover crops under limiting
   pedoclimatic conditions for a possible development into EU Operational Groups.
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- 121

#### 2.1 Land uses, climates, and soils

122 According to Eurostat (2018), agricultural land use is the most common primary land use 123 category in the EU-28, accounting for 41.1 % of the total area in 2015. Arable land made up 124 60% of the Utilized Agricultural Area (UAA) in 2013 with 104 million hectares, although the 125 distribution of the main types of agricultural land use types (arable land, permanent grasslands, permanent crops and kitchen gardens) varied widely between Member States. The 126 127 highest values observed in Finland and Denmark (98% and 91% of the UAA respectively). 128 The lowest share of arable land (21% of the UAA) was observed in Ireland, the United 129 Kingdom, Slovenia and Luxembourg (79%, 63%, 59% and 51% respectively). Close to half 130 of the utilized agricultural area (UAA) was reported in France, Spain, the United Kingdom, 131 and Germany. The lowest share of arable land (21% of the UAA) was observed in Ireland. 132 Within agricultural land, cropland covered, on average, aboutsome 22.2% of the total area, although the share of cropland varies greatly among countries and NUT2 territories (from 133 134 50.6% in Denmark to 9.5% in Slovenia and 1.2% in Sweden). In most member states, the 135 share of cropland was between 10% and 35% of overall land cover. Arable land made up 136 60% of the UAA in 2013 with 104 million hectares, although the distribution of the main 137 uses of agricultural land types (arable land, permanent grassland, permanent crops and 138 kitchen gardens) varied widely between Member States. The highest values observed in Finland and Denmark (98% and 91% of the UAA respectively). The lowest share of arable 139 140 land (21% of the UAA) was observed in Ireland, the United Kingdom, Slovenia and

Luxembourg (79%, 63%, 59% and 51% respectively). Finally, in relation to the production system, the total area under organic farming in the EU-28 was 11.9 million ha in 2016 and is still expected to grow in the coming years. The increase in organic area between 2012 and 2016 was 18.7%.

According to the European Soil Data Centre (ESDAC), in Europe there are at least 24
different <u>major</u> types of soil, this variety representing the variability in climatic conditions
and pedogenetic factors.

Considering the topsoil (0-30 cm), it has been estimated that 45% of European soils have a 148 low or very low organic carbon concentration (from 0 to 2 g 100 g<sup>-1</sup>) and 45% have a medium 149 150 concentration (2 to 6 g 100 g<sup>-1</sup>) (JRC, 2011). The databases reflect the broad scale influence 151 of climate on SOC, with a manifest decreasing gradient from the Boreal to the Mediterranean 152 climates. In fact, it has been largely demonstrated that drier and hotter climates favor SOC 153 depletion in agricultural lands (Pellegrini et al., 2018; Francaviglia et al., 2019). Lugato et al. 154 (2014b) estimated the content of SOC in European agricultural topsoils under different uses 155 conditions [A2]through a modelling approach, which allowed for upscaling single spot 156 measures of SOC content, taking into account also management practices and official 157 statistics. According to their predictions, arable land was predicted to storeck 7.65 Gt of SOC 158 (43% of total) in the first 30 cm of depth. The distribution of this SOC was however seen as 159 rather heterogeneous among territories. Another study on the potential for SOC sequestration 160 in European arable land (Lugato et al., 2014a) illustrated that, among land--uses changes that 161 do not imply converting arable land into other uses (i.e., grasslands), some strategies (i.e., ley 162 cropping systems and cover crops) seem to have a greater potential to increase SOC stocks 163 than others (i.e., straw incorporation and reduced tillage). The efficiency of these strategies 164 was however found to be highly variable in-across different regions (Lugato et al., 2014a).

Besides climate and soil typeology, soil degradation is acknowledged as a main driver of
SOC impoverishment (FAO, 2017a; Sanz et al., 2017). A review on major soil degradation
problems in Europe issuing from many-of these sources can be found in Virto et al. (2015),
who concluded that no single soil management strategy to cope with soil degradation is
suitable for all regions, soil types and soil uses.

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171

#### 2.2. Organic fertilization strategies

172 In addition to the returning carbon to the soil return of C from by decomposition of crop 173 residues or cover crops, supplementing the soil with carbon from external sources (i.e., 174 organic fertilizers, manure, composts, slurries, sewage sludge) is another complementary 175 strategy with huge potential to increase SOC in many cropping systems. In a their-review of 176 studies conducted under<del>produced in</del> Mediterranean conditions, Francaviglia et al. (2019) 177 reported that the application of external C sources to the soil has a great potential in terms of SOC storage rate, with highest results obtained by compost (+334.02-10<sup>-3</sup> MkMg C ha<sup>-1</sup> yr<sup>-1</sup> 178 179  $\times$  1000) and sewage sludge (+101.58  $\cdot$  10<sup>-3</sup> M-kMg C ha<sup>-1</sup> yr<sup>-1</sup>  $\times$  1000), whilst manure was less effective (+18.70•10<sup>-3</sup> M-kMg C ha<sup>-1</sup> yr<sup>-1</sup> x 1000) and slurry even negative (-0.07•10<sup>-3</sup> M-kMg 180 C ha<sup>-1</sup> yr<sup>-1</sup>  $\times$  1000), especially when combined with mineral fertilizers (-7.02 •10<sup>-3</sup> MkMg C 181 182 ha<sup>-1</sup> yr<sup>-1</sup> <del>x 1000</del>) (Table 1).

The use of organic amendments is normally linked to organizational and economic aspects of farm management and supply chains. The availability of farmyard manures or slurries is obviously connected to the presence of animal husbandry in the farm or in neighbor farms, whilst purchasing these products on broader markets is normally unviable for farmers due to their high volume and consequently high transportation and spreading costs. The choice to include or not animal production in the single farm or in <u>a local</u> network of <del>local</del> farms is normally an option linked to farm diversification strategies, <u>and</u> availability of <u>manpower as</u>
 well as local markets for animal products.

191 Nevertheless, given their high potential in terms of SOC storage rate, other external
192 sources of carbon as composts and sewage sludge could be valuable fertilization options for
193 farms not connected to animal production.

194

#### 195 **2.3. Tillage systems**

196 Reduced tillage is still a limited practice in Europe. In relation to tillage systems used in 197 arable land, inversion tillage (mostly ploughing) was the most wide-spread tillage practice in EU-27 in 2010, being practiced on almost two-third of the arable land in EU, while almost a 198 199 fifth was tilled with conservation practices, but pure zero tillage was not practiced in most 200 areas. Conservation agriculture in Europe, in particularalthough in increase, is estimated to 201 interest onlyabout 2.04Mha in 2013 (Kassam et al. 2015). In relation to tillage systems used 202 in arable land, inversion tillage (mostly ploughing) was the most wide-spread tillage practice 203 in EU-27 in 2010, being practiced on almost two-third of the arable land in EU, while almost 204 a fifth was tilled with conservation practices, but pure zero tillage was not practiced in most 205 areas. The share of arable land on which conservation tillage, which includes minimum soil 206 disturbance coupled with crop rotation and permanent maintenance of soil mulch cover 207 (http://www.fao.org/conservation-agriculture/en/), is applied also varied greatly among 208 countries within regions, for instance, from 0% in the Azores, Madeira, Malta and 209 Montenegro, to 65% in Thüringen (Germany) and the West Midlands (United Kingdom) in 210 2010. In a fifth of the regions, conservation tillage was practiced on more than 29.5% of the 211 arable land. In More recent estimations (year 2013) report that -2013, conservation agriculture 212 in Europe, although in increase, is estimated to interest cover [A3]about 2.04 Mha -(Kassam et

<u>al. 2015</u>). In most countries, the largest share of arable area on which conservation or zero tillage is applied was found on farms specialized in cereals, oilseed and protein crops.

215

#### 216 **2.4. Irrigation**

Since the 1990s, recurrent droughts in Europe, along with the increased need to enhance crops economic sustainability, have forced the implementation of additional irrigation, growing the proportion of the total cultivated area that is irrigated. The need of irrigation has become particularly important for the Mediterranean countries. At present, much of the food production (about 40%) in the Mediterranean area is associated with irrigation. The amount of water used accounts for 72% of the current freshwater withdrawals across the

223 Mediterranean area (Antonopoulos et al., 2017).

As described by Eurostat (2018), in 2013 the total irrigable area in EU-28 was 18.7 million ha (11.3% total agricultural area), although only 10.2 million ha (6.2% of total) was-were actually really irrigated. These values are in fact the result of an expansive trend, which represents an average increase of 13.4% since 2003 in all countries except Portugal. In particular, Spain and Italy increased their irrigable area by 19.7 and 15.5%, respectively. In the context of climate change, irrigation demand in Southern Europe is projected to <u>further</u> increase (Füssel et al., 2017).

231

#### 232 **2.5 Precision farming**

High Tech Farming can be a useful tool to maintain SOC, using best available
technologies to adapt crop and soil management to specific conditions and variability at field
scale. Site-specific management through precision agriculture (PA) can lead to optimization
of quantity and quality of crop yields, whilst local variability of SOC is detected and
corrected by means of fertilization, applied on a detailed soil-by-soil basis. The contribution

238 of PA can be referred to a series of tools sharing an integrated use of Information and

239 Communication Technologies (ICT) to identify and properly manage small areas with

240 homogeneous conditions, for instance in terms of soil fertility.

#### **3. Effect of crop type and management on SOC at the national and regional scale**

#### 242

#### 3.1 The influence of different crop types on SOC

243 Many authors suggest that changes in land use and crop management might be responsible 244 for the variations in SOC at the country and regional levels (Smith et al., 2012).

Comprehensive studies recently published online (FAO, 2017b) indicated marked variations in organic carbon stocks among crop types.- In particular, topsoils (0-30 cm) of arable lands at the global scale show mean values of  $51.0 \pm 0.66$  Mg C ha<sup>-1</sup>, while rice fields and forests stands-account for  $55.6 \pm 3.01$  and  $71.1 \pm 2.10$  Mg C ha<sup>-1</sup> respectively. Most authors agree that, under equal pedoclimatic conditions, SOC content is generally smaller in cropland soils than in forests, grasslands or shrublands (see, for instance, Pellegrini et al., 2018).

251 A revision of the values given for SOC concentration in the first 30 cm of Spanish soils 252 showed a high variability in the medians, from 0.82 g SOC 100 g soil<sup>-1</sup> in fallow areas to 1.24 and 1.29 g SOC 100 g<sup>-1</sup> in horticultural land (mostly under irrigation) and grain legumes, 253 254 respectively (González-Sánchez et al., 2018) (Table 1). The comprehensive study of the 255 potential of different management strategies to increase SOC in Spain (González-Sánchez et 256 al., 2018) also showed a great influence of crop types, farming systems and/or crop rotation. 257 The information stored in the Italian soil database confirms the influence of land use type 258 on soil carbonSOC stocks (Costantini and Lorenzetti, 2013). Recently updated studies for the 259 Italian map of soil organic carbon, published in the framework of the FAO global assessment 260 (FAO, 2017b), indicate mean SOC concentration values (0-30 cm) in paddies and other 261 arable lands (urban soils included) between 1.16 and 1.33 g 100 g<sup>-1</sup>, between 1.74 and 2.26 g

100 g<sup>-1</sup> in meadows and other less intensively or not cultivated areas, whereas in different
kinds of woodlands and natural areas they can reach hig<u>1</u>her values up to 3.48 g SOC 100 g<sup>-1</sup>
(<u>Table 1</u>). Though, the large values of standard deviation indicate that variations of land
management and local conditions play a great role in regulating SOC.

In NW Portugal, in temperate areas with adequate rainfall for winter cereal cultivation, recent trials on the introduction of legume crops in a rotation of winter cereals have resulted in no gains in SOC after three years (Oliveira et al., 2019), very likely because of soil conditions including low clay contents, and low-reactive minerals in the clay fraction.

270

#### 3.2 The influence of tillage on SOC

271 In some cases, strategies widely adopted and recognized to increase SOC content in arable 272 land, such as no-till adoption or the diversification of rotations by including legume crops, 273 seem not to get the expected responses (Dimassi et al., 2014; Oliveira et al., 2019). Within 274 no-till systems, it has been observed that the gain in crop productivity associated to the 275 adoption of no-till is the major driver of SOC gains, explaining more than 30% of the 276 observed increment in a worldwide meta-analysis (Virto et al., 2012). Recent research for the 277 Mediterranean region (Francaviglia et al., 2019) supports this view. Crops sensitive to this 278 changebenefiting from shifting from conventional tillage to conservation tillage in terms of 279 productivity seem therefore more effective for SOC enhancement under no-tillage systems. 280 Usually, these crops are represented by summer crops (e.g. maize, soybean) when grown in 281 areas prone to drought stress. Compared to inversion tillage, the application of no-till through 282 the direct sowing of these crops in dry conditions can result then in earlier and better crop 283 establishment, higher nutrient availability (mediated by faster mineralization of organic 284 matter in the hottest season) and then also higher yields if sufficient water availability is also 285 ensured across the season (Pareja-Sánchez et al., 2019). Nevertheless, besides increasing

286 plant biomass-derived C inputs (i.e. crop residues returning to the soil), reducing SOC 287 mineralization rates (i.e. the major SOC output) by tillage operations can also play a major 288 role, especially in Mediterranean areas prone to high oxidative conditions (Mazzoncini et al., 289 2011). Reduced or nil response to no-till adoption in the long-term has been also verified in 290 loam soils in temperate areas of Europe with extensive wheat and maize cropping (e.g. 291 Dimassi et al., 2014). This result has been associated to the lack of response of crop 292 productivity to tillage strategies, as well as to the existence of a climate with a positive water 293 balance inducing mineralization of crop residues left at the soil surface. As crop productivity 294 is also the result of a successful crop protection, low or null increase in crop yields observed 295 in no-till and minimum tillage systems under different pedoclimatic conditions should be 296 related to difficulties in controlling weeds, especially perennial species, pests and diseases 297 (Chinseu et al., 2019). Spatial and temporal variability in soil and climatic conditions clearly 298 plays a key role in the selection of target noxious organisms in each specific case. 299 Other technical barriers actually hindering a wider adoption of no-till among farmers in 300 Europe could be identified in peculiar combinations of soil characteristics and unavailability 301 of proper machinery (i.e. direct drilling machines or machinery to manage crop residues or 302 cover crops) (Sanz et al., 2017). For instance, it is well known that soils prone to crust and 303 heavy compaction in topsoil (e.g. soils rich in silt, or with low ability of the soil structure to 304 regenerate naturally because of a high content of illite-type clay, which has low shrinkage-305 swelling capacity) are not well adapted to the direct sowing of many crops, above all small 306 seeds crops (Sasal et al., 2017). Due to their weight, direct drilling machines may cause soil 307 compaction themselves, hampering seeds to germinate and plantlets to establish well 308 (Chinseu et al., 2019), and need to be adapted to specific soil conditions to reduce soil 309 compaction (e.g. by mounting shanks in front of the furrower), but then increasing purchase 310 costs. Also clay soils on hillslopes can be difficult to manage under no-tillage, due to

difficulties in field operations (e.g. powerful tractors are needed due to high traction forces,
the high clay content make narrower the windows where the soils can be seeded) which
translate into frequent poor crop establishment and yield.

314 Besides these technical constraints, there is a number of other barriers faced by farmers 315 willing to adopt conservation agriculture practices. First of all, the absence in specific regions 316 of financial incentives or subsidies to motivate or compensate farmers for possible yield 317 losses. No till normally encompasses the use of agrochemicals (pesticides but also mineral 318 fertilisers fertilizers), which makes its environmental impact less positive and sometimes 319 farmers cannot afford the costs because crop yields also are also reduced or maintained in the 320 short term, which makes the economic balance negative, at least in the short run (Sanz et al., 321 2017; Ingram et al., 2014).

From <u>a</u> socio-cultural point of view, in some areas no-till conflicts with an important cultural symbol for hard work, as tillage is generally believed to <u>symbolisesymbolize</u> a hard worker, and with the social recognition that a field properly ploughed is "clean" (Chinseu et al., 2019; Schoonhoven & Runhaar, 2018).

Uncertainty about the weather, policy and market developments in addition to internal farm
factors (such as debt, tenure, and family status) are other important barriers to overcome.

328

#### 329 **3.3 SOC in irrigated and non-irrigated arable lands**

In relation to soils and organic C storage, the adoption of irrigation has different potential effects, including alterations of the organic C cycle (Entry et al., 2002; Denef et al., 2008), as it can increase the amount of organic C entering the soil through a greater plant productivity, but can also favor mineralization by providing moisture and thus stimulating microbial activity. Irrigation is also associated to some intensive cropping systems such as vegetables,

which need intensive and frequent tillage operations and high fertilization inputs, which canconcur to alter the carbon cycle.

337 The consequences of these practices have been reported to affect the soil chemical fertility 338 (McDowell et al., 2011), its physical condition and biological indicators (Manono and 339 Moller, 2015), which can indirectly affect the stabilization of SOC. Soil quality indicators sensitive to management can also therefore change when dryland is converted to irrigation 340 341 (Apesteguía et al., 2017). As pointed out by Chenu et al. (2019), different observations 342 suggest that these alterations are site- and management-dependent, as the changes observed in 343 SOC stocks are not always directly related to the increment observed in crop yields (Follett et al., 2013). 344

345 The effect of irrigation on SOC has been indeed observed to vary according to local 346 conditions and specific water and soil management. In Mediterranean Europe, soil C losses 347 associated to the implementation of irrigation have been reported in Portugal (Nunes et al., 348 2007) as well as in Italy (Costantini and Lorenzetti, 2013). The negative effect of irrigation 349 might be related to the consequent intensification of agricultural management and it is 350 particularly evident under warm and dry climates. The data stored in the Italian soil database 351 indicate for the regions of central and southern Italy lower SOC values for all irrigated crops, 352 particularly for vegetables, row-crops, and orchards (Costantini and Lorenzetti, 2013). 353 Modeling scenarios also point out the possibility of C losses upon a wider irrigation 354 adoption in the long term in terms of acreage (Álvaro-Fuentes and Paustian, 2011; Muñoz-355 Rojas et al., 2017). A study conducted in Navarre (NE Spain) showed that the turn-over rates 356 of organic C can be accelerated in the short-term, very likely because of changes induced in 357 the shoot-to-root ratios of some crops, and the less limiting conditions for soil C 358 mineralization (Apesteguía et al., 2015) when irrigation is adopted.

359	Some other studies have shown a positive effect of irrigation on SOC. Aguilera et al.					
360	(2013b), when comparing organic and conventional cropping systems in Mediterranean					
361	conditions, observed that SOC stock increment in organic systems was greater under					
362	irrigation than under rainfed conditions (25% vs. 13% increase over conventional,					
363	respectively).					
364	In summary, we can conclude that					
365						
366	the overall consequence of the described As a summary, interactions between soil, climate,					
367	crop type, and agricultural management, make is that that strategies to ensure SOC					
368	stabilization in arable land need $fora$ local assessments.					
369						

## 370 4. Needs for adaptation at the local scale: problematic soil, climate, and managementβ71 conditions

372 <u>In Table 1 we summarized the results of the literature search performed to address the</u>

373 <u>issue of the importance of adapting strategies to increase SOC storage in peculiar local</u>

374 <u>conditions. Each management practices has been reported with its ownrelated results in terms</u>

375 <u>of SOC concentration, SOC storage and SOC storage rate.</u>

#### **4.1 C management in soils subjected to wind erosion**

377 Wind erosion affects both the semi-arid areas of the Mediterranean region as well as the

- temperate climate areas of the northern European countries (Borrelli et al., 2018).
- The North-West parts of Europe are the most vulnerable to wind erosion and almost 40% of
- the agricultural area in Denmark is deemed to be affected (Riksen & De Graaff, 2001). This
- depends on the combination of wind trades, light soil texture, and intensity of the farming

382 practice. As most of the carbon is lost with the shallowest eroded soil (Borrelli et al., 2016), 383 The soil lost through wind erosion can counteract the mitigation of soil carbon sequestration -384 if the management is not also adapted to limit erosion problems, as most of the carbon is lost 385 with the shallowest eroded soil (Borrelli et al., 2016). To mitigate wind erosion, the soil 386 surface needs to be covered throughout the year and it is suggested to use -conservation 387 tillage for promoting aggregate stability of the soils. This will at the same time increase the 388 carbon sequestration of the soils, thanks to the inputs of C from plant residues and roots. Also 389 organic amendments would contribute a lot to improve structure stability and to increase 390 SOC conservation. Unfortunately, we are not aware of any case study reporting success 391 stories combining agricultural practices aimed at reducing wind erosion and at the same time 392 increasing SOC in arable lands. Setting Operational Groups to address this issue in 393 representative lands, especially of Northern Europe, would be highly recommended. 394

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#### 395 **4.2 C management in water eroded soils**

396 Enhancement of soil organic carbon in water eroded soils <del>could</del>-can be effective <del>also</del>-to 397 increment soil water infiltration and storage capacity and as well as to reduce soil and water 398 losses by erosion (FAO, 2017a). However, halting soil erosion in sloping lands is a 399 fundamental prerequisite. Actually, it must be considered that the topsoil keeps the major part of SOC. For instance, a soil loss of 10 Mg ha<sup>-1</sup> y<sup>-1</sup> in a soil with a bulk density of 1.4 g cm<sup>-3</sup> 400 401 corresponds to an annual loss of about 0.5% of the carbon stock of the first 30 cm, and the 402 map of soil erosion of Europe shows many parts having erosion rates far larger than 10 Mg ha<sup>-1</sup> y<sup>-1</sup> (<u>https://esdac.jrc.ec.europa.eu/themes/erosion</u>). The visual comparison of the maps of 403 404 soil erosion rates and SOC contents in Europe reveals a strong inverse relationship, especially 405 in the Mediterranean countries (https://esdac.jrc.ec.europa.eu/themes). Many experimental 406 data confirm the relationship between high soil erosion rate and low SOC content in different

arable and tree crops, and the importance of preventing soil erosion to improve the ability of
the agro-ecosystems to incorporate SOC (Le Bissonnais et al., 2002; Cerdan et al., 2010;
Costantini et al., 2018; Chenu et al., 2019).

410 The success of the wide variety of soil conservation practices to be adopted will depend on 411 local combinations of soil type, climate, and management practices, together with the socio-412 economic context. In a comprehensive overview on different soil and water conservation 413 techniques in Europe and the Mediterranean, it was found that crop and vegetation 414 management (i.e., cover crops, mulching, grass buffer strips) and mechanical techniques (i.e., 415 terraces, contour bounds, geotextiles) were more effective in reducing annual runoff and soil 416 loss rates than soil management (i.e., no tillage, reduced tillage, contour tillage, deep tillage, 417 soil amendment) -(Maetens et al., 2012). Regarding soil management techniques, it was also 418 found that no tillage and conservation tillage become less effective in reducing annual runoff 419 - but not annual soil loss - over time. A conclusion drawn from this meta-analysis was that 420 the more erosion-prone conditions are (i.e., erodible soils, steeper slopes, areas with high-421 intensity low-frequency rainfall events occurrence), the most effective in reducing runoff and 422 soil erosion rates these soil and water conservation techniques -are. Arable cropland can be 423 considered among the most erodible land uses, so that any management ensuring a permanent 424 soil cover could result in a dramatic reduction of actual soil loss rates (Panagos et al., 2015) [A4] [A5]. 425 On the other hand, due to their usual low SOC content, eroded soils can be considered as 426 show higher SOC sequestration rates in the short term, potential new C sinks for C as C 427 incorporation is potentially quicker faster just in soils where SOC values are lowerpoorer in 428 SOC (Francaviglia et al., 2019).

429

#### 430 **4.3 C management in soils with shallow groundwater and limited drainage**

431 Some agronomic strategies to increase SOC content in arable soils, such as the adoption of 432 no-till, have been observed to be unsuitable to poorly drained soils (Soane et al., 2012), as 433 well as in soils with shallow groundwater (Costantini and Dazzi, 2013). In poorly drained 434 soils, reasons for this unsuitability are related to the difficulty of soil management when the 435 seeding dates coincide with the rainiest season. Areas with poor drainage also result in low 436 productivity, as water excess in topsoil limits the presence of oxygen in soil pores, hampering 437 soil microbial activity, nutrient availability (e.g., the inhibited activity of nitrifying bacteria 438 can reduce nitrate concentration in the soil) and root functioning. A reduced crop productivity 439 will also result also in reduced C return to the soil. Areas of winter cereals grown on soils 440 with high content of clay and/or silt in Mediterranean climates, with the highest precipitation 441 peaks in fall, are good examples of soils of that kind. If winter crops follow a spring crop 442 harvested in the fall, problems of poor drainage may become extremely severe, thus 443 preventing the adoption of reduced tillage.

444 Gleysols (soils affected by groundwater) share the same lowland environment of Histosols 445 (organic soils) and are still rich of organic carbon. Many Glevsols, in particular, are just 446 degradation forms of Histosols caused by reclamation activities, in particular, drainage, 447 addition of mineral material, and repeated ploughing, which mineralized the most elaborated 448 parts of the organic matter. Therefore, the management of organic matter in Gleysols goes 449 along with that of groundwater (Costantini and Dazzi, 2013) as drainage can at the same time 450 facilitate the adoption of cropping strategies increasing crop productivity, but also SOC 451 losses through mineralization.

452 Some soil management strategies have been developed in these areas to improve drainage 453 and allow for higher productivity. Surface drains are common in many of these areas and 454 have to be done every year before seeding. Subsurface de-compaction can be done with 455 different types of subsoilers, and subsurface drainage can also be improved by mole-ploughs.

Some experiences in NE Spain (Pérez de Ciriza, personal communication) have shown that these techniques can efficiently improve soil conditions at seeding, and therefore allow for increased productivity and the adoption of some strategies of conservation agriculture. A group of arable farmers from NE Italy (Life HelpSoil, <u>http://www.lifehelpsoil.eu/</u>) -even demonstrated to be able to apply continuous no-till for several years in clay soils thanks to occasional subsoiling.

- 462
- 463 **4.4 C management in stony soils**

Rock fragments are frequently found as a part of the soil volume in many areas. Soils with
rock fragments are usually less productive than other soils in the same conditions of texture.
Their presence at the soil surface can limit both crop development and soil management by
impeding tillage. Direct seeding can be impossible in stony soils.

In arid and semi-arid land, the reduced water-holding capacity has led stony soils to be considered marginal soils for agriculture since long ago (Arias et al., 2017). In terms of <u>SOC</u> stock, their limitations arise from their reduced primary productivity when rock fragments occupy a significant part of the soil volume, and also from their reduced proportion of fine soil able to effectively protect organic C from mineralization.

473 Different management strategies can be adopted on these soils to overcome these
474 limitations, especially in semi-arid and arid lands. However, the interaction between
475 conservation agriculture and soils with rock fragments has been seldom addressed, and its
476 adoption can be challenging in this context (e.g., Schwilch et al., 2015).

A recent study on the effect of the simultaneous adoption of irrigation and no-till in a
stony soil in NE Spain (Arias et al., 20174) [A6] has shown that these soils can be reactive to
the improved <u>carbon</u> inputs from crop residues resulting from the combination of irrigation
with no-till in the very short-term (2 years). In particular, an increment of 10 Mg SOC ha<sup>-1</sup>

481 (from  $37.9 \pm 0.7$  to  $47.9 \pm 1.1$  Mg SOC ha<sup>-1</sup>) in the upper 30 cm (after correction for stone 482 content using the hybrid method for bulk density described by Throop et al., 2012) was 483 observed after <u>3-2</u> years (Table 1).

484

#### 485 **4.5 C management in saline, gypsiferous and alkaline soils**

Many soils in arid and semi-arid areas display significant accumulations of soluble salts in the upper horizons. Although these soils are generally not common in arable land, some are cultivated in marginal areas of Southern Europe. More frequently, salinization occurs as a result of agricultural management (especially irrigation). These soils are naturally not suited for SOC stabilization. For instance, Solonchaks displayed the lowest values of SOC (11.1 Mg SOC ha<sup>-1</sup>) in the upper 25 cm among all arable soil types in Andalusia (Muñoz-Rojas et al.,

492 2012).

493 As a result of high sodium percentage on the cation exchange complex lets the clay 494 particles lose their tendency to stick together when wet. Soils became impermeable in depth 495 to both water and roots, and geomorphologically unstable. Therefore, SOC management in 496 these soils must be accompanied by important measures to prevent soil water erosion. 497 Gypsisols are not frequent in Europe, being the dominant soil type in less than 0.1 % of 498 the total area, mostly concentrated in the Ebro Basin and other areas of Spain, as well as in 499 Sicily. They are limited for SOC concentration (between 0.4 and 1 g 100 g<sup>-1</sup> for Ap horizons) 500 because of the low reactivity of its mineral fraction, their low water-holding capacity and 501 other physical limitations. Their location in arid areas makes them unsuitable for agriculture 502 without irrigation (Herrero, 2017). Otherwise, they are mostly not used, or used for extensive 503 grazing. Fertilization needs to be used at higher rates than usually calculated from crop needs. 504 With irrigation, drainage and heavy fertilization, satisfactory yields of gypsum-tolerant crops

can be obtained. Increasing SOC in these soils can be particularly challenging, as their
physicoal-chemical properties, which can impose limitation on water retention (MoretFernández and Herrero, 2015), and their mineralogical compositions do not favor the
development of stable soil structure or organic matter complexation.

509

#### 510 **4.6 A stock to be protected: SOC in black soils and peats**

511 Peat lands are common in Northern Europe, due to wet and cold climate, while are 512 marginal but still present in Mediterranean countries. In both cases peatlands have been 513 drained for a long time to get agricultural land and by that they are prone to mineralisation at 514 a rate leading to a loss of from 0.5 to 4 cm of peat soil per year (Regina et al., 2016). This 515 means that the 10-20% of peatlands drained for agriculture and forestry are currently is now 516 losing carbon and producing net greenhouse gas (GHG) emissions such as CO<sub>2</sub> and N<sub>2</sub>O, 517 whilst CH<sub>4</sub> decreases with drainage (Berglund and Berglund, 2010; Regina et al., 2016). 518 In most peat lands, the agricultural exploitation is limited to the most marginal lands, 519 which in some cases are converted to forestry, but still large areas in the Nordic countries are 520 cultivated and will lose carbon if they are not managed properly (Regina et al., 2016). 521 Suggestions on how to keep the carbon in these soils are either to keep the soil covered 522 with grasslands, perennial crops instead of annual cropping, but also by rewetting, so that the 523 decomposition of the peat slows down (FAO, 2012; Rumpel et al., 2019). 524 Black soils, or highly base-saturated mineral soils, rich in organic carbon (Phaeozems, 525 Chernozems, Kastanozems) are frequent in eastern Europe but are also present in 526 Mediterranean countries, where soil erosion is not intense and summer drought limits the 527 mineralization of the soil organic matter (i.e., the so-called "Mediterranean steppe"). A 528 common relevant feature of black soils is the relatively high organic carbon content of the 529 topsoil, but also of the subsoil, , reaching values of where values of SOC concentration of

530 0.5% in weightw/w even until at 1 m deep are very frequentand more, which confirms the 531 great potentiality for carbon sequestration of this kind of soils. Actually, the SOC 532 concentration in topsoil of modal Italian Phaeozem, Chernozems and Kastanozems is 1.75, 1.34 and 1.29 dag kg100 g<sup>-1</sup>, respectively, while density A7-SOC storage in the first meter is 533 respectively 12.69126.90, 127.60.76 and 12.141.40 kg Mg C mha<sup>-21</sup> in the first meter, 534 respectively, and reaches 17.181.80, 12.99.00 and  $13.66.90 \text{ kg-Mg m}ha^{-2}$  in the whole 535 536 profile (Costantini and Dazzi, 2013). As shown by the results obtained in China by Liu et al. 537 (2003), the SOC of black soils can be restored by adopting the right crop rotation and an 538 intensive return of organic material to the soil through amendments. To the best of our 539 knowledge, we are not aware of any study produced in European black soils.

540

#### 541 **4.7 Organic soil amendment and organic farming conditions**

542 Although the relationship between the application of exogenous organic matter to soil and 543 the gains in SOC is very clear in the reviewed literature (e.g., Dignac et al., 2017), attention 544 must be paid on associated GHG emissions from soils (Aguilera et al., 2013a). In their 545 review, Aguilera et al. (2013a) clearly showed how, for instance, that the typology of organic 546 amendment itself can have a strong effect on the level of N<sub>2</sub>O emissions. Liquid organic 547 amendments (i.e., slurry), for instance, were reported to lead to N2O emissions comparable to 548 those produced by mineral fertilisers fertilizers, whilst the application of solid forms of 549 organic amendments, especially when coupled with use of cover crops, resulted in 550 significantly lower N<sub>2</sub>O emissions and higher SOC stocks (Aguilera et al., 2013b). 551 In addition, the interplay between increasing SOC content without substantially increasing 552 the emission of GHGs becomes a challenge and depends on the crop and soil types and 553 management option.

-In this context, fine textured and poorly drained soils are particularly prone to, respectively, limited SOC incorporation and high GHG emissions.

-Sandy soils have limited capacity to store SOC due to usually low organic matter return,
on one hand, and high SOC mineralization rates, on the other hand. Low return of organic
matter is normally typical of sandy soils due to their lower chemical fertility and lower
productivity compared to loam or clay soils, which normally implies alsoalso implies a lower
return of crop residues into the soil. High aeration and fast mineralization of organic matter
normally contribute to keep low the content of SOC in sandy soils, but alsosimilarly increase
the risk of high GHG emissions.

563 -Poorly drained soils (e.g. soils with high silt content and prone to shallow crust or clay 564 soils with low permeability) are also-well known to have limited capacity of SOC stock storage due to typically poor establishment of the crops and consequently low crop residue 565 566 return. On the other hand, N<sub>2</sub>O and methane-CH<sub>4</sub> emissions from these soils can be high due 567 to low aeration (Krichels et al, 2019). To overcome these limitations, the improvement of 568 soil drainage (Kumar et al., 2014) is of paramount importance. Furthermore, poorly drained 569 soils can be managed through liming practices aimed to increase soil nutrient availability for 570 plants and to enhance soil microbial activity and N<sub>2</sub>O reductase, which counteracts the 571 emission of N<sub>2</sub>O from the soil, can contribute to reduce GHG emissions from poorly drained 572 soils (García-Marco et al., 2016)-or drainage (Kumar et al., 2014). 573 574 Apart from all the above mentioned environmental limiting conditions for applying

#### Apart from an the above mentioned environmental miniting conditions for apprying

575 <u>organic fertilizers, there are other important barriers such as the lack of access to manure or</u>

576 <u>organic wastes and the lack of technical know-how for proper processing of manure before</u>

577 <u>applying it to the field, causing weed infestation and pest occurrence and/or increasing labour</u>

578 <u>demand and costs of implementing it (Chinseu et al., 2018).</u>

579

#### 580 <u>4.8 Organic farming</u>

581 As for the management options, organic farming can reach a good trade-off between the 582 instances of high SOC increase and reduced GHG emissions and is increasingly adopted by 583 European farmers. Organic farming, as defined by the latest EU Regulation 2018/848, is "an 584 overall system of farm management and food production that combines best environmental 585 and climate action practices, a high level of biodiversity, the preservation of natural resources 586 and the application of high animal welfare standards and high production standards in line 587 with the demand of a growing number of consumers for products produced using natural 588 substances and processes". Reduced use of synthetic external inputs (mainly mineral 589 fertilisers fertilizers and pesticides) and augmented return of organic matter to the soil 590 (through organic amendments and fertilisers fertilizers, green manures, and crop residues) are 591 the most relevant farming practices with respect to the objectives of increasing SOC increase 592 and reduced-reducing GHG emissions (Aguilera et al., 2015). Nevertheless, the magnitude of 593 SOC increase and GHG reduction that organic farming management can achieve strongly 594 depends on other variables, such as crop type and management intensity. Despite organic 595 inputs are on average higher in organic than in conventional agriculture, often in this type of 596 soil management there is the a need to till the soil frequently in order to avoid the use of 597 chemical herbicides to control weeds-without chemical herbicides, and this can cause 598 significant SOC losses by erosion and mineralization processes (Stavi et al., 2016). 599 For example, dData from Spain revised by Aguilera et al. (2015) showed how in rainfed 600 cereals (wheat, barley), business as usual conventional management led to higher N<sub>2</sub>O 601 emissions (mostly due to the exclusive use of mineral fertilisers fertilizers) compared to 602 organic management, while soil carbon sequestration rates were similar between both types 603 of management practices (Aguilera et al., 2013b). However, the low use of synthetic inputs

604 under conventional legume management leads to similar GHGs balance when comparing 605 with organic management. On the contrary, although SOC content could be increased by 606 means of incorporating rice straw and manures to the soil in rice fields, the increase in CH<sub>4</sub> 607 emissions derived from these practices could not be overcome by the enhancement of SOC 608 stock (Aguilera et al., 2015). As horticulture requires high input management in terms of 609 irrigation, fertilizers and pesticides, there is a potential for increasing SOC stocks and 610 mitigate GHGs emissions when conversion to organic management is adopted. According to 611 Aguilera et al. (2015) estimations, a decrease of 59% of GHGs emission while a three-fold 612 increase in carbon-SOC stock can be reached per ha when passing from conventional to 613 organic management in horticultural cropping systemse. 614 Apart from all the above mentioned environmental limiting conditions for applying 615 organic fertilizers, there are other important barriers such as the lack of access to manure or 616 organic wastes and lack of technical know-how for proper processing of manure before 617 applying it to the field, causing weed infestation and pest occurrence and/or increasing labour 618 demand and costs of implementing it (Chinseu et al., 2018). despite organic inputs are on 619 average higher than in conventional agriculture, often in this type of soil management is 620 needed to till the soil frequently to control weeds without chemical herbicides, and this can 621 cause significant SOC losses.

#### 622 **5. Case studies and potential EIP-AGRI operational groups**

<u>Table 2 summarizes the results achieved by selected good practices tested in Hereby we</u>
 identify interesting case studies concerning local adaptation of strategies intended for<u>aimed to</u>
 <u>SOC</u>-increase <u>SOC</u> in different pedoclimatic and agronomic conditions. For each case study,
 the most relevant research gaps are identified. As EIP-AGRI OGs could actively contribute to
 complement research activities by implementing the most promising practices, and collecting

- 628 <u>and through</u>-validating on and data collection at field scale<u>farm</u> level, also ideas for potential
  629 OGs are illustrated.
- 630

631 <u>Table 2 - Management practices and SOC: state of the art, identified good practices,</u>

632 <u>research gaps, ideas for EIP-AGRI Operational Groups (OGs)</u>

#### 633 5.1 Management of irrigated crops to increase SOC in dry climates

634 In irrigated systems, the combination between irrigation and cultivation strategies can have 635 contrasting effects on SOC stocks, in comparison with dryland management (Chenu et al., 636 2018). Information is needed on the influence of the alteration of the soil water regime on 637 SOC cycling and stockstorage. This includes research to fill the gaps in understanding SOC 638 dynamics and its determinants at all scales from basic soil processes to landscape- and regional scale. EIP-AGRI OGs could supply consistent<sup>[A8]</sup> models for irrigation applied to 639 640 different cropping systems, suitable to be disseminated in similar conditions (e.g. EIP-AGRI, 641 2019).

An example of project aimed at demonstrating how strategies for climate change

643 mitigation in irrigated agriculture can be developed in drylands was conducted in Navarre

644 (NE Spain), between 2013 and 2016 in the regional-scale project Life Regadiox (<u>http://life-</u>

645 <u>regadiox.es/en/)</u>, led by a regional farmers association (Fundagro, https://uagn.es/fundagro/).

646 As part of the work was to quantify climate change mitigation, an inventory of SOC stocks in

647 the most representative cultivated soil types under dryland and irrigated agriculture was

648 conducted.

649 The results of this project can be summarized as follows:

650 - Overall, compared to rainfed conditions, irrigation resulted in a greater SOC storage in
651 the tilled layer (0-30 cm). The extent of SOC gain upon irrigation varied among sites, ranging

from  $\pm 19.2 \pm 1.4$  Mg SOC ha<sup>-1</sup> in one site more recently transformed to irrigation (9 <u>years)4 ± 2.5 Mg SOC ha<sup>-1</sup> in the site most recently transformed to irrigation</u>, to  $\pm 42.3 \pm 2.8$ Mg SOC ha<sup>-1</sup> in the site with 20 years of irrigation and NT corn under irrigation, both on a Haplic Calcisols (Antón et al., 2019) (Table 1);

656 - Within irrigated systems in arable crops, great differences were observed, mostly related 657 to the intensity of cultivation. For instance, within one site, the soil under horticultural crops with two crops per year and little residue restitution stocked  $68.3 \pm 2.7$  Mg SOC ha<sup>-1</sup> in the 658 659 upper 30 cm after 20 years of irrigation, whereas no-till corn stocked 99.6  $\pm$  5.6 Mg SOC ha<sup>-</sup> <sup>1</sup>, and rainfed organic wheat  $74.5 \pm 5.5$  Mg SOC ha<sup>-1</sup>. In this sense, it is noteworthy that the 660 661 introduction of no-till in irrigated land is less frequent than in rainfed semi-arid areas, where 662 one of the main reasons for introducing no-till is the optimization of water retention. Also, alfalfa stocked  $63.1 \pm 4.0$  Mg SOC ha<sup>-1</sup> after 6 years of continuous cropping with irrigation in 663 664 an area where rainfed cereals on the same soil had  $43.9 \pm 2.3$  Mg SOC ha<sup>-1</sup>. In a previous 665 study (Virto et al., 2006), the effect of irrigation with wastewater from vegetable canning 666 industry, which contained moderated amounts of organic C mostly in the form of particulate 667 organic C, was observed to be inexistent or very low on SOC, compared with that of the 668 implementation of a permanent alfalfa crop.

669 In arid and semi-arid conditions, inclusion of permanent crops, less intensive crop 670 rotations and tillage strategies seems therefore the major driving variables determining the 671 possibilities to increase SOC stock in soils when transformed from dryland to irrigated. 672 Nevertheless, it must be highlighted that in dry conditions permanent non-woody crops, and 673 especially pastures and forage crops, can be profitably grown only if irrigation is 674 implemented. In this case other crops may result more profitable, hindering their wider 675 adoption by farmers. OGs should consider ways to make irrigated permanent crops more 676 profitable and appealing for farmers, e.g. by opening new market opportunities (e.g. alfalfa

677 protein concentrate), estimating the amount of subsidies to be paid under RDPs, or testing 678 cultivation techniques oriented to increase the yield of such crops in rainfed conditions. 679 Making irrigated systems more effective in terms of SOC stock increase may also imply a 680 redesign of the proper irrigation system, aiming at reaching the best tradeoff between 681 production-related targets and soil conservation. For instance, the combination between no-682 till and sub-irrigation seems very promising in overcoming some technical constraints typical 683 of no-till (e.g. limited deepening of crop roots, with consequently scarce water uptake due to 684 lower water infiltration compared to tilled soils, high weed competition for water supplied on 685 topsoil).

686 On top of that, environmental and socio-economic barriers need to be assessed to evaluate the expected net effects associated to SOC increments in comparison to non-irrigated systems 687 688 or previous condition under irrigation (Antón et al., 2019). In addition to the profitability 689 issues described above, the net balance in GHG emissions and other possible outcomes of 690 irrigation (soil loss, nutrients leaching, salinization, etc.) are to be considered. While some 691 basic consequences on SOC cycling as affected by irrigation adoption are still unclear and 692 need more research (Chenu et al., 2019; Rumpel et al., 2019), the technical strategies to 693 overcome these limitations (such as erosion control or fertilization management) are 694 promising topics for OGs, in addition to those described above.

695

## 5.2 Management of rainfed and low input crops in dry climates and erosion prone soils

In dry climates and poor soils, the enhancement of SOC in rainfed, low input cropping systems is constrained by severe limiting (water and nutrient availability) conditions and thus very low SOC sequestration rates are normally observed. Therefore, reducing the SOC losses caused by water erosion and mineralization processes is of utmost importance. In order to 702 adapt semiarid rainfed systems and increase their resilience against climate change, several 703 sustainable agricultural management practices to control soil erosion and promote SOC 704 sequestration and water harvesting are being implemented in South-eastern Spain. Reducing 705 tillage, intercropping, green manuring, crop diversification, the selection of new and local 706 varieties better adapted to dry climates, crop residue retention, as well as the implementation 707 of vegetative buffer strips, application of swales and ponds for soil and water conservation, as 708 well as the selection of new and local varieties better adapted to dry climates, can be 709 promising options to make agro-ecosystems more resilient against climate change and market 710 price fluctuations in the long-term. However, the success of these sustainable agricultural 711 land-management practices will depend on the local conditions (soil, climate, and 712 management) together with the socio-economic context. In this regard, monitoring programs 713 and integrated assessments are needed to demonstrate the long-term beneficial effects of such 714 practices from farm hillslope to landscape regional scale and could be the target of OGs. 715 A good example of carbon-SOC management in semiarid rainfed low input systems under 716 eroded soils is that implemented in Southeastern Europe as an outcome of the DESIRE 717 European project (Ritsema and Stroosnidjer, 2008 [A9] [A10]), in which different sustainable 718 agricultural practices such as reduced tillage, green manuring during fallow periods, straw 719 mulch, and traditional water harvesting techniques (e.g., swales) were successfully 720 implemented to increase SOC and reduce soil and water loss through runoff in cereal fields 721 (de Vente et al., 2012). Specifically, passing from conventional moldboard ploughing at 430 cm depth (5-7 passes yr<sup>-1</sup>) to minimum tillage at 20 cm depth (2 passes yr<sup>-1</sup>) has reduced 722 723 carbon losses by soil erosion and runoff by 56% and increased the SOC stocks at 30 cm depth by 11% (15 and 16.7 Mg SOC ha<sup>-1</sup> in the former and in the latter, respectively) since 2010<del>08</del> 724 725 to 20161 (Martínez-Mena et al., 2020) (Table 2). [A11] Given the outcomes of these 726 experimental plots, these agricultural practices are being currently implemented in other

similar areas in SE Spain as part of a monitoring programme in collaboration with the local
farmer association Alvelal (<u>www.alvelal.net</u>) and the Commonland Foundation
(www.commonland.com).

730 Soil and crop management strategies intended to increase SOC in low input rainfed 731 systems need to be adapted to the local conditions, being aware of the high variability of 732 pedoclimatic conditions, which may constrain their potential to effectively increase SOC in a 733 specific year, so a long-term perspective is encouraged. For example, in areas more prone to 734 water erosion, because of erodible soils with significant slopes, the incorporation of plant 735 residues through minimum tillage together with the implementation of swales are 736 recommended (Figure 1-2). However, in flat areas with soils less prone to compaction, the 737 implementation of ponds together with no tillage can be a suitable option (Figure 3). Also, 738 demonstration farms in which different mixed cropping systems are tested before being 739 implemented at larger scales are mandatory. 740

Figure 1 – Keyline opened in an arable field to reduce water erosion in flat land in Spain.
Source: Maria Almagro

Figure 2 – Swale opened in a hilly land to reduce the speed of water runoff in Spain.

744 Source: Maria Almagro

Figure 3 – Pond realized in flat soil less prone to soil compaction in Spain. Source: Maria
Almagro

# 5.3 High Tech and Precision Farming to maintain SOC on farm and reduce GHG emissions

749Some research/demonstration projects combined PA and conservation agriculture (CA)

techniques by means of ICT in Northern Italy (Veneto Region), and in Southern Spain

751 (Andalusia). The project Agricare (Furlan, 2017) proposed a wheat/canola/maize/soybean 752 rotation at large field scale near Venice, comparing minimum tillage (MT), strip tillage (ST) 753 and no-till (NT) against conventional tillage (CT), under uniform and variable rate 754 application (VRA) of inputs. The project Agricarbon (González-Sánchez et al., 2012) 755 compared 3 farms in Andalusia region applying, on large plots, conventional soil 756 management against the combination of direct drill+PA (GNSS-assisted machinery, sensor-757 assisted maps, VRA for fertilizers and herbicides by prescription maps, etc.), on a 758 wheat/sunflower/broad bean rotation. 759 The project Agricare obtained the best results where NT+VRA was applied, with emissions savings of 0.5 t CO<sub>2</sub>eq ha<sup>-1</sup> and the same gross income than CT. SOC content was 760 761 assessed through time and type of management, to model its mid-term (15-yrs) dynamics, 762 resulting in relevant emission savings by CA+PA techniques (lower direct and indirect 763 energy consumption, reduced losses of SOC due to oxidation, higher fertilization efficiency). 764 The project Agricarbon showed specific-, for each crop, reductions of energy consumption 765 per product unit of 12, 26 and 18%, and production costs savings of 10, 21 and 15%, -

respectively for wheat, sunflower and broad bean, respectively. On a 4-yr average, CA+PA
 resulted in a 30% increase of SOC stock, as compared with conventional management (Table
 2).

The combination between PA, organic fertilizers application and soil conservation techniques (Pezzuolo et al., 2017), could achieve the best results in terms of both direct and indirect reduction of GHGs emissions, and eventually increase in-SOC storage. In this case, more information is needed about the mineralization rate of organic fertilisers fertilizers under different combinations of soil, climate, crop type and management to make VRA applications suitable also for farming systems based on organic fertilisation fertiliz ation as organic farming systems.

PA techniques have the potential to support the <u>decision makingdecision-making</u> process of also organic farmers dealing with spatial variability in their soil conditions, triggering finetuning and adaptation of crop technique to small-scale soil fertility level. This could be included in specific EIP's OGs at different sites.

Digitization of farming and conservation practices need to be adapted anyway at farm level to obtain tailor-made solutions, through the coordination of experts aimed at enhancing environmental and economic performance. Another important challenge is to make PA technologies accessible also to smallholders. The engagement of contractors managing large pieces of land and networking activities specifically aimed to connect small holdings in an information hub exploring also the issue of SOC will be key actions in that sense.

786

#### 787 **5.4 Using cover crops to increase SOC under limiting pedoclimatic conditions.**

788 Although widely recognized as effective tools to increase SOC stocks, the use of cover 789 crops and mulches may be limited in practice by several reasons, above-all limiting 790 pedoclimatic conditions and socio-economic constraints. For instance, in Mediterranean dry 791 areas of Europe, dry summers can hinder the adoption of strategies including double cropping 792 or summer cover crops. In sub-humid areas, where rainfall normally occurs also in 793 summertime, spring/summer cover crops can be profitably grown even in rainfed conditions, 794 instead, but normally they are not, because of farmers' attitudes and economic reasons (e.g., 795 farmers prefer to grow cash crops as maize or soybean instead of cover crops, determining 796 intensive soil exploitation). To enhance the adoption of cover crops among farmers, financial 797 instruments (e.g. specific subsidies included among the agri-environmental measures of the 798 RDPs of several Countries/Regions of Europe) can play an important role, but only at an 799 initial stage. Farmers should be rather convinced about the importance of cover crops in

sustaining the fertility of their soils and the yield of their most important cash crops in the
long run, in a context of market uncertainty and climatic fluctuations triggering farming
unprofitability. Adequate levels of financial support and effective education efforts should be
tailored to each specific pedoclimatic and socio-economic context through OGs involving all
the target stakeholders.

805 In this framework, peer-to-peer knowledge transfer among farmers can be crucial. Several 806 innovative strategies developed by local farmers with a wide knowledge of their soils and 807 climate conditions have resulted in a win-win strategy allowing for a continuous soil cover 808 and an optimization of the storage of soil C under rainfed conditions. Some particular 809 examples of these strategies are summarized in Table 24. They represent different case 810 studies developed in Italy and Spain under reduced tillage systems, aiming to maximize soil 811 cover and reduce as much as possible the period without living plants on the soil, without any 812 detrimental effect on farm profitability (Figures 4-5). Agronomic solutions included the use 813 of spontaneous weeds or cover crop mixtures inter-sown in double crops or after main crop 814 harvest and terminated mechanically or chemically immediately before the following cash 815 crop, ensuring a continuous soil cover all year around and diversifying the quality of the 816 residues returned to the soil. These experiences constitute good examples to be replicated at a 817 broader scale in the same or in other regions, with similar environmental conditions, possibly 818 involving other farmers and stakeholders to form OGs.

819

Figure 4 – Mechanical termination of a hairy vetch (*Vicia villosa* Roth.) cover crop by
roller crimper and simultaneous sod-seeding of sunflower (*Helianthus annuus* L.). Source:
Daniele Antichi

823

Figure 5 – Dead mulch provided by hairy vetch (*Vicia villosa* Roth.) cover crop terminated
by roller crimper and reducing weed pressure in sunflower (*Helianthus annuus* L.). Source:
Daniele Antichi

827

828 An observation arising from the examples in Table 24 is that adoption of diversified crop 829 rotations and inclusion of cover crops are essential tools in CA systems to achieve a 830 continuous soil cover and, consequently, an increase in SOC content. Anyway, the entire crop 831 rotation and all the related agronomic strategies must be designed with a holistic approach. 832 This needs to consider the specific climatic and soil conditions, as well as the economic 833 targets of the farm, to path the way for an agronomically and economically efficient 834 application of CA principles. Although with some extra efforts when starting to use them, 835 cover crops can be efficiently introduced in many kinds of crop rotations also in spring or 836 summer time, also in water-limiting conditions. In extreme cases, also keeping growing 837 spontaneous weed species might be a win-win strategy both from biomass production and 838 economic viability points of view. The key factor of success is a proper technical guidance 839 about best solutions for each specific local context in terms of choice of cover crop species 840 and establishment/termination technique and timing, made according to a specific and 841 realistic soil water balance. Finally, it is also recommended that OGs would envisage a 842 support of machinery builders, advisors and plant breeders in providing the best solutions for 843 each specific local context, targeting also also targeting a significant reduction in herbicide 844 and fertilizer use to benefit GHG mitigation and prevent on- and off-site soil and water 845 pollution. Selecting and tTesting and selection of improved genotypes of cover crops, adapted 846 to the local conditions and targeted to increase their potential to supply high C inputs to the 847 soil and to grow in limited water availability are also important steps further to increase the 848 adoption rate of cover crops among farmers.

Furthermore, increasing the awareness among farmers about the benefits of avoiding bare
soil during fallow periods would be a likely effective action. This could be pursued, for
example, by running an economic assessment to quantify the negative economic impacts of
losing a significant amount of soil, and associated carbon and nutrients, after extreme erosion
events when the soil is unprotected by vegetation.

854

Table 1 - Examples of successful case-studies on the use of cover crops under limiting
 pedoclimatic conditions for a possible development into EU Operational Groups.
 857

#### 5.4-<u>5</u> Management adaptation in areas subjected to water bombs and hail risk

859 Dry lands that experience extreme thunderstorms, even if infrequent, are the most 860 susceptible to the negative effects of both rainfall and hail in terms of soil compaction, 861 erosion and loss of the C-richest layer, i.e. the topsoil. In Europe, many areas are frequently 862 affected by such events. Among these, the Mediterranean region has the highest risk of 863 erosive events and flooding events, yet, at the same time, water scarcity — because of the 864 low infiltration of very intense local rainstorms- and loss of soil fertility. This is because of frequent thunderstorms associated with water-bombs and hailstorms that typically occur in 865 866 fall after long dry conditions in spring-summer. Farmers are more and more often 867 experiencing severe crop damages due to late summer-early fall thunderstorms, but also, their 868 soils are reported to be degraded.

In field vegetable cropping systems, soil structure could be better protected, and SOC maintained or increased, by the so called "permanent raised bed" technique, i.e. the combination of reduced tillage and permanent soil cover achieved through mulching (plastic films or organic material) (Sayre and Moreno, 1997) (Figure 6). In this technique, the soil is

873 tilled only once at the beginning of cultivation to establish high macro-porosity and then is 874 covered by plastic or biodegradable films, or even organic material as cereal straw, wood 875 chips, etc. This mulching material will stay on top of the soil until the end of its lifecycle, 876 which is normally about one year for the plastic mulch, or a single season for the 877 biodegradable film and the organic materials. Organic fertilizers are usually incorporated into 878 the soil at the time of the initial tillage. Each crop is manually transplanted into the mulch and 879 the raised bed is never trampled by field workers/machines, in order to protect soil structure. 880 Once the first crop has been harvested manually, also its residues are removed from the fields 881 and used for composting. Then, the second crop can be transplanted in the same positions or 882 in new ones, without any replacement of the mulch. As long as If the mulch material is 883 sufficiently covering the soil, it could be kept on place and the next crop transplanted.

884

Figure 6 – Permanent seedbed implemented in Veneto, North-East of Italy. Source:
Daniele Antichi

887

888 A group of farmers practicing this technique in Veneto (NE Italy) reports that with this 889 management they could improve their soils in terms of organic C content and biological and 890 physical fertility (Luca Conte, personal communication). As a proof of that, the spade test, 891 usually performed to evaluate visually soil fertility, always gave excellent results (Figure 7). 892 Soil structure was dramatically improved, soil depth reached at least 30 cm, earthworms were 893 abundant and organic materials were well decomposed. This promising technique can be 894 applied not only in small farms but also in larger ones and practiced with business as usual 895 machinery (e.g. standard transplanting machines) and the use of organic material instead of 896 films. For large vegetable farms willing to use films instead of organic mulch material, the 897 use of PA technologies for detection of transplant patterns can make it possible to perform

the transplant of vegetables into permanent raised beds also mechanically, with hugeadvantages in terms of costs saving.

900

901 Figure 7 – Soil structure improved after one year of implementation of permanent seedbed
902 on plastic film in Veneto, North-East of Italy. Source: Daniele Antichi

903

Although the use of organic mulch materials (e.g. biodegradable films, wood chips, straw, 905 <u>etc.</u>...) should be preferred to plastic mulch to reduce the environmental impact of non-906 renewable materials, it is not clear yet whether this could be more profitable for farmers from 907 an economic point of view, due to the short lifetime of organic materials. This aspect needs 908 further investigation both at scientific and demonstration levels.

909 Another technical issue to be carefully considered is the management of the space between 910 the raised beds. This space is much prone to soil compaction (and then to water logging) due 911 to the huge traffic by field workers and/or machines. Some benefits in terms of water 912 infiltration, but also of weed suppression, may come from sowing perennial living mulch 913 (e.g. white clover, black medic) on this space when establishing the raised beds. Clearly, this 914 might imply the development of proper machines adapted to these conditions for sowing and 915 management of the living mulch, that is currently not included in research projects. 916 Following this development, OGs could be focused in developing adequate strategies for the 917 implementation of the technique described above in different farm typologies and 918 pedoclimatic conditions. 919 920 6. Conclusions and perspectives

We have identified a series of lLocal climatic and soil limitations conditions asean be
 limiting <u>factors</u> for the adoption of some strategies to increase SOC in European arable lands-

923 They are: i) water scarcity or seasonal imbalances, which need the tailoring of strategies 924 under irrigation, ii) high risk of wind or water soil erosion, reducing SOC in the surface soil 925 layers, iii) presence of a shallow groundwater and limited drainage, which pose specific 926 management problems and lower crop productivity, iv) high stoniness, limiting the use of 927 certain machinery, v) saline, gypsiferous and alkaline soils, where SOC is naturally more 928 difficult to be stabilized, and vi) manure fertilization, increasing GHG emissions in 929 Mediterranean climate and fine textured soils. 930 To overcome some of this limiting factors we have illustrated Some five case studies 931 illustrate together with potential proposalssible references for Operational Groups to support 932 a successful local adaptation of measures agronomic practices to improve carbon-SOC 933 storage in arable lands. They include: i) 934 -Adoption of soil management strategies to improve SOC storage in irrigated systems, 935 <u>ii) -</u>

936

937 <u>—</u>Management of rainfed and low input crops to maintain and increase SOC in 938 dry climates and erosive prone soils, iii)

Precision farming and other high-tech solutions able to generate local diagnosis and
 adaptive strategies for increasing SOC and reducing GHG emissions, iv) -

941 —Innovative strategies for extending soil cover periods and introducing cover crops in

p42 rotations in areas with limited water availability or prone to harsh weather conditions, and v) -

943 —Adaptation of soil management to cope with water bombs and hail risk.

944 Other-Additional possible OGs could should deal with no-till based cropping systems of

heavy soils, in order to find affordable ways to improve <u>the soil</u> structure in <u>of surface soil</u>

946 surface <u>horizons</u> soil (i.e. the first 10 cm-layer, the most prone to compaction in the transition

947 to no-till), which is crucial for a good early establishment of the crops and for improving

948 water infiltration and drainage. Furthermore, effective combinations of no-till/reduced tillage 949 with other components of cropping systems -(e.g. crop rotation, fertilization, irrigation 950 strategy and technique, cover cropping, mechanical weed control) have to be identified, 951 aiming at maximizing the return to the soil of large amounts of C, on the one hand, and at 952 modulating soil organic matter mineralization rates, in order to synchronize them with crop 953 nutrient demands, on the other hand. 954 In general, increasing no-till crop productivity in areas with limiting pedoclimatic 955 characteristics conditions requests an additional knowledge effort on the climate-soil-crop 956 interactions, and permitting which can lead to could be achieved through the tuning of

957 intensive application of specific agro-ecological strategies, like <u>in the case study of the</u>

958 permanent raised bed technique applied in small holding vegetable production.

959

#### 960 Acknowledgements

961 The work was initiated in the framework of the EIP-AGRI Focus Group on Moving from

962 source to sink in arable farming (<u>https://ec.europa.eu/eip/agriculture/en/focus-groups/moving-</u>

- 963 <u>source-sink-arable-farming</u>).
- 964 One author was supported by the Juan de la Cierva Program (Grant IJCI-2015-23500).

#### 965 **References**

966 Aguilera, E., Guzman, G., and Alonso, A. (2015). Greenhouse gas emissions from

967 conventional and organic cropping systems in Spain. II. Fruit tree orchards. Agronomy for

- 968 Sustainable Development, 35 (2), 725–737. https://doi.org/10.1007/s13593-014-0265-y.
- 969 Aguilera, E., Lassaletta, L., Sanz-Cobena, A., Garnier, J., and Vallejo, A. (2013a).
- 970 The potential of organic fertilizers and water management to reduce N<sub>2</sub>O emissions in

- 971 Mediterranean climate cropping systems. A review. Agriculture, Ecosystems & Environment,
  972 164, 32–52. http://dx.doi.org/10.1016/j.agee.2012.09.006.
- Aguilera, E., Lassaletta, L., Gattinger, A., Gimeno, B.S. (2013b). Managing soil
  carbon for climate change mitigation and adaptation in Mediterranean cropping systems: A
  meta-analysis. Agriculture, Ecosystems & Environment, 168, 25–36.
- 976 https://doi.org/10.1016/j.agee.2013.02.003
- Álvaro-Fuentes, J., Paustian, K. (2011). Potential soil carbon sequestration in a
- 978 semiarid Mediterranean agroecosystem under climate change: Quantifying management and
- 979 climate effects. Plant and Soil, 338, 261–272. <u>https://doi.org/10.1007/s11104-010-0304-7</u>.
- 980 Antón, R., Virto, I., Gonzalez, J., Hernandez, I., Enrique, A., Bescansa, P., Arias, N.,
- 981 Orcaray, L., Campillo, R. (2019). Extension of irrigation in semi-arid regions: What
- 982 challenges for soil security? Perspectives from a regional-scale project in Navarre (Spain), in:
- 983 Anne Richer de Forges, Florence Carré, Alex B. McBratney, Johan Bouma, D.A. (Ed.),
- 984 Global Soil Security. Towards More Science-Society Interfaces. pp. 79–87.
- 985 Antonopoulos, I. S., Canfora, P., Dri, M., Gaudillat, P., Styles, D., Williamson, J.,
- Jewer, A., Haddaway, N., Price, M. (2017). European Commission. Best environmental
- 987 management practice for the agriculture sector crop and animal production. Final draft.
- 988 <u>http://susproc.jrc.ec.europa.eu/activities/emas/documents/AgricultureBEMP.pdf</u>
- 989 Apesteguía, M., Virto, I., Orcaray, L., Enrique, A, Bescansa, P. (2015). Effects of the
- 990 conversion to irrigation of semiarid Mediterranean dryland agroecosystems on soil carbon
- dynamics and soil aggregation. Arid Land Resource Management, 2015, 29, 339–414.
- 992 https://doi.org/10.1080/15324982.2015.1016245.
- 993 Apesteguía, M., Virto, I., Orcaray, L., Bescansa, P., Enrique, A., Imaz, M.J., Karlen,
- 994 D.L. (2017). Tillage Effects on Soil Quality after Three Years of Irrigation in Northern Spain.
- 995 Sustainability, 9(8), 1476. https://doi.org/10.3390/su9081476.

- 996 Arias, N., Orcaray, L., Bescansa, P., Enrique, A., & Virto, I. (2017). Implications of
- 997 Rock Fragments for Soil Quality Evaluation: Assessing Changes in a Gravelly Irrigated Soil
- 998 Following No Till Adoption. Communications in Soil Science and Plant Analysis, 48:22,
- 999 2663-2677. https://doi.org/10.1080/00103624.2017.1416140.
- 1000 Berglund, K., Berglund, Ö. (Eds.) (2017). Proceedings of the International Conference
- 1001 on Climate Smart Agriculture on Organic Soils. Uppsala, Sweden, 23-24 November 2017.
- 1002 https://pub.epsilon.slu.se/14739/1/berglund\_et\_al\_171121.pdf
- 1003 Berglund Ö., Berglund K. (2010). Distribution and cultivation intensity of agricultural
- 1004 peat and gyttja soils in Sweden and estimation of greenhouse gas emissions from cultivated
- 1005 peat soils, Geoderma, Volume 154 (3-4), 173-180.
- 1006 <u>https://doi.org/10.1016/j.geoderma.2008.11.035</u>.
- Bescansa, P., Imaz, M.J., Virto, I., Enrique, A., Hoogmoed, W. (2006). Soil water
  retention as affected by tillage and residue management in semiarid Spain. Soil and Tillage
  Research, 87, 19-27.
- 1010 Borrelli, P., Lugato, E., Montanarella, L., Panagos, P. (2018). A new assessment of
- 1011 soil loss due to wind erosion in European agricultural soils using a quantitative spatially
- 1012 distributed modelling approach. Land Degradation and Development, 28, 335-344.
- 1013 https://doi.org/10.1002/ldr.2588.
- 1014 Borrelli, P., Panagos, P., Ballabio, C., Lugato, E., Weynants, M., Montanarella, L.
- 1015 (2016). Towards a Pan-European assessment of land susceptibility to wind erosion. Land
- 1016 Degradation and Development, 27: 1093–1105. https://doi.org/10.1002/ldr.2318.
- 1017 Cerdan, O., Govers, G., Le Bissonnais, Y., Van Oost, K., Poesen, J., Saby, N., Gobin,
- 1018 A., Dostal, T. (2010). Rates and spatial variations of soil erosion in Europe, a study based on
- 1019 erosion plot data. Geomorphology 122, 167–177.

1020	Chenu, C., Angers, D. A., Barré, P., Derrien, D., Arrouays, D., & Balesdent, J. (2019).					
1021	Increasing organic stocks in agricultural soils: Knowledge gaps and potential innovations.					
1022	Soil and Tillage Research, 188, 41-52.					
1023	Chinseu, E., Dougill, A., & Stringer, L. (2019). Why do smallholder farmers dis-adopt					
1024	conservation agriculture? Insights from Malawi. Land Degradation and Development, 30(5),					
1025	533–543.					
1026	Costantini, E. A., Castaldini, M., Diago, M. P., Giffard, B., Lagomarsino, A.,					
1027	Schroers, H. J., Priori, S., Valboa, G., Agnelli, A. E., Akça, E., D'Avino, L., Fulchin, E.,					
1028	Gagnarli, E., Kiraz, M. E., Knapič, M., Pelengić, R., Pellegrini, S., Perria, R., Puccioni, S.,					
1029	Simoni, S., Tangolar, S., Tardaguila, J., Vignozzi, N., Zombardo, A. (2018). Effects of soil					
1030	erosion on agro-ecosystem services and soil functions: A multidisciplinary study in nineteen					
1031	organically farmed European and Turkish vineyards. Journal of environmental management,					
1032	223, 614-624.					
1033	Costantini E.A.C., Dazzi C. (Eds.) (2013). The Soils of Italy. World Soils Book					
1034	Series, Springer, 354.					
1035	Costantini, E.A.C. & Lorenzetti, R. (2013). Soil degradation processes in the Italian					
1036	agricultural and forest ecosystems. Italian Journal of Agronomy, 8(4), 28.					
1037	https://doi.org/10.4081/ija.2013.e28.					
1038	de Vente, J., Solé-Benet, A., López, J., Boix-Fayos, C. (2012). Biophysical and					
1039	socioeconomic impacts of soil and water conservation measures. An evaluation of					
1040	Sustainable Land Management in SE Spain. Geophysical Research Abstracts Vol. 14,					
1041	EGU2012-12497, EGU General Assembly 2012.					
1042	Denef, K., Stewart, C.E., Brenner, J., Paustian, K. (2008). Does long-term center-					
1043	pivot irrigation increase soil carbon stocks in semiarid agro-ecosystems? Geoderma, 145,					
1044	121-129. https://doi.org/10.1016/j.geoderma.2008.03.002.					

1045	Dignac, M.F., Derrien D., Barré, P., Barot, S., Cécillon, L., Chenu, C., Chevallier, T.,						
1046	Treschet, G.T., Garnier, P., Guenet, B., Hedde, M., Klumpp, K., Lashermes, G., Maron, P-A.,						
1047	Nunan, N., Roumet, C., Basile-Doelsch, I. (2017). Increasing soil carbon storage:						
1048	mechanisms, effects of agricultural practices and proxies. A review. Agronomy for						
1049	Sustainable Development, 37: 14.						
1050	Dimassi, B., Mary, B., Wylleman, R., Labreuche, L., Coutoure, D., Piraux, F., Cohan,						
1051	JP. (2014). Long-term effect of contrasted tillage and crop management on soil carbon						
1052	dynamics during 41 years. Agriculture, Ecosystems and Environment, 188, 134–146.						
1053	https://doi.org/10.1016/j.agee.2014.02.014.						
1054	EIP-AGRI, Agriculture and Innovation (2019). Operational Group idea: Best practices						
1055	and advisory service on irrigation with C sequestration. Available at						
1056	https://ec.europa.eu/eip/agriculture/en/find-connect/project-ideas/operational-group-idea-						
1057	best-practices-and-advisory [Accessed September 6, 2019].						
1058	Entry, J.A., Sojka, R., Shewmaker, G.E. (2002). Management of irrigated agriculture						
1059	to increase organic carbon storage in soils. Soil Science Society of America Journal, 66,						
1060	<u>1957–1964. https://doi.org/10.2136/sssaj2002.1957.</u>						
1061							
1062	Eurostat (2018). Statistics explained. Soil, land cover and land use.						
1063	https://ec.europa.eu/eurostat/statistics-						
1064	explained/index.php?title=Category:Soil,_land_cover_and_land_use (accessed on 18						
1065	November 2018).						
1066	European Environmental Agency (EEA) (2018). Glossary.						
1067	https://www.eea.europa.eu/help/glossary/eea-glossary/agroecosystem (accessed on 1						
1068	December 2018).						

- 1069 Entry, J.A., Sojka, R., Shewmaker, G.E. (2002). Management of irrigated agriculture
- 1070 to increase organic carbon storage in soils. Soil Science Society of America Journal, 66,
- 1071 <del>1957–1964. https://doi.org/10.2136/sssaj2002.1957.</del>
- 1072 European Commission (2017). COM(2017) 713 final The Future of Food and
- 1073 Farming, https://ec.europa.eu/agriculture/sites/agriculture/files/future-of-
- 1074 <u>cap/future\_of\_food\_and\_farming\_communication\_en.pdf</u> (accessed on 1 December 2018).
- 1075 FAO (2012). Peatlands guidance for climate change mitigation by conservation,
- 1076 rehabilitation and sustainable use (H. Joosten, M. L. Tapio-Biström, & S. Tol, Eds.).
- 1077 Retrieved from http://www.fao.org/docrep/015/an762e/an762e00.htm
- 1078 FAO (2017a). Voluntary Guidelines for Sustainable Soil Management.
- 1079 http://www.fao.org/global-soil-partnership/resources/highlights/detail/en/c/472458/
- 1080 FAO (2017b). Global Soil Organic Carbon Map (GSOCmap).
- 1081 http://www.fao.org/global-soil-partnership/pillars-action/4-information-and-data-new/global-
- 1082 soil-organic-carbon-gsoc-map/en/
- 1083 Finke, P., Hartwich, R., Dudal, R., Ibánez, J., Jamagne, M., King, D., Montanarella, L.
- 1084 & Yassoglou, N. (1998). Georeferenced soil database for Europe; manual of procedures
- 1085 <u>Version 1.0.</u>
- 1086 http://eusoils.jrc.ec.europa.eu/esdb\_archive/eusoils\_docs/esb\_rr/n05\_ManualVer11EN4.pdf.
- 1087 Follett, R.F., Jantalia, C.P., Halvorson, A.D. (2013). Soil carbon dynamics for
- 1088 irrigated corn under two tillage systems. Soil Science Society of America Journal, 77, 951-
- 1089 963. https://doi.org/10.2136/sssaj2012.0413.
- 1090 Francaviglia, R., Di Bene, C., Farina, R., Salvati, L. & Vicente-Vicente, J.L. (2019).
- 1091 Assessing "4 per 1000" soil organic carbon storage rates under Mediterranean climate: a
- 1092 comprehensive data analysis. Mitigation and Adaptation Strategies for Global Change, 1-24.
- 1093 https://doi.org/10.1007/s11027-018-9832-x.

1094	Furlan, L. (2017). LIFE+ AGRICARE: Introducing innovative precision farming					
1095	techniques in AGRIculture to decrease CARbon Emissions: technical document. Veneto					
1096	Agricoltura – Agency of the Veneto Region for the innovation in the Primary Sector.					
1097	Füssel, H-M., Jol, A., Marx, A. et al. (2017). Climate change, impacts and					
1098	vulnerability in Europe 2016 - An indicator-based report (Hans-Martin Füssel, André Jol,					
1099	Andreas Marx, Mikael Hildén Eds.), Vol. 1/2017 (January 2017).					
1100	García-Marco, S., Abalos, D., Espejo, R., Vallejo, A., Mariscal-Sancho, I., 2016. No					
1101	tillage and liming reduce greenhouse gas emissions from poorly drained agricultural soils in					
1102	Mediterranean regions. Science of Total Environment, 566–567, 512–520					
1103	González-Sánchez, E.J., Ordóñez-Fernández, R., Carbonell-Bojollo, R., Veroz-					
1104	González, O., Gil-Ribes, J.A. (2012). Meta-analysis on atmospheric carbon capture in Spain					
1105	through the use of conservation agriculture. Soil and Tillage Research. 122, 52-60.					
1106	https://doi.org/10.1016/j.still.2012.03.001.					
1107	González-Sánchez, E.J., Veroz-González, O., Gil-Ribes, J.A., Ordóñez-Fernández, R.					
1108	(2018). Iniciativa 4 por mil: el carbono orgánico del suelo como herramienta de mitigación y					
1109	adaptación al cambio climático en España. Ed. Oficina Española de Cambio Climático.					
1110	Ministerio de Agricultura y Pesca, Alimentación y Medio Ambiente. Madrid, Spain, 262.					
1111	Henry, B., Murphy, B. and Cowie, A. (2018). Sustainable Land Management for					
1112	Environmental Benefits and Food Security. A synthesis report for the GEF.					
1113	http://www.stapgef.org/sustainable-land-management-environmental-benefits-and-food-					
1114	security-synthesis-report-gef					
1115	Herrero, J. (2017). On the early irrigation of gypseous lands in Spain. Land					

1116 Degradation and Development, 28, 1152–1155. <u>https://doi.org/10.1002/ldr.2683</u>.

1117	Ingram, J., Mills, J., Frelih-Larsen, A., Davis, M., Merante, P., Ringrose, S., Molnar,
1118	A., Sánchez, B., Ghaley, B.B. and Karaczun, Z. (2014). Managing soil organic carbon: a farm
1119	perspective. EuroChoices, 13(2), 12-19.
1120	JRC (Joint Research Centre), 2011. The State of Soil in Europe. Luxembourg:
1121	Publication Office of the European Union. pp 78.
1122	Kassam A., FriedrichT., Derpsch R. and KienzleJ. (2015). Overview of the world
1123	widespread of conservation agriculture. Field Actions Science Reports [Online], Vol.8/2015,
1124	verified on 20 April 2020.
1125	Krichels, A., DeLucia, E.H., Sanford, R., Chee-Sanford, J. and Yang, W. H. (2019).
1126	Historical soil drainage mediates the response of soil greenhouse gas emissions to intense
1127	precipitation events. Biogeochemistry 142, 425-442. https://doi.org/10.1007/s10533-019-
1128	<u>00544-x.</u>
1129	
1130	Kumar, S., Nakajima, T., Kadono, A., Lal, R., Fausey, N., 2014. Long-term tillage
1131	and drainage influences on greenhouse gas fluxes from a poorly drained soil of central Ohio.
1132	Journal of Soil Water Conservation, 69, 553-563. https://doi.org/10.2489/jswc.69.6.553
1133	Lal, R. (2008). Soil carbon stocks under present and future climate with specific
1134	reference to European ecoregions. Nutrient Cycling in Agroecosystems, 81, 113-127.
1135	Le Bissonnais, Y., Montier, C., Jamagne, M., Daroussin, J., King, D. (2002). Mapping
1136	erosion risk for cultivated soil in France. Catena 46, 207–220.
1137	Liu, X., Han, X., Song, C., Herbert, S. J. and Xing, B. (2003). Soil organic carbon
1138	dynamics in black soils of china under different agricultural management systems.
1139	Communications in Soil Science and Plant Analysis, 34, 973-984.
1140	Lugato, E., Bampa, F., Panagos, P., Montanarella, L., Jones, A. (2014a). Potential
1141	carbon sequestration of European arable soils estimated by modelling a comprehensive set of

1 142 management practices. Global Change Biology, 20: 3557–3567.

1 143 https://doi.org/10.1111/gcb.12551.

1144	Lugato, E., Panagos, P., Bampa, F., Jones, A., Montanarella, L. (2014b). A new
1145	baseline of organic carbon stock in European agricultural soils using a modelling approach.
1146	Global Change Biology, 20, 313–326. https://doi.org/10.1111/gcb.12292.
1147	Maetens, W., Poesen, J., Vanmaercke, M. (2012). How effective are soil conservation
1148	techniques in reducing plot runoff and soil loss in Europe and the Mediterranean? Earth-
1149	Science Reviews, 115, 21-36. https://doi.org/10.1016/j.earscirev.2012.08.003.
1150	Manono, B.O., Moller, H. (2015). Effects of stock type, irrigation and effluent
1151	dispersal on earthworm species composition, densities and biomasses in New Zealand
1152	pastures. Pedobiologia, 58, 187–193. https://doi.org/10.1016/j.pedobi.2015.09.002.
1153	Martínez-Mena, M., Carrillo-López, E., Boix-Fayos, C., Almagro, M., García Franco,
1154	N., Díaz-Pereira, E., Montoya, I. & de Vente, J. 2020. Long-term effectiveness of sustainable
1155	land management practices to control runoff, soil erosion, and nutrient loss and the role of
1156	rainfall intensity in Mediterranean rainfed agroecosystems. Catena, 187.
1157	https://doi.org/10.1016/j.catena.2019.104352
1158	
1159	Mazzoncini, M., Sapkota, T.B., Bàrberi, P., Antichi, D., & Risaliti, R. (2011). Long-
1160	term effect of tillage, nitrogen fertilization and cover crops on soil organic carbon and total
1161	nitrogen content. Soil and Tillage Research, 114(2), 165-174. https://doi.org/
1162	10.1016/j.still.2011.05.001.
1163	McDowell, R.W., van der Weerden, T.J., Campbell, J. (2011). Nutrient losses
1164	associated with irrigation, intensification and management of land use: A study of large scale
1165	irrigation in North Otago, New Zealand. Agriculture Water Management, 98, 877-885.
1166	https://doi.org/10.1016/j.agwat.2010.12.014.

- 1167 Moret-Fernández, D., Herrero, J. (2015). Effect of gypsum content on soil water
- 1168 retention. Technical Note. Journal of Hydrology 528, 122–126.

1169 Muñoz-Rojas, M., Abd-Elmabod, S.K., Zavala, L.M., de la Rosa, D., Jordán, A.

1170 (2017). Climate change impacts on soil organic carbon stocks of Mediterranean agricultural

1 areas: A case study in Northern Egypt. Agriculture, Ecosystems and Environment, 238, 142–

1 172 152. https://doi.org/10.1016/j.agee.2016.09.001.

1/173 Muñoz-Rojas, M., Jordán, A., Zavala, L.M., de la Rosa, D., Abd-Elmabod, S.K.,

1 Anaya-Romero, M. (2012). Organic carbon stocks in Mediterranean types under different

1 and uses (Southern Spain). Solid Earth 3, 375-386. <u>https://doi.org/10.5194/se-3-375-2012</u>.

1/176 Nunes, J.M., López-Piñeiro, A., Albarrán, A., Muñoz, A., Coelho, J. (2007). Changes

1177 in selected soil properties caused by 30 years of continuous irrigation under Mediterranean

1178 conditions. Geoderma, 139, 321–328. <u>https://doi.org/10.1016/j.geoderma.2007.02.010</u>.

1179 Oliveira, M., Barré, P., Trinidade, H., Virto, I. (2019). Different efficiencies of grain

1180 legumes in crop rotations to improve soil aggregation and organic carbon in the short-term in

a sandy Cambisol. Soil and Tillage Research, 186, 23-25.Panagos, P., Borrelli, P.,

1182 Meusburger, C., Alewell, C., Lugato, E., Montanarella, L. (2015). Estimating the soil erosion

1183 cover-management factor at European scale. Land Use Policy, 48C, 38-50.

1184 <u>https://doi.org/10.1016/j.landusepol.2015.05.021</u>.

1185 Panagos, P., Borrelli, P., Poesen, J., Ballabio, C., Lugato, E., Meusburger, K.,

Alewell, C. (2015). The new assessment of soil loss by water erosion in Europe.

1187 Environmental science & policy, 54, 438-447.

1/188 Pareja-Sánchez, E., Plaza-Bonilla, D., Álvaro-Fuentes, J., Cantero-Martínez, C.

1 [189 (2019). Is it feasible to reduce tillage and N use while improving maize yield in irrigated

1190 Mediterranean agroecosystems? European Journal of Agronomy, 109, 125919.

- 1191 Pellegrini, S., Agnelli, A.E., Andrenelli, M. C., Barbetti, R., Papa, G.L., Priori, S., &
- 1192 Costantini, E.A.C. (2018). Using present and past climosequences to estimate soil organic
- 1193 carbon and related physical quality indicators under future climatic conditions. Agriculture,
- 1194 Ecosystems & Environment, 266, 17-30. https://doi.org/10.1016/j.agee.2018.07.015.
- 1195 Pezzuolo, A., Dumont, B., Sartori, L., Marinello, F., De Antoni Migliorati, M., Basso
- 1196 B. (2017). Evaluating the impact of soil conservation measures on soil organic carbon at the
- 1197 farm scale. Computers & Electronics in Agriculture, 135, 175–182.
- 1198 https://doi.org/10.1016/j.compag.2017.02.004.
- 1199 Regina, K., Budiman, A., Greve, M.H., Grønlund, A., Kasimir, Å, Lehtonen, H.,
- 1200 Petersen, S.O., Smith, P. & Wösten, H. (2016). GHG mitigation of agricultural peatlands
- 1201 requires coherent policies. Climate Policy, 16(4), 522-541.
- 1202 https://doi.org/10.1080/14693062.2015.1022854.
- 1203 Riksen, M.J.P.M., De Graaff, J. (2001). On-site and off-site effects of wind erosion on
- 1204 European light soils. Land Degradation & Development 12, 1–11.
- 1205 <u>https://doi.org/10.1002/ldr.423</u>.
- 1206 Ritsema, C. J., Stroosnijder, L. (2008). DESIRE: desertification mitigation and
- 1207 remediation of land: a global approach for local solutions. In D. Gabriels, & WM Cornelis
- 1208 (Eds.), Book of Abstracts of Conference on Desertification (pp. 66). Ghent, Belgium.
- 1209 Rodríguez-Murillo, J.C. (2001). Organic carbon content under different types of land
- 1210 use and soil in peninsular Spain. Biology and Fertility of Soils, 33(1), pp. 53-61.
- 1211 https://doi.org/ 10.1007/s003740000289.
- 1212 Rumpel, C., Amiraslani, F., Chenu, C., Garcia Cardenas, M., Kaonga, M., Koutika,
- 1213 L.-S., Ladha, J., Madari, B., Shirato, Y., Smith, P., Soudi, B., Soussana, J.-F., Whitehead,
- 1214 D., Wollenberg, E. (2019). The 4p1000 initiative: Opportunities, limitations and challenges

- 1215 for implementing soil organic carbon sequestration as a sustainable development strategy.
- 1216 Ambio, 1-11. https://doi.org/10.1007/s13280-019-01165-2.

1217 Sanz, M.J., de Vente, J., Chotte, J.-L., Bernoux, M., Kust, G., Ruiz, I., Almagro, M.,

- 1218 Alloza, J.-A., Vallejo, R., Castillo, V., Hebel, A. and Akhtar-Schuster, M. (2017). Sustainable
- 1219 Land Management contribution to successful land-based climate change adaptation and
- 1220 mitigation. A Report of the Science-Policy Interface. United Nations Convention to Combat
- 1221 Desertification (UNCCD), Bonn, Germany.
- 1222 Sasal, M.C., Léonard, J., Andriulo, A., Boizard, H. (2017). A contribution to
- 1223 understanding the origin of platy structure in silty soils under no tillage. Soil and Tillage
- 1224 Research, 173, 42-48.
- Sayre, K.D., Moreno, R.O. (1997). Application of raised-bed planting system towheat. Wheat Special Report no. 31. Mexico, DF:CIMMYT. 362.
- 1227 Schoonhoven, Y. & Runhaar, H. (2018). Conditions for the adoption of agro-
- 1228 ecological farming practices: a holistic framework illustrated with the case of almond farming
- in Andalusia, International Journal of Agricultural Sustainability, 16:6, 442-454,
- 1230 https://doi.org/ 10.1080/14735903.2018.1537664.
- 1231 Schwilch, G., Laouina, A., Chaker, M., Machouri, N., Sfa, M., & Stroosnijder, L.
- 1232 (2015). Challenging conservation agriculture on marginal slopes in Sehoul, Morocco.
- 1233 Renewable Agriculture and Food Systems, 30(3), 233-251. doi:10.1017/S1742170513000446
- 1234 Smith, P., Davies, C.A., Ogle, S., Zanchi, G., Bellarby, J., Bird, N., Boddey, R.M.,
- 1235 McNamara, N.P., Powlson, D., Cowie, A., van Noordwijk, M., Davis, S.C., Richter, D.D.B.,
- 1236 Kryzanowski, L., van Wijk, M.T., Stuart, J., Kirton, A., Eggar, D., Newton-Cross, G., Adhya,
- 1237 T.K., Braimoh, A.K. (2012). Towards an integrated global framework to assess the impacts
- 1238 of land use and management change on soil carbon: Current capability and future vision.
- 1239 Global Change Biology, 18: 2089–2101. https://doi.org/10.1111/j.1365-2486.2012.02689.x.

1240	Soane, B.D., Ball, B.C., Arvidsson, J., Basch, G., Moreno, F., Roger-Estrade, J.						
1241	(2012). No-till in northern, western and South-Western Europe: A review of problems and						
1242	opportunities for crop production and the environment. Soil and Tillage Research, 118, 66-						
1243	87. <u>https://doi.org/10.1016/j.still.2011.10.015</u> .						
1244	Stavi, I., Bel, G., & Zaady, E. (2016). Soil functions and ecosystem services in						
1245	conventional, conservation, and integrated agricultural systems. A review. Agronomy for						
1246	sustainable development, 36(2), 32.						
1247	Throop, H.L., Archer, S.R., Monger, H.C. and Waltman, S. (2012). When bulk						
1248	density methods matter: Implications for estimating soil organic c arbon pools in rocky soils.						
1249	Journal of Arid Environments, 77, 66–71. https://doi.org/10.1016/j.jaridenv.2011.08.020.						
1250	Verheye, W., & Boyagdiev, T. (1997). Evaluating the land use potential of						
1251	gypsiferous soils from field pedogenic characteristics. Soil Use and Management, 13(2), 97						
1252	103. https://doi.org/10.1111/j.1475-2743.1997.tb00565.x.						
1253	Virto, I., Barré, P., Burlot, A., Chenu, C. (2012). Carbon input differences as the main						
1254	factor explaining the variability in soil organic C storage in no-tilled compared to inversion						
1255	tilled agrosystems. Biogeochemistry, 108, 17-26. https://doi.org/10.1007/s10533-011-9600-						
1256	4.						
1257	Virto, I., Bescansa, P., Imaz, M.J., Enrique, A. (2006). Soil quality under food-						
1258	processing wastewater irrigation in semi-arid land, northern Spain: aggregation and organic						
1259	matter fractions. Journal of Soil and Water Conservation, 61, 398-407.						
1260	Virto, I., Imaz, M.J., Fernández-Ugalde, O., Gartzia-Bengoetxea, N., Enrique, A.,						
1261	Bescansa, P. (2015). Soil Degradation and Soil Quality in Western Europe: Current Situation						
1262	and Future Perspectives. Sustainability, 7, 313-365. https://doi.org/10.3390/su7010313.						

1263	Table 1 - Results from scientific literature and projects dealing with management practices
1264	and soil organic carbon (SOC) concentration (g 100 g <sup>-1</sup> ), stock (Mg C ha <sup>-1</sup> ) and storage rate
1265	(Mg C ha <sup>-1</sup> y <sup>-1</sup> )Examples of successful case-studies on the use of cover crops under limiting
1266	pedoclimatic conditions for a possible development into EU Operational Groups.
1267	Table 1 - Case-studies for a possible development into EU Operational GroupsTable 2 -
1268	Management practices and SOC: state of the art, identified good practices, research gaps,
1269	ideas for EIP-AGRI Operational Groups (OGs)
1270	
1271	
1272	Figure 1 – Keyline opened in an arable field to reduce water erosion in flat land in Spain.
1273	Source: Maria Almagro
1274	
1275	Figure 2 – Swale opened in a hilly land to reduce the speed of water runoff in Spain.
1276	Source: Maria Almagro
1277	
1278	Figure 3 – Pond realized in flat soil less prone to soil compaction in Spain. Source: Maria
1279	Almagro
1280	
1281	Figure 4 – Mechanical termination of a hairy vetch (Vicia villosa Roth.) cover crop by
1282	roller crimper and simultaneous sod-seeding of sunflower (Helianthus annuus L.). Source:
1283	Daniele Antichi
1284	
1285	Figure 5 – Dead mulch provided by hairy vetch (Vicia villosa Roth.) cover crop terminated
1286	by roller crimper and reducing weed pressure in sunflower (Helianthus annuus L.). Source:
1287	Daniele Antichi

1289	Figure 6 – Pe	rmanent seedl	ed imple	mented in V	leneto North	-East of Italy	Source:
1207	$r_{1}guic 0 - 1c$	manent secu	in mpici		r cheto, north	-Last Of Italy	. Source.

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- 1292 Figure 7 Soil structure improved after one year of implementation of permanent seedbed
- 1293 on plastic film in Veneto, North-East of Italy. Source: Daniele Antichi