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Suitability of discrete choice experiments for landscape

management under the European Landscape Convention

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ABSTRACT:

Landscapes throughout the world are increasingly being altered as a result of human actions and natural processes, therefore necessitating urgent management. Acknowledging this situation, the European Landscape Convention (ELC) was approved in 2000 with the explicit objective of protecting, managing and planning European landscapes. In this paper, we provide empirical evidence on the suitability of the Discrete Choice Experiment (DCE) methodology for valuing multidimensional landscapes under the ELC. An application in the Basque Country, Spain, serves as an example to illustrate that the DCE methodology may be suitable for supporting the aims of the ELC because: (i) it is a tool for public consultation; (ii) it offers an insight into the relative attractiveness of key landscape attributes, such as native forests and farming activities; and (iii) it provides policy-makers with quantitative information on the public preferences for potential future landscape protection, management and planning programmes. The results highlight the important role that the conservation of native forests and the promotion of organic farming may play in the management of European landscapes, which in turn is found to be strongly culturally dependent.

Keywords: European Landscape Convention; discrete choice experiment; Llanada Alavesa; random parameter logit; willingness to pay; welfare measure.

1 **1. Introduction**

2 Landscapes, designated as collections of forests, fields, wetlands, urban areas and more, and 3 the areas where they intersect, are an integral part of individual and communal well-being. 4 These areas are continuously changing because they are the expression of the dynamic 5 interaction between environmental and cultural forces. More specifically, the transformations 6 of landscapes can be attributed to factors such as population growth, changes in lifestyle 7 preferences, competing demands for agricultural land, fuel security and natural disasters (van 8 der Heide and Heijman, 2013). Moreover, the current changes are increasingly regarded as a 9 threat because they are often characterised by the loss of diversity, coherence and identity of 10 the existing landscapes (Balej et al., 2010). These rapid and sometimes chaotic landscape 11 changes can result in the need to manage landscapes based on a balanced relationship between 12 social needs, economic activity and the environment.

13 This is the case for European landscapes, which are complex assemblages of forests, fields, 14 wetlands and human settlements that have been subjected to a high level of transformation 15 throughout history. Forest areas in particular render a number of extremely important 16 economic, ecological and social services, along with being part of Europe's identity and cultural 17 heritage (Edwards, 2006). In fact, forest landscape restoration should be carried out under the 18 assumption that improving the flow of forest goods and services requires a balance between 19 livelihoods and environmental protection, and that this is best achieved within dynamic, 20 multifunctional landscapes (Déjeant-Pons, 2006a).

The approval of the European Landscape Convention (ELC) (Council of Europe, 2000) initiated more research and action programmes relating to landscapes in most European countries than ever before. A significant difference from older regulations regarding landscape protection is that all kinds of landscapes are involved, not only especially valuable locations, such as natural protected sites. Natural, rural, urban and peri-urban areas are considered, which

encompass land, inland water and marine areas. European landscapes are recognised as a basic
component of the European natural and cultural heritage that can be a favourable resource for
economic activity.

29 The ELC introduces a series of formal landscape definitions as well as a series of 30 recommendations. Landscape is defined as "an area, as perceived by people, whose character is 31 the result of the action and interaction of natural and/or human factors". The aim of the ELC is 32 to promote landscape protection, management and planning, and to organise European co-33 operation on landscape issues. The general public, as opposed to solely academic experts or 34 political officials, is set as the main information source to guide the ELC process (see Jones, 35 2007). Although general and specific measures are proposed, the ELC is not explicit regarding how 36 to proceed, and gives flexibility for determining how the agreement is implemented (Déjeant-Pons, 37 2006b). It is clear that the competent public authorities are responsible for organising the public's 38 participation, by collecting their aspirations and translating them into policy actions.

39 Opened for signature on 20 October 2000, the Convention entered into force on 1 March 2004, 40 after 10 member states had ratified it. Spain ratified the Convention in 2008, and the Basque 41 Country in 2009. Therefore, with the adherence of the Basque Country, the Basque authorities 42 made a commitment to promote the ELC principles and a landscape law so that the quality of 43 different landscapes, including native forests, is preserved and improved. However, due to the 44 nature of landscapes as public goods (i.e. the landscape itself is non-excludable and non-rival in 45 nature) and the complex definition and holistic nature of landscapes, effective governance is 46 often complex and challenging (for different interpretations and connotations of the term 47 "landscape" see van der Heide and Heijman, 2013). The challenge for policy-makers is to find a 48 way to quantify the value of landscapes to the local residents and to incorporate these values 49 into their landscape protection, management and planning decisions (Johnston and Duke, 2007).

50 In this context, the Discrete Choice Experiment (DCE) valuation technique can enrich the 51 process of landscape decision-making mainly due to its flexibility and ability to take into account 52 the multidimensional nature of landscapes and to provide detailed information about marginal 53 changes in landscapes as well as trade-offs between the landscape attributes themselves and 54 between the landscape attributes and money (Adamowicz et al., 1998; Bennett and Blamey, 55 2001; Bateman et al., 2002). DCE has been argued to be ideally suited to informing both the election and the design of multidimensional policies (Hanley et al., 2001; Horne et al., 2005) and 56 57 seems to be an appropriate valuation method to achieve the objective of managing landscapes 58 under the ELC framework, especially given the emphasis it puts on the general public's 59 preferences as the main information source to guide it.

60 In this paper, we provide empirical evidence on the suitability of the Discrete Choice 61 Experiment (DCE) methodology for valuing multidimensional landscapes under the ELC. An 62 application in the Basque Country, Spain, serves as an example to illustrate that the DCE 63 methodology may be suitable for supporting the aims of the ELC because: (i) it is a tool for public 64 consultation; (ii) it offers an insight into the relative attractiveness of key landscape attributes, 65 such as native forests and farming activities; and (iii) it provides policy-makers with quantitative 66 information on the public preferences for potential future landscape protection, management 67 and planning programmes. The results highlight the important role that the conservation of 68 native forests and the promotion of organic farming may play in the management of European 69 landscapes, which in turn is found to be strongly culturally dependent.

The paper is structured as follows. Section 2 describes the main landscape characteristics of the case study area. Section 3 is devoted to the DCE methodology, concerning the survey design, data collection and econometric specification. Section 4 presents a detailed analysis of the results, and Section 5 provides a discussion and some concluding remarks.

74 2. Case study description

75 The DCE involves the landscapes of a specific area known as Llanada Alavesa. This area is 76 located in the province of Araba, one of the three provinces that form the Basque Autonomous 77 Community (BAC) in Spain (see Figure 1). Llanada Alavesa is a wide plain extending into the 78 central and north-eastern part of Araba and encompasses 256,003 people according to Basque 79 Statistic Office (EUSTAT), almost 80% of the population of the province of Araba in 2013. 80 Different types of landscapes, natural habitats and human activities coexist in this area, including 81 forests, farming activities, industry, urban areas, infrastructure and swamps (Latasa et al., 2012; 82 Latasa et al., 2014).

83 The landscapes of Llanada Alavesa exhibit great richness of flora, fauna and habitat diversity 84 (Loidi et al., 2011; Lozano et al., 2012). Its native forests, which are the cornerstone of the 85 society, culture and landscape, take up 39% of the surface. These forests support centenary and 86 even millenary trees, such as oaks (Quercus robur) and gall oaks (Quercus faginea). However, 87 the number of these millenary tree species has been reduced mainly due to extraction for 88 firewood, coal and plank production. Ash trees (Fraxinus excelsior), maple trees (Acer 89 pseudoplatanus) and beech trees (Fagus sylvatica) are also found in this area. Moreover, these 90 native forests are home to threatened, vulnerable and rare animals, such as otters (Lutra lutra) 91 and European minks (mustela vitrelola), and plant species, such as Pentaglottis sempervirens 92 and Littorella uniflora (Uribe-Echebarría, 2010).

Given the favourable topography of Llanada Alavesa, 45% of its area is devoted to farming,
which produces mainly potatoes, cereal and beetroot. Approximately 29% of the total surface
of Llanada Alavesa is devoted to intensive farming, which employs chemical fertilisers and
pesticides as well as making high use of machinery to achieve a higher level of production. The
forest plantation surface, which is mainly devoted to timber crops, is included within this area.
However, it represents a very small percentage of the total intensive farming surface (less than
3%). Approximately 16% of the land of Llanada Alavesa is covered by organic farming, which

does not use non-natural synthetic products or genetically modified organisms and makes more
 limited use of machinery, resulting in environment-friendly and high-quality production.

Urban and industrial sites, in addition to infrastructure are also present. Companies working in the automotive industry, wind generation and machine tools are located in this area largely due to its good connectivity. It is estimated that currently 14% of the area of Llanada Alavesa is occupied by urban, industrial and economic activity sites as well as by infrastructure. Henceforth, this landscape attribute definition will be referred to as cemented surface (i.e. 14% is cemented surface).

108 Recreation areas are also common in this area, including the *Ullibarri* and *Urrunaga* 109 reservoirs. These swamps depict a unique leisure location where fishing, canoeing, cycling and 110 hiking are possible. Llanada Alavesa also presents a large cultural heritage. Romanesque chapels 111 and churches, megalithic monuments, medieval towns, typical Basque farms (*baserri*) or a 112 branch of the way of St James are located in this historical area. In addition, many of the villages 113 of Llanada Alavesa hold ancient rural traditions.

114 The largest transformations of the landscapes of this area have always been spurred by 115 actions within the primary sector, particularly from agriculture. The most drastic changes have 116 been initiated by land consolidation that took place during the 1960s (i.e. causing field 117 boundaries and bank vegetation to largely disappear). To a lesser extent, changes can also be 118 attributed to agricultural mechanisation and intensive farming practices. Today, the biggest 119 alterations come from infrastructure, urban and non-residential land development projects. In 120 addition, the Partial Territorial Plan of Central Araba and General Urban Planning Plans are also leading to landscape transformation processes of great territorial magnitude. In the last 30 121 122 years, the cemented surface of Llanada Alavesa has approximately tripled, mainly due to 123 economic activities, infrastructure and residential land uses (Ruiz and Galdós, 2013).

124 The current land use trend in Llanada Alavesa is a decrease in the agricultural land together 125 with an increase in the cemented surface and to a lesser extent, in the native forest area (Latasa 126 et al., 2014; Ruiz and Galdós, 2013). As in many parts of Europe, the cemented surface for 127 residential land use, economic activities, and urban and developable areas has potential to 128 continue expanding at a high rate. A gain in the native forest area is also expected due to both 129 forest protection regulations (including the European Nature 2000 network) and a large 130 decrease in cattle farming in the area (which is resulting in meadow and scrublands changing to 131 a native forest state).

The persistent changes in landscapes have raised concerns about the sustainability of development based on social needs, economic activity and the environment. With the adherence of the Basque Country to the ELC, the Basque authorities ought to adopt management measures to preserve and improve the quality of different landscapes, including native forests, farming (intensive and organic), cemented surface as well as recreation areas and cultural heritage.

3. The discrete choice experiment for landscape valuation

139 In recent years, various DCEs have been applied to the economic valuation of landscape 140 changes in Europe. To our knowledge, the first European DCE study was reported by Bergland 141 (1997) and attempted to measure the value of certain attributes (hedgerows, creeks, fences, 142 vegetation islands and paths) of the agricultural landscape in Norway. A year later, Bullock et al. 143 (1998), Hanley et al. (1998a) and Hanley et al. (1998b) estimated the willingness to pay (WTP) 144 for different attributes of the deer stalking experience in the Scottish highlands, for the 145 conservation and landscape benefits of environmentally sensitive areas in Scotland, and for 146 changes in landscape elements in the UK's public forests, respectively.

Since then, a growing number of DCE studies have been published in Europe. For example,
Rambonilaza and Dachary-Bernard (2007) used the DCE method to examine public preferences

149 for three landscape features (hedgerows, farm buildings and scrubland) in the Monts d'Arrèe 150 region in Brittany (France). More recently, Grammatikopoulou et al. (2012) employed a DCE to 151 evaluate a management programme that provides certain landscape attributes (proportion of 152 uncultivated land, number of plant species, grazing animals, water protection zones and state 153 of production buildings) in the southern Finnish agricultural landscape. Another recent DCE 154 application was conducted by Liekens et al. (2013) to evaluate the public preferences associated 155 with land use changes from agricultural land to different types of nature in the Flemish region 156 in Belgium.

157 Spain also presents widespread use of DCE applications concerning landscape valuation. For 158 instance, Kallas et al. (2007) applied this stated preference (SP) method for valuing the 159 multifunctionality of the agricultural landscape of Tierra de Campos in Castile and Leon. 160 Similarly, Arriaza et al. (2008) estimated the social demand for agricultural multifunctionality 161 from mountain olive groves of Andalusia. Another Spanish example is the DCE conducted by 162 Domínguez-Torreiro and Soliño (2011), which estimated the welfare change associated with 163 multifunctional rural development programmes in Cantabria. In the Basque Country, it is 164 possible to find DCE applications focused on the economic valuation of the Basque forests 165 (Pascual, 2007), the natural area of Mount Jaizkibel (Hoyos et al., 2009) and a regional Natura 166 2000 network site (Hoyos et al., 2012). Although these studies have been valuable inputs for the 167 survey design, the present study contributes to the recent landscape valuation literature by 168 providing policy-makers with a specific tool to manage landscapes in the context of the ELC.

169 **3.1 Survey design**

A valuation survey was conducted in Araba to evaluate the public's preferences for the main attributes of the landscapes of Llanada Alavesa. The respondents were first informed about the current situation of the landscapes of Llanada Alavesa and the need for a policy for the

protection, planning and management of them. Further on in the questionnaire, the attributesto be valued and their levels of provision were described.

For the definition of the landscape attributes and levels of provision, we carried out an extensive literature review on European DCEs valuing landscapes, investigated the landscape features of Llanada Alavesa and considered expert advice obtained from bio-geographers and economists as well as conducted a focus group discussion. The focus group was led by a professional in October 2012 to test the appropriateness of the landscape attributes (and their levels), photographic materials, valuation context and payment vehicle.

181 The attributes and levels considered in the final version of the questionnaire were (see Table 182 1): (1) Native forests, represented by the percentage of the area of Llanada Alavesa covered by 183 native forests; (2) Intensive farming, represented by the percentage of the land devoted to 184 intensive farming (including forest plantations); (3) Organic farming, measured by the 185 percentage of the land of Llanada Alvesa taken up by organic farming; (4) Cemented surface, 186 represented by the percentage of the surface occupied by urban, industrial and economic 187 activity sites as well as by infrastructure; and (5) Recreation areas, measured by the level of 188 conservation and protection of the recreation areas (e.g. swamps, picnic areas) and the cultural 189 heritage (e.g. megalithic monuments, the branch of the way of St James) found in the area of 190 Llanada Alavesa. All these attributes were specified on four different levels, the first of which 191 corresponded to the current situation (status quo), whereas the remaining three represent 192 hypothesised changes. Note that the hypothesised future levels of the attributes were 193 estimated and provided by experts in this field and these were found to be both credible and 194 understandable by the focus group participants.

A monetary attribute was also included. The proposed payment vehicle was an annual payment through a new tax to be paid by the citizens of Araba to an organisation exclusively dedicated to coordinating the action plans. This payment vehicle was preferred to voluntary

donations since respondents may have incentives to free-ride with the latter (Whitehead, 2006).
The usual reminder of the budget constraint was also incorporated. The focus group participants
found the payment vehicle and cost levels credible.

The six attributes and their varying levels allowed a large number of alternatives to be constructed $(4^5 \times 5^1)$. In order to reduce the number of alternatives and choice sets, we applied a *D*-efficient main-effects fractional factorial design (Scarpa and Rose, 2008). The design was also constrained so that the sum of the land devoted to the first four attributes considered would not exceed the total land percentage in the current situation (98%). As a result, 120 versions of the choice sets were randomly divided into 20 blocks.

207 Figure 2 shows an example of a choice set used in the questionnaire. Each respondent was 208 presented with six choice sets and asked to choose one of the three options presented for each 209 set. The choice set included some pictures of the attributes considered and a graphical 210 representation of their levels to facilitate the choice task understanding. The questionnaire also 211 tested the understanding of the choice task by including an additional "rationality" choice set in 212 which respondents faced the same attribute levels as in option A (status quo) but with a higher 213 cost, so we expected respondents always to choose the status quo (Hoyos, 2010). Data on the 214 respondents' social and economic characteristics were collected at the end of the survey.

215 **3.2 Data collection and final number of observations**

Data was gathered through face-to-face interviews in order to encourage participation and facilitate the understanding of the choice task. The relevant population considered was that of the Basque province of Araba, accounting for 266,014 residents aged at least 18 (EUSTAT). The pilot survey was conducted in November 2012 with 60 individuals, and the final questionnaire was administered in January 2013 with a stratified random sample of 521 individuals selected from the relevant population. The strata used included the age, gender and size of the town of residence, following official statistical information provided by EUSTAT. It is important to analyse the validity of responses we obtained from the choice sets for the further analysis of welfare measures. Out of the 521 survey respondents, everyone understood the objective of contributing or not to the proposed initiative. However, a total of 317 respondents chose option A (status quo) in the first choice card. We used a set of follow-up questions to identify protest respondents based on the state of practice in environmental valuation with SP methods (e.g. Hanley et al., 2007; Lindberg et al., 2009; Jacobsen et al., 2011; Martin-Ortega et al., 2011; Schaafsma et al., 2012).

230 278 respondents were identified as protesters, of whom most stated that they already paid 231 too much tax followed by the opinion that this initiative should be financed entirely by the 232 Government. We devote this high presence of protest responses to the adverse economic 233 situation at the time the survey was conducted. In order to obtain reliable and unbiased welfare 234 estimates, we excluded protest responses from the sample, as is the common practice in the 235 literature (e.g. Glenk et al., 2011; Martin-Ortega et al., 2011; Liekens et al., 2013). Moreover, 236 respondents who failed to pass the rationality test (25 respondents) were also excluded from 237 the final sample.

All in all, the data analysis used 218 completed questionnaires, yielding 1308 observations as each respondent faced six choice sets. The benefit of higher reliability of the results comes at the expense of slightly reducing the representativeness of the sample, which has been the case in many DCE studies (e.g. Glenk and Colombo, 2011; Glenk et al., 2011; Jacobsen et al., 2011; Hoyos et al., 2012). The total number of observations is comparable with other European DCE studies for landscape valuation (e.g. Campbell, 2007; Colombo and Hanley, 2008; Elsasser et al., 2010; Domínguez-Torreiro and Soliño, 2011; Hoyos et al., 2012; Olschewski et al., 2012).

245 3.3 Econometric specification

246 In order to convert the individual choice responses into estimated parameters, the DCE 247 employs the behavioural framework of the Random Utility Theory (RUT) developed by

248 McFadden (1974). The utility function for individual i choosing alternative j on choice 249 occasion t is given by:

$$U_{it,j} = V_{it,j} + \varepsilon_{it,j}, \tag{1}$$

where $V_{it,j}$ is the deterministic part of the latent utility that contains factors observable by the analyst and $\varepsilon_{it,j}$ is an error term. In order to analyse the data, a Random Parameter Logit (RPL) model, which has recently been used in the field of landscape valuation, was applied (Bliem et al., 2012; Schaafsma et al., 2012; Garrod et al., 2013).

In contrast to the traditional Multinomial Logit (MNL) model (McFadden, 1974; Louviere et al., 2000), the RPL specification is not subject to the undesirable Independence of Irrelevant Alternatives (IIA); it accounts for unobserved heterogeneity by allowing (some of) the parameters of the utility function to vary according to some distributions as well as considering that a respondent makes choices in more than one choice situation (Train, 2003; Siikamäki and Layton, 2007).

260 In the RPL model, the probability that individual i chooses alternative j in choice situation 261 t is:

$$P_{it,j} = \frac{\exp(x'_{it,j}\beta_i)}{\sum_{j=1}^{J_i} \exp(x'_{it,j}\beta_i)},$$
(2)

where $x_{it,j}$ is a $(K \times 1)$ vector of the attribute levels of alternative j (from a total of J_i alternatives) for individual i (from a total of N individuals) in choice card t (from a total of T_i choice cards). In our application, the full vector of K parameters is continuously distributed across individuals with:

$$\beta_i = \beta + \Delta z_i + \Gamma v_i, \qquad (3)$$

where β is a parameter vector representing the fixed means of the random parameter distribution, z_i is the vector of observed individual-specific characteristics, Δ is the associated parameter matrix and Γ is the unknown lower triangular matrix to be estimated. The random unobserved taste variation is represented by v_i , a vector of uncorrelated random variables characterised by:

$$E(v_i) = 0, \quad \text{Var}(v_i) = \Sigma = \text{diag}\left[\sigma_1, \sigma_2, ..., \sigma_K\right].$$
(4)

The estimation procedure by maximising the simulated log-likelihood function is described in
Greene and Hensher (2003), Hensher and Greene (2003) and Train (2003).

273 **4. Results**

274 4.1 Basic statistics

275 More than half of the sample respondents (54.6%) live and work or study in Llanada Alavesa. 276 Moreover, we found that almost 57% of the respondents appreciate quite or a great deal that 277 products are protected under a quality label (certificate of origin, organic farming, Basque label 278 ...); almost 58% of the respondents are very concerned about biodiversity loss and 279 approximately 57% about landscape quality loss. Table 2 shows the descriptive statistics of 280 socio-economic variables obtained with this application along with their corresponding 281 population values.

The sample mean age (42.28 years); gender decomposition (50.4% male and 49.6% female); personal income (the majority, 47.4%, earning between €500 and €1,500 per month and 24.6% with no personal income); employment status (54.4% salaried employee or self-employed and 11.5% unemployed); education level (6.4% with no official education and 18.8% with a university degree or equivalent qualification); and the percentage of the population of Araba residing in 287 Llanada Alavesa (76.2%) are generally in line with those of the overall population of Araba in
288 2013 according to EUSTAT (see Table 2).

289 Moreover, the last four rows of Table 2 present the additional dummy variables created for 290 the socio-economic variables considered in the model estimation stage: Resident (taking the 291 value 1 if the respondent lives in a township of Llanada Alavesa and 0 otherwise), Low income 292 (for respondents' net monthly disposable income lower than €500 or respondents with no 293 income at present and 0 otherwise), Basque identity (taking the value 1 if the respondent 294 considers himself/herself as having a Basque cultural identity at an above average level and 0 295 otherwise) and *High recreation* (taking the value 1 if the respondent uses the recreation areas 296 of Llanada Alavesa frequently or very frequently and 0 otherwise).

297 4.2 Model specifications and estimation results

The data obtained in the experiment were examined using the NLOGIT version 4.0 software (Greene, 2007). Table 3 shows the estimations corresponding to the MNL and different RPL model specifications. We included an alternative-specific constant (ASC) in all the utility specifications for the non-status quo options (*ASC1* and *ASC2*). Their significant and positive coefficients under all the estimated models suggest that all else being equal, respondents tend to favour moving away from the current situation to a situation with change.

The MNL was first estimated in order to obtain a first insight into the data. Then, different RPL specifications were estimated. The general indirect utility function used for the RPL specifications was the following:

$$V_{it,j} = ASC_{j} + \beta_{NFi} \text{Native forests}_{it,j} + \beta_{IFi} \text{Intensive farming}_{it,j} + \beta_{OFi} \text{Organic farming}_{it,j}$$
(5)
+ $\beta_{CSi} \text{Cemented surface}_{it,i} + \beta_{RAi} \text{Recreation areas}_{it,i} + \beta_{Cosi} \text{Cost}_{it,i},$

307 where ASC_{j} is the alternative-specific constant under options B and C, β_{NFi} , ..., β_{Costi} are the 308 attribute parameters defined according to (3) and *Native forests*,..., *Cost* are the attribute 309 levels described in Table 1. The possible randomness of the attribute parameters was tested 313 using the Lagrange Multiplier (LM) test proposed by McFadden and Train (2000) which presents
correct empirical size (Mariel et al., 2013). As a result, the coefficients associated with the *Native forests, Cemented surface* and *Cost* attributes are random.

313 We applied the empirical approach proposed by Hensher and Greene (2003) to describe 314 graphically the empirical distributions for the random parameters. Figure 3 shows the empirical 315 shape of each distribution. The parameters associated with the Native forests and Cemented 316 surface attributes were assumed to be normally distributed, while the lognormal distribution 317 was chosen for the *Cost* attribute parameter. Moreover, the lognormal distribution (with a sign 318 change) for the cost parameter assures finite moments for the distributions of WTPs (Daly et al., 319 2012). The stability and precision of each and every model was confirmed when using 2,000 320 Halton draws (Hensher and Greene, 2003).

321 In the first RPL specification (RPL-1), individual preference heterogeneity is covered by 322 unobserved influences affecting respondents' utility. As expected given the results of the LM 323 test for selecting random parameters, the standard deviations of the assumed random 324 parameters are significant, suggesting the presence of unobserved heterogeneity in the 325 preferences for changes in the levels of the Native forests, Cemented surface and Cost attributes. As a further step, we explored possible factors explaining the preference heterogeneity by 326 327 estimating the RPL-2 model including the socio-economic variables affecting the parameters' 328 mean. Under this RPL-2 specification, all the estimated coefficients associated with landscape 329 attributes are on average significant and positive except the coefficient for the *Cemented surface* 330 attribute, which on average is insignificant. However, the standard deviation of its random 331 parameter is significant, which suggests that *Cemented surface* attribute is controversial with 332 positive and negative preferences toward percentage increases in it.

Furthermore, the significant coefficient estimates of the interactions between socioeconomic and attribute variables means that, all else being equal, (i) having Basque cultural

335 identity increases the utility regarding increases in land devoted to native forests and/or organic 336 farming, (ii)living in a township of Llanada Alavesa results in a utility decrease toward increases 337 in land covered by intensive farming , (iii) earning less than €500 per month or having no income 338 decreases the utility with regard to an increase in the level of conservation and protection of 339 recreation areas and (iv) using recreation areas frequently or very frequently results in a utility 340 increase concerning an improvement in the level of conservation and protection of recreation 341 areas. Explaining the individual preference heterogeneity by both socio-economic 342 characteristics and other unobserved influences improves the log-likelihood, AIC, BIC and 343 Pseudo R-squared (see Table 3). Moreover, we allowed for free correlation between random 344 parameters but as no significant correlations were found, we used the RPL-2 specification for 345 welfare analysis.

346 **4.3 Welfare measures**

The welfare change, either positive or negative, related to a hypothetical choice scenario can be estimated by using the compensating surplus (CS) following Small and Rosen (1981) and Hanemann (1984):

$$\mathbf{CS} = -\frac{1}{\mu} \left[\ln \left(\sum_{j=1}^{J_i} \exp\left(x_{ij}^{\prime 0} \lambda\right) \right) - \ln \left(\sum_{j=1}^{J_i} \exp\left(x_{ij}^{\prime 1} \lambda\right) \right) \right], \tag{6}$$

350 where μ is the marginal utility of income (usually represented by the coefficient of the cost attribute, β_{Cost}), λ represents the vector of parameters corresponding to landscape attributes 351 and x_{ij}^0 and x_{ij}^1 correspond to the vector of landscape attributes before and after the change 352 353 under consideration. Thus, the Hicksian CS measures a change in the expected utility due to a 354 change in the level of provision of the attribute(s) by weighting this change by the marginal 355 utility of income. Simplifying the above equation, the WTP for a marginal change in the level of 356 provision of each landscape attribute is obtained by dividing the coefficient of the landscape 357 attribute by the coefficient of the cost attribute (Haab and McConnell, 2002).

In this section, we present the unconditional simulation of the WTP results derived from out-of-sample populations by randomly sampling each individual from the full distribution (Krinsky and Robb, 1986). Table 4 reports the simulated WTPs for the RPL-2 model with the tenth and ninetieth percentile points from the resulting distribution of each WTP because its distribution is non-standard (Daly et al., 2012).

363 In the presence of the RPL-2 model specification, WTP calculations should take into account 364 both the effect of socio-economic variables and the possible randomness of the parameter. A 365 baseline scenario was first specified with all the socio-economic dummy variables equated to 366 zero. By setting the dummies of the socio-economic variables to one, their effect can be 367 examined (see Table 4). Hence, for example, the WTP for a 1% increase in the Native forests 368 attribute, with its parameter distributed as normal and the Cost attribute parameter as 369 lognormal, when the respondent has Basque cultural identity at an above average level (Basque 370 *identity* = 1) was specified as:

WTP_{NF} =
$$-\frac{\left(\hat{\beta}_{NF} + \hat{\delta}_{BI} \cdot 1 + \hat{\sigma}_{\beta_{NF}} \cdot \upsilon\right)}{-\exp\left(\hat{\beta}_{Cost} + \hat{\sigma}_{\beta_{Cost}} \cdot \upsilon\right)},$$
 (7)

where $\hat{\beta}_{NF}$ and $\hat{\beta}_{Cost}$ are the estimated means of the *Native forests* and *Cost* attribute parameters, respectively, $\hat{\sigma}_{\beta_{NF}}$ and $\hat{\sigma}_{\beta_{Cost}}$ are their corresponding estimated standard deviations, $\hat{\delta}_{BI}$ is the estimate associated with having Basque cultural identity and $\upsilon \sim N(0,1)$. In order to calculate the WTP means (reported in the last row of Table 4), we used weights corresponding to the proportion of each group in the population.

The mean annual WTP for a 1% increase in the land area covered by native forests is estimated at €1.00 per person, in 2013 values. The WTP increases to €2.58 if the respondent's cultural identity is Basque. Similarly, the mean annual WTP for a 1% increase in the land devoted to organic farming is estimated at €1.64, but the WTP increases to €3.32 if the individual again has Basque cultural identity. The important role that Basque cultural identity plays in this
economic valuation study is in line with previous DCE studies conducted in the Basque Country
(Hoyos et al., 2009; Hoyos et al., 2012).

383 The mean annual WTP for a 1% increase in the land area devoted to intensive farming is 384 estimated at €0.37. However, the respondents residing in one of the municipalities of Llanada 385 Alavesa have on average almost a null annual WTP (€0.03). Although the mean WTP to increase 386 the cemented surface in Llanada Alavesa is zero on average, from its distribution we can observe 387 that there are people who are willing to pay for increases in the cemented surface and there are 388 other people who are willing to prevent this increase, leading therefore to a compensated effect. 389 Eventually, the annual WTP to improve the level of conservation and protection of the 390 recreation areas and cultural heritage is estimated at €6.75. In this case, the preferences also 391 differ among the population: respondents with an income less than €500/month or no income 392 have a WTP of €1.06 while the WTP of respondents using the recreation areas frequently or very 393 frequently is estimated at €10.57.

394 **4.4 Compensating surplus for alternative landscape management scenarios**

395 The welfare measures obtained by estimating the mean marginal WTP in the previous section 396 also allow us to estimate the changes in the welfare of the population of Araba associated with 397 different landscape protection, management and planning options. In order to estimate the CS 398 for different options of interest, we proposed three different future scenarios, altering the 399 attribute levels and considering feasible land use changes: (1) Promotion of native forests and 400 organic farming; (2) Promotion of intensive farming and cemented surface; and (3) Promotion of 401 recreation areas and cultural heritage. Table 5 presents the changes proposed in relation to the 402 status quo by showing the attribute levels corresponding to each of these three scenarios. Note 403 that the levels we used to construct the different scenarios are those provided by experts and 404 used in the DCE sets.

405 Table 6 shows the estimated annual CS (i.e. the estimated social benefits) for each scenario 406 proposed using equation (6). Under scenario 1, which promotes native forests and organic 407 farming and can be associated with the ecological features of the area, the total welfare benefits 408 for the population of Araba are estimated at €5.05 million per annum. Scenario 2, characterised 409 by the promotion of intensive farming and cemented surface and which can be associated with 410 economic (or production) interests, would entail on average a social loss estimated at €6.46 411 million per annum for the citizens of Araba. Finally, with scenario 3, which enhances the cultural 412 heritage and recreation areas, keeping the rest of the landscape attributes at their current levels, 413 the citizens of Araba would experience on average a social benefit estimated at €4.35 million 414 per annum.

415 In order to assess the robustness of the welfare estimates, a sensitivity analysis was also 416 carried out. Given that the robustness of the estimated welfare benefits depends on the sample 417 representativeness, we have conducted a sensitivity analysis for the observed sample 418 proportions of the dummy variables that could not be verified with official statistics (i.e. Basque 419 Identity and High Recreation variables). The 95% confidence interval for the case of Basque 420 identity is (10.17%, 20.15%), while that for High recreation is (40.39%, 54.11%). Changing the 421 Basque identity sample proportion to the lower and upper bound of its corresponding 95% 422 confidence interval, we estimate the social benefits under scenario 1 between €4.53 million and 423 €5.46 million (i.e. -10% and 8%). Similarly, when modifying the corresponding sample proportion 424 of *High recreation*, the social benefits corresponding to scenario 3 vary between €4.15 million 425 and €4.53 million (i.e. -4% and 4%). Given that these are considerable variations, welfare 426 estimates should be seen with a certain degree of caution. Furthermore, the sensitivity analysis 427 highlights the importance of collecting this type of attitudinal information with particular 428 attention.

429 **5. Discussion and conclusions**

430 Landscapes, which are a key element of individual and social well-being, continuously change 431 and evolve through natural and human-induced processes and activities. The ELC highlights that 432 the protection, management and planning of landscapes can provide a range of benefits in the 433 cultural, ecological, environmental and social fields as well as contributing to job creation. In 434 order to aid the optimal design of landscape protection, management and planning, policy-435 makers need a proper means of accounting for all these benefits in the public preferences for 436 the policy intervention. This paper estimates the public preferences for landscape protection, 437 management and planning under the ELC using the DCE valuation method as a case study of the 438 Basque area of Llanada Alavesa, Spain.

439 For the purpose of policy, this study presents useful information to help policy-makers 440 resolve the issue of disaggregating protection, management and planning policies into suitable 441 landscape attributes and levels. Taking into account the public's preferences of the population 442 of Araba, the annual per capita mean marginal WTP to increase the land area of Llanada Alavesa 443 under native forests and organic farming is estimated at €1.00 and €1.64, respectively, while for 444 the land area under intensive farming is estimated at €0.37. Further, we found a null mean 445 marginal WTP for percentage increases in the cemented surface. Finally, we estimated the mean 446 marginal WTP to improve the level of conservation and protection of the recreation areas by 447 €6.75 per person and year.

In the face of landscape protection, management and planning, we found different and sometimes conflicting landscape preferences. We applied the RPL model in order to investigate and explain the individuals' preference heterogeneity. Firstly, we identified unobserved influences affecting respondents' utility for changes in the levels of the native forests, cemented surface and cost attributes. A further exploration of the heterogeneity through an RPL model with relevant socio-economic variables helped to describe part of the unobserved

454 heterogeneity. This model identified a strong Basque cultural identity as a main explanatory 455 variable for higher values regarding native forests and organic farming. Interestingly, the mean 456 annual WTP of individuals residing in one of the Llanada Alavesa's municipalities for increases in 457 the land devoted to intensive farming seems to be lower than the mean annual WTP of the 458 overall Araba population. The mean annual WTP for cultural heritage and recreation areas' 459 conservation seems to be higher for high-frequency users of the recreation areas, but as 460 expected by economic theory, this is lower for individuals with low income or no income. 461 Moreover, we found opposite preferences toward percentage increases in cemented surface.

462 The expected welfare benefits in monetary terms of specific landscape scenarios are useful for 463 facilitating future cost-benefit examinations. Among the three different hypothetical scenarios 464 proposed, we found that the promotion of native forests and organic farming, and the 465 development of recreation areas and cultural heritage would bring on average welfare benefits 466 estimated at €5.05 and €4.35 million per year, respectively. In contrast, the specified promotion 467 of intensive farming and cemented surface would entail on average an annual social loss from 468 the population's point of view estimated at ≤ 6.46 million. These results suggest that policy-469 makers might attempt to enhance the ecological values of the landscapes that could be 470 compatible with an improvement in the conservation level of the recreation areas and cultural 471 heritage.

Under the current trend of increasing cemented surface and forest areas at the expense of decreasing agricultural land in Llanada Alavesa, our study shows that these land transformations are partially in line with public preferences. The promotion of the area under native forests is supported by the population of Araba, while the one under urban areas seems to be much disputed. With regards to the consequence of more limited farm land, the citizens of Araba seem to give more recognition to the organic farming than the intensive farming. Hence, our results can be seen as the starting point for dealing with the trends in the landscapes of Llanada Alavesa

under the ELC. It is also important to note that the costs of implementing a given landscape
programme would have to be considered to evaluate whether that programme would succeed
in a cost-benefit analysis.

482 Policy-makers should also bear in mind the economic and cultural context in which the 483 landscape management policies will be implemented as well as the potential landscape use 484 conflicts. On the one hand, this study has been developed in a context of economic recession, 485 which was partly reflected in the high number of protest responses. Given that the 486 unemployment rate in the first guarter of 2013 was around 13% both in Araba and in the Basque 487 Country (EUSTAT), and about 27% in Spain (National Institute of Statistics, INE), the society might 488 be more concerned about unemployment and social assistance while environmental issues 489 seem to be in the background in times of economic crisis.

490 On the other hand, given that the natural environment plays a central role in the Basque 491 cultural tradition and that a matriarchal culture such as that of the Basques feels close 492 attachment to the land, cultural identity is found to be a key factor explaining the social benefits 493 related to the main ecological features of the landscapes of Llanada Alavesa. The positive 494 influence of Basque cultural identity on native forests is in line with other DCEs conducted in the 495 Basque Country (Hoyos et al., 2012), while that on organic farming is found first in this study. In 496 fact, people who support organic farming production may do so for health reasons or just 497 because they want to protect the environment. Native forests are an important pillar of the 498 Basque society and culture, and this is also true in Llanada Alavesa. This is reflected in the local 499 language "euskera", which contains the word "basoa" (forest) in the roots of several significant 500 words, as well as in the mythology of the country (Hoyos et al., 2009; Palacios, I., 2011). Thus, 501 considering and understanding the cultural values that predominate in a specific area, region or 502 country could help in designing landscape-related policies.

503 Eventually, since designers and policy-makers have to make choices between different and 504 usually competing uses of landscapes, they should be aware of the preferences of different 505 segments of the population to reduce the potential conflicts that may arise through the 506 landscape changes proposed. For example, the funds for landscape protection, management 507 and planning could be transferred not only to invest in the corresponding landscape programme, 508 but also to compensate for the losses that local farmers could experience if the proposed 509 landscape programme significantly affects their activity. Moreover, some local farmers might 510 need support in terms of information and economic incentives to move from intensive farming 511 to organic farming. In fact, nowadays, both at the European level through the Common Agricultural Policy (see European Commission, 2013) and at the Spanish Autonomous 512 513 Communities level, there are different programmes that motivate farmers to introduce organic 514 farming. In the Basque Country, the Rural Development Plan (2007–2013) specifies different aids 515 that provide financial support depending on the number of hectares under organic farming, 516 livestock and socio-economic characteristics of the farmers.

This study concludes that the DCE may be a useful tool for establishing the principles and objectives of the ELC since: (i) it supports public participation; (ii) it highlights the trade-offs between different landscape attributes from the general public's point of view; (iii) it takes into account the heterogeneous preferences of the population; and (iv) it offers well-informed advice in terms of the welfare benefits derived from different landscape management programmes. The present study has illustrated how useful results can be successfully obtained from a carefully designed DCE in a specific European landscape.

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533 **References**

- Adamowicz, W., Boxall, P., Williams, M. and Louviere, J., 1998. Stated preference approaches
 for measuring passive use values: choice experiment and contingent valuation. American
 Journal of Agricultural Economics 80, 64–75.
- Arriaza, M., Gómez-Limón, J.A., Kallas, Z. and Nekhay, O., 2008. Demand for non-commodity
 outputs from mountain olive groves. Agricultural Economics Review 9, 5–23.
- 539 Balej, M., Raska, P., Andel, J. and Chvátalová, A., 2010. Memory of a landscape a constituent
- 540 of regional identity and planning? In Andel, J., Bicik, I., Dostál, P., Lipský, Z. and Shahneshin,
- 541 S.G. (Eds), Landscape Modelling: Geographical Space, Transformation and Future Scenarios.
- 542 Springer, Dordrecht, pp. 107–121.
- Bateman, I.J., Carson, R.T., Day, B.H., Hanemann, W.M., Hanley, N., Hett, T. et al., 2002.
 Economic Valuation with Stated Preferences Techniques: A Manual. Edward Elgar,
 Cheltenham.
- Bennett, J. and Blamey, R., 2001. The Choice Modelling Approach to Environmental Valuation.
 Edward Elgar, Cheltenham.
- Bergland, O., 1997. Valuation of landscape elements using a contingent choice method. Paper
 presented at the 1997 EAERE conference, Tilburg, Holland, June.
- 550 Bliem, M., Getzner, M. and Rodiga-Laßnig, P., 2012. Temporal stability of individual preferences 551 for river restoration in Austria using a choice experiment. Journal of Environmental 552 Management 103, 65–73.
- Bullock, C., Elston, D.A. and Chalmers, N.A., 1998. An application of economic choice
 experiments to a traditional land use deer hunting and landscape change in the Scottish
 Highlands. Journal of Environmental Management 52, 335–351.

556 Campbell, D., 2007. Willingness to pay for rural landscape improvements: combining mixed logit

and random-effects models. Journal of Agricultural Economics 58, 467–483.

558 Colombo, S. and Hanley, N., 2008. How can we reduce the errors from benefits transfer? An

investigation using the choice experiment method. Land Economics 84, 128–147.

- 560 Council of Europe, 2000. European Landscape Convention. In Report and Convention.
- Daly, A.J., Hess, S. and Train, K.E., 2012. Assuring finite moments for willingness to pay in random
 coefficients models. Transportation 39, 19–31.
- 563 Déjeant-Pons, M., 2006a. The European Landscape Convention and forests. Proceedings of the

564 Forestry and our Cultural Heritage Seminar, 13-15 June 2005, Sunne, Sweden, pp 23–30.

565 Ministerial Conference on the Protection of Forests in Europe (MCPFE) Liaison Unit,

566 Warsaw, ISBN10 83-922396-3-6, ISBN13 978-83-922396-3-5.

- 567 Déjeant-Pons, M., 2006b. The European landscape convention. Landscape Research 31, 363–
 568 384.
- Domínguez-Torreiro, M. and Soliño, M., 2011. Provided and perceived status quo in choice
 experiments: implications for valuing the outputs of multifunctional rural areas. Ecological
 Economics, 70 2523–2531.
- Edwards, D., 2006. Social and cultural values associated with European forests in relation to key
 indicators of sustainability. EFORWOOD: Tools for Sustainability Impact Assessment,
 Deliverable D2.3.1, Forest Research, UK.
- Elsasser, P., Englert, H. and Hamilton, J., 2010. Landscape benefits of a forest conversion
 programme in North East Germany: results of a choice experiment. Annals of Forest
 Research 53, 37–50.

- European Commission, 2013. The Common Agricultural Policy after 2013. Agriculture and Rural
 Development. Available online at: <u>http://ec.europa.eu/agriculture/cap-post-2013/</u> (Last
 access: November, 2014).
- Garrod, G., Ruto, E., Willis, K., and Powe, N., 2013. Investigating preferences for the local delivery
 of agri-environment benefits. Journal of Agricultural Economics 65, 177–190.
- Glenk, K. and Colombo, S., 2011. Designing policies to mitigate the agricultural contribution to
 climate change: an assessment of soil based carbon sequestration and its ancillary effects.
 Climatic Change 105, 43–46.
- Glenk, K., Lago, M. and Moran, D., 2011. Public preferences for water quality improvements:
 implications for the implementation of the EC Water Framework Directive in Scotland.
 Water Policy 13, 645–662.
- Grammatikopoulou, I., Pouta, Salmiovirta, M. and Soini, K., 2012. Heterogeneous preferences
 for agricultural landscape in southern Finland. Landscape and Urban Planning 107, 181–
 191.
- 592 Greene, W.H. and Hensher, D.A., 2003. A latent class model for discrete choice analysis:
- 593 contrasts with mixed logit. Transportation Research Part B: Methodological 37, 681–698.
- 594 Greene, W.H., 2007. Nlogit version 4.0 reference guide. Econometric Software, Inc., Castle Hill.
- Haab, T.C. and McConnell, K.E., 2002. Valuing Environmental and Natural Resources: The
 Econometrics of Non-Market Valuation. Edward Elgar, Cheltenham and Northampton.
- Hanemann, W.M., 1984. Discrete/continuous models of consumer demand. Econometrica 52,
 541–561.
- Hanley, N., Colombo, S., Mason, P. and Johns, H., 2007. The reform of support mechanisms for
 upland farming: paying for public goods in the severely disadvantaged areas of England.
 Journal of Agricultural Economics 58, 433–453.

- Hanley, N., MacMillan, D., Wright, R.E., Bullock, C., Simpson, I., Parsisson, D. and Crabtree, B.,
- 603 1998a. Contingent valuation versus choice experiments: estimating the benefits of
 604 environmentally sensitive areas in Scotland. Journal of Agricultural Economics 49, 1–15.
- Hanley, N., Mourato, S. and Wright, R.E., 2001. Choice modelling approaches: a superior
 alternative for environmental valuation? Journal of Economic Surveys 15, 435–462.
- Hanley, N., Wright, R.E. and Adamowicz, V., 1998b. Using choice experiments to value the
 environment. Environmental and Resource Economics 11, 413–428.
- Hensher, D.A. and Greene, W.H., 2003. The mixed logit model: the state of
 practice. Transportation 30, 1331–76.
- Horne, P., Boxall, P.C. and Adamowicz, W.L., 2005. Multiple-use management of forest
 recreation sites: a spatially explicit choice experiment. Forest Ecology and Management
 207, 189–199.
- Hoyos, D., 2010. The state of the art of environmental valuation with discrete choice
 experiments. Ecological Economics 69, 1595–1603.

Hoyos, D., Mariel, P. and Fernández-Macho, J., 2009. The influence of cultural identity on the
WTP to protect natural resources: some empirical evidence. Ecological Economics 68,
2372–2381.

- Hoyos, D., Mariel, P., Pacual, U. and Etxano, I., 2012. Valuing a Natura 2000 network site to
 inform land use options using a discrete choice experiment: an illustration from the Basque
 Country. Journal of Forest Economics 18, 329–344.
- Jacobsen, J.B., Lundhede, T.H., Martinsen, L., Hasler, B. and Thorsen, B.J., 2011. Embedding
 effects in choice experiment valuations of environmental preservation projects. Ecological
 Economics 70, 1170–1177.

- Johnston, R.J. and Duke, J.M., 2007. Willingness to pay for agricultural land preservation and
 policy process attributes: does the method matter? American Journal of Agricultural
 Economics 89, 1098–1115.
- Jones, M., 2007. The European Landscape Convention and the question of public
 participation. Landscape Research 32, 613–633.
- Kallas, Z., Gómez-Limón, J.A. and Arriaza, M., 2007. Are citizens willing to pay for agricultural
 multifunctionality? Agricultural Economics 36, 405–419.
- Krinsky, I. and Robb, A., 1986. On approximating the statistical properties of elasticities. The
 Review of Economics and Statistics 68, 715–719.
- 634 Latasa, I., Lozano, P.J., Cadiñanos, J.A., Meaza, G. Varela, and Gómez, D.C., 2014: El catálogo de 635 paisaje de La Llanada Alavesa. Ejemplo de desarrollo de una figura de planificación 636 paisajística en el contexto de un territorio del País Vasco / The landscape catalogue of 637 Llanada Alavesa. Example of the development of a landscape-planning figure within a rural 638 area of the Basque Country. Proceedings of the VIII Congreso Español de Biogeografía: 639 Biogeografía de Sistemas litorales. Dinámica y conservación, 23-26 September, Seville, 640 Spain, pp. 257–266. Universidad de Sevilla y Grupo de Geografía Física de la AGE, ISBN: 978-641 84-617-1068-3.
- Latasa, I., Lozano, P.J., Del Val, M., Cadiñanos, J.A., Varela, R., Davila, N. and Fernández, P., 2012.
 La protección, gestión y ordenación de los paisajes de Álava. El catálogo de paisaje de la
 Llanada Alavesa (País Vasco) / The protection, management and planning of the landscapes
 of Álava. The landscape catalogue of Llanada Alavesa (Basque Country). Porceedings of the
 VII Congreso Español de Biogeografía: Las zonas de Montaña. Gestión y Biodiversidad, 3-7
 September 2012, Lérida, Spain, pp. 343–349. GRAMP-Departamento de Geografía,
 Universitat Autònoma de Barcelona, ISBN: 978-84-616-0354-1.

Liekens, I., Schaafsma, M., De Nocker, L., Broekx, S., Staes, J. and Aertsens, J., 2013. Developing
a value function for nature development and land use policy in Flanders, Belgium. Land Use
Policy 30, 549–559.

Lindberg, K., Fredman, P. and Heldt, T., 2009. Facilitating integrated recreation management:
assessing conflict reduction benefits in a common metric. Forest Science 55, 201–209.

Loidi, J., Biurrun, I., García-Mijangos, I., Campos, J. A. and Herrera, M., 2011. La vegetación de la
Comunidad Autónoma del País Vasco. Leyenda del mapa de series de vegetación a escala
1: 50.000 / The vegetation of the Basque Autonomous Community. Legend of the map of
vegetation series at 1:50,000 scale. University of the Basque Country (Electronic Edition).

Louviere, J., Hensher, D.A. and Swait J.D., 2000. Stated Choice Methods Analysis and Application.
 Cambridge University Press, Cambridge.

660 Lozano, P.J., Cadiñanos, J.A., Cámara-Artigas, R. y La Roca, N., 2012. Aplicación de una 661 metodología de inventariación biogeográfica a diversos ejemplos de hayedos del norte 662 peninsular para su caracterización y comparación / An application of a bio-geographic 663 inventorying methodology to various examples of northern peninsular beech woods for their characterization and comparison. Porceedings of the VII Congreso Español de 664 Biogeografía: Las zonas de Montaña. Gestión y Biodiversidad, 3-7 September 2012, Lérida, 665 666 Spain, pp. 238–243. GRAMP-Departamento de Geografía, Universitat Autònoma de 667 Barcelona, ISBN: 978-84-616-0354-1.

Mariel, P., de Ayala, A., Hoyos, D. and Abdullah, S., 2013. Selecting random parameters in
discrete choice experiment for environmental valuation: a simulation experiment. Journal
of Choice Modelling 7, 44–57.

Martin-Ortega, J., Giannoccaro, G. and Berbel, J., 2011. Environmental and resource costs under
water scarcity conditions: an estimation in the context of the European Water Framework
Directive. Water Resource Management 25, 1615–1633.

- 674 McFadden, D. and Train, K., 2000. Mixed MNL models for discrete response. Journal of Applied
 675 Econometrics 15, 447–470.
- 676 McFadden, D., 1974. Conditional logit analysis of qualitative choice behaviour. In Zarembka, P.
 677 (Ed.), Frontiers in Econometrics. Academic Press, New York, pp. 105–142.
- Olschewski, R., Bebi, P., Teich, M., Hayek, U.W. and Grêt-Regamey, A., 2012. Avalanche
 protection by forests a choice experiment in the Swiss Alps. Forest Policy and Economics
 15, 108–113.
- Palacios, I., 2011. Los bosques del País Vasco y cambio climático / The forests of the Basque
 Country and climate change. In Bernal, H., Sierra, C.H., Onaindia, M. and Gonzales, T.A.
 (Eds), Bosques del Mundo, Cambio Climático y Amazonía. University of the Basque Country,
 Bilbao, Spain, pp. 23–33.
- Pascual, U., 2007. El valor económico total del bosque en la comunidad autónoma del País Vasco
 / The total economic value of forests in the Basque Autonomous Community. Technical
 Report, IKT Institute, Arkaute, Vitoria-Gasteiz.

688 Rambonilaza, M. and Dachary-Bernard, J., 2007. Land-use planning and public preferences: what

can we learn from choice experiment method? Landscape and Urban Planning 83, 318–326.

- Ruiz, E. and Galdós, R.., 2013. La pérdida de los espacios agrarios, artificialización del suelo y
 forestación en España y en el País Vasco / The loss of agricultural areas, land artificialisation
 and forestry in Spain and in the Basque Country. Lurralde: Investigación y Espacio 36, 121–
 133.
- Scarpa, R. and Rose, J.M., 2008. Design efficiency for non-market valuation with choice
 modelling: how to measure it, what to report and why*. Australian Journal of Agricultural
 and Resource Economics 52, 253–282.

- 697 Schaafsma, M., Brouwer, R. and Rose, J., 2012. Directional heterogeneity in WTP models for
 698 environmental valuation. Ecological Economics 79, 21–31.
- Siikamäki, J. and Layton, D.F., 2007. Discrete choice survey experiments: a comparison using
 flexible methods. Journal of Environmental Economics and Management 53, 122–139.
- Small, K.A. and Rosen, H.S., 1981. Applied welfare economics with discrete choice models.
 Econometrica 49, 105–130.
- 703 Train, K., 2003. Discrete Choice Methods with Simulation. Cambridge University Press, New York.
- 704 Uribe-Echebarría, P.M., 2010. Estudio de caracterización botánica de los bosques de fondo de
- valle del municipio de Vitoria-Gasteiz (Álava) / Study on botanical characterization of the
- valley-bottom forests of the municipality of Vitoria-Gasteiz (Álava). Centro de Estudios
- 707 Ambientales, Ingurugiro Gaietarako Ikastegia, Vitoria-Gasteiz.
- van der Heide, C.M. and Heijman, W., 2013. Landscape and economics. Perceptions and
 perspectives. In van der Heide, C.M. and Heijman, W. (Eds.), The Economic Value of
 Landscapes. Routledge, London and New York, pp. 1–19.
- 711 Whitehead, J.C., 2006. A practitioner's primer on contingent valuation. In Alberini, A. and Kahn,
- 712 J. (Eds.), Contingent Valuation Handbook. Edward Elgar, Cheltenham.

713 Figures



Figure 1. Location of the area of Llanada Alavesa in the Basque Autonomous Community (Spain)



716 Figure 2. Example of a choice set (translated into English)



717 Figure 3. Empirical distributions for 218 individuals derived non-parametrically for three

718 random parameters

719 Tables

Attribu	utes	Levels				
Intensive farming (IF)	% surface	29%*	20%	15%	35%	
Organic farming (OF)	% surface	16%*	25%	30%	8%	
Native forests (NF)	% surface	39%*	45%	50%	30%	
Cemented surface (CS)	% surface	14%*	16%	20%	25%	
Recreation areas (RA)	Conservation & protection level	Medium*	Very high	High	Low	
Cost	Annual payment	€0*	€5	€15	€30	€50

720 Table 1. Attributes and levels considered in the DCE

721 * Levels with an asterisk represent the status quo scenario.

722 Table 2. Socio-economic variables and summary statistics

Variable	Mean	Std. Dev.	Min.	Max.	Sample	Population values
Age	42.284	14.678	18	76	218	42.2
Gender	0.504	0.500	0	1	218	49.7% (male), 50.3% (female)
Medium income	0.474	0.500	0	1	211	Average disposable personal income = €17,209 per year
No income	0.246	0.431	0	1	211	13.6% (last census of 2011)
Employed	0.544	0.498	0	1	217	63% (occupancy rate)
Unemployed	0.115	0.319	0	1	217	12.2% (unemployment rate)
Higher education	0.188	0.391	0	1	218	10.8%
No education	0.064	0.245	0	1	218	10.2%
Resident	0.762	0.426	0	1	218	79.8%
Low income	0.308	0.462	0	1	211	
Basque identity	0.157	0.363	0	1	217	
High recreation	0.472	0.500	0	1	218	

Table 3. Estimated models

	MNL	RPL-1	RPL-2
Variable	Coef.	Coef.	Coef.
No. 11 - 1 - 1 - 1 - 1 - 1 - 1 - 1 - 1 - 1	(Std. Error)	(Std. Error)	(Std. Error)
Native forests	(0.036 ***		
Intensive farming	0.016	0.014	0.068 ***
Organic farming	(0.012) 0.056 ***	(0.016) 0.072 ***	(0.023) 0.062 ***
Cemented surface	(0.013) 0.008	(0.016)	(0.017)
	(0.017)		
Recreation areas	0.256 *** (0.043)	0.307 *** (0.047)	0.293 *** (0.078)
Cost	-0.051 *** (0.003)		
ASC1	0.690 ***	1.505 ***	1.570 ***
4500	(0.123)	(0.134)	(0.143)
ASCZ	(0.118)	(0.132)	(0.141)
Mean of parameter distribution	· · · ·	· · ·	
Native forests (normal dist.)		0.044 **	0.034 *
		(0.018)	(0.019)
Cemented surface (normal dist.)		-0.015 (0.027)	-0.017 (0.029)
Cost (lognormal dist.)		-2.487 ***	-2.469 ***
Std. Deviation of parameter distribution		(01200)	(0.10.1)
Native forests (normal dist)		0 063 ***	0 062 ***
Native forests (normal dist.)		(0.014)	(0.015)
Cemented surface (normal dist.)		0.169 *** (0.024)	0.183 *** (0.024)
Cost (lognormal dist.)		1.108 ***	1.091 ***
		(0.087)	(0.087)
Heterogeneity in mean			
Native forests: Basque identity			0.087 *** (0.030)
Organic farming: Basque identity			0.093 *** (0.022)
Intensive farming: Resident			-0.066 ***
Recreation areas: Low income			-0.244 **
Pocreation areas: High recreation			(0.101)
			(0.097)
Log-likelihood	-1228.319	-1115.613	-1095.605
AIC	1.890	1.723	1.699
BIC Bernard	1.922	1./66	1./63
Pseudo K-squared	210	0.224	U.237
Observations	210 1308	210 1308	210 1308
K	8	11	16

***, **, *: significance at the 1%, 5% and 10% level. AIC is the "Akaike Information Criterion" and BIC is the "Bayesian Information Criterion" for the selection of models.

Group	Native forests	Intensive farming	Organic farming	Cemented surface	Recreation areas
Baseline scenario	0.71 (-0.61, 2.45)	1.45 (0.20, 3.23)	1.33 (0.18, 2.96)	0 (-4.17, 4.10)	6.30 (0.86, 13.98)
Basque identity = 1	2.58 (0.20, 6.01)		3.32 (0.45, 7.37)		
Resident = 1		0.03 (0.004, 0.07)			
Low income = 1					1.06 (0.14, 2.35)
High recreation = 1					10.57 (1.44, 23.53)
Mean €/person/year	1.00 (-0.37, 2.94)	0.37 (0.05, 0.82)	1.64 (0.23, 3.65)	0 (-4.17, 4.10)	6.75 (1.55, 13.84)

728 Table 4. Simulated WTP based on RPL-2 (€2013/person/year)

729 The tenth and ninetieth percentile points of the WTP distributions are in brackets.

730 Table 5. Alternative management scenarios for the landscapes of Llanada Alavesa

Attributes	Status quo	Scenario 1 Promotion of native forests and organic farming	Scenario 2 Promotion of intensive farming and cemented surface	Scenario 3 Promotion of recreation areas and cultural heritage
Native forests	39%	45% (个)	30% (↓)	39% (=)
Intensive farming	29%	15% (↓)	35% (个)	29% (=)
Organic farming	16%	25% (个)	8% (↓)	16% (=)
Cemented surface	14%	14% (=)	25% (个)	14% (=)
Recreation areas	Medium	Medium (=)	Medium (=)	Very high (个)
TOTAL surface	98%	99%	98%	98%

731 \uparrow/\downarrow : increase/decrease compared with the status quo level; =: equal level.